## Monterey Bay Aquarium SeafoodWatch.

## Seafood Watch ${ }^{\circledR}$ Standard for Fisheries

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## Introduction

The Monterey Bay Aquarium is committed to inspiring conservation of the ocean. To this end, Seafood Watch ${ }^{\circledR}$, a program of the Monterey Bay Aquarium, researches and evaluates the environmental impact of wild fisheries and aquaculture products and shares these seafood recommendations with the public and other interested parties in several forms, including regionally specific Seafood Watch pocket guides, smartphone apps and online at www.seafoodwatch.org.

This document houses the Seafood Watch Standard for Fisheries as approved on February 25, 2020 by the Seafood Watch Multi-Stakeholder Group. The Standard allows assessment of the relative environmental sustainability of wild-capture fisheries according to the conservation ethic of the Monterey Bay Aquarium. It includes background and rationale text explaining how the assumptions and Seafood Watch values are reflected within the calculations and scoring options. Sources from aquaculture operations and wild caught salmon are evaluated with different standards. The Standard for Aquaculture, the Standard for Fisheries and the Standard for Salmon, in addition to our assessment process, assessments and recommendations, are available at www.seafoodwatch.org.

The Standard version approved on February 25, 2020 will be used for all wild fisheries assessments beginning on or after April 1, 2020. The Standard consists of:

1. Defined guiding principles
2. Science-based performance criteria that are regularly revised based on the input from fisheries experts
3. A robust and objective scoring methodology that that results in a transparent assessment of a wildcapture fishery operation against the performance criteria

The Seafood Watch Standard for Fisheries is used to produce assessments for wild-capture fisheries resulting in a Seafood Watch rating of Best Choice (green), Good Alternative (yellow), or Avoid (red). Seafood Watch uses the assessment criteria to determine a final numerical score as well as numerical subscores and color ratings for each criterion. These scores are translated to a final Seafood Watch color rating according to the methodology described in the table below. The table also describes how Seafood Watch defines each of these categories. The narrative descriptions of each Seafood Watch color rating category, and the guiding principles listed below, compose the framework on which the criteria are based, and should be considered when providing feedback on any aspect of the criteria.

| Best Choice | Final Score $>3.2$, and either <br> Criterion 1 or Criterion 3 (or both) is <br> Green, and no Red Criteria, and no <br> Critical scores | Wild-caught and farm-raised seafood on the "Best <br> Choice" list are ecologically sustainable, well managed <br> and caught or farmed in ways that cause little or no <br> harm to habitats or other wildlife. These operations <br> align with all of our guiding principles. |
| :--- | :--- | :--- |
| Good <br> Alternative | Final score >2.2, and no more than <br> one Red Criterion, and no Critical <br> scores, and does not meet the <br> criteria for Best Choice (above) | Wild-caught and farm-raised seafood on the "Good <br> Alternative" list cannot be considered fully sustainable <br> at this time. They align with most of our guiding <br> principles, but there is either one conservation concern <br> needing substantial improvement, or there is significant <br> uncertainty associated with the impacts of this fishery or <br> aquaculture operations. |


| Avoid | Final Score $\leq 2.2$, or two or more <br> Red Criteria, or one or more Critical <br> scores. | Wild-caught and farm-raised seafood on the "Avoid" list <br> are caught or farmed in ways that have a high risk of <br> causing significant harm to the environment. They do <br> not align with our guiding principles and are considered <br> unsustainable due to either a critical conservation <br> concern, or multiple areas where improvement is <br> needed. |
| :--- | :--- | :--- |

## Seafood Watch Guiding Principles

Seafood Watch ${ }^{\circledR}$ defines "sustainable seafood" as seafood from sources, whether fished or farmed, that can maintain or increase production without jeopardizing the structure and function of affected ecosystems.

Sustainable wild capture fisheries:

1. Follow the principles of ecosystem-based fisheries management

The fishery is managed to ensure the integrity of the entire ecosystem, rather than solely focusing on maintenance of single species stock productivity. To the extent allowed by the current state of the science, ecological interactions affected by the fishery are understood and protected, and the structure and function of the ecosystem is maintained.
2. Ensure all affected stocks ${ }^{1}$ are healthy and abundant

Abundance, size, sex, age and genetic structure of the main species affected by the fishery (not limited to target species) is maintained at levels that do not impair recruitment or long-term productivity of the stocks or fulfillment of their role in the ecosystem and food web.

Abundance of the main species affected by the fishery should be at, above, or fluctuating around levels that allow for the long-term production of maximum sustainable yield. Higher abundances are necessary in the case of forage species, in order to allow the species to fulfill its ecological role.

## 3. Fish all affected stocks at sustainable levels

Fishing mortality for the main species affected by the fishery should be appropriate given current abundance and inherent resilience to fishing while accounting for scientific uncertainty, management uncertainty, and non-fishery impacts such as habitat degradation.

The cumulative fishing mortality experienced by affected species must be at or below the level that produces maximum sustainable yield for single-species fisheries on typical species that are at target levels.

Fishing mortality may need to be lower than the level that produces maximum sustainable yield in certain cases such as forage species, multispecies fisheries, highly vulnerable species, or fisheries with high uncertainty.

[^0]For species that are depleted below target levels, fishing mortality must be at or below a level that allows the species to recover to its target abundance.
4. Minimize bycatch

Seafood Watch defines bycatch as all fisheries-related mortality or injury other than the retained catch. Examples include discards, endangered or threatened species catch, pre-catch mortality and ghost fishing. All discards, including those released alive, are considered bycatch unless there is valid scientific evidence of high post-release survival and there is no documented evidence of negative impacts at the population level.

The fishery optimizes the utilization of marine and freshwater resources by minimizing post-harvest loss and by efficiently using marine and freshwater resources as bait.
5. Have no more than a negligible impact on any threatened, endangered or protected species The fishery avoids catch of any threatened, endangered or protected (ETP) species. If any ETP species are inadvertently caught, the fishery ensures and can demonstrate that it has no more than a negligible impact on these populations.
6. Are managed to sustain the long-term productivity of all affected species.

Management should be appropriate for the inherent resilience of affected marine and freshwater life and should incorporate data sufficient to assess the affected species and manage fishing mortality to ensure little risk of depletion. Measures should be implemented and enforced to ensure that fishery mortality does not threaten the long-term productivity or ecological role of any species in the future.

The management strategy has a high chance of preventing declines in stock productivity by taking into account the level of uncertainty, other impacts on the stock, and the potential for increased pressure in the future.

The management strategy effectively prevents negative population impacts on bycatch species, particularly species of concern.
7. Avoid negative impacts on the structure, function or associated biota of aquatic habitats where fishing occurs.
The fishery does not adversely affect the physical structure of the seafloor or associated biological communities.

If high-impact gears (e.g., trawls, dredges) are used, vulnerable seafloor habitats (e.g., corals, seamounts) are not fished, and potential damage to the seafloor is mitigated through substantial spatial protection, gear modifications and/or other highly effective methods.
8. Maintain the trophic role of all aquatic life

All stocks are maintained at levels that allow them to fulfill their ecological role and to maintain a functioning ecosystem and food web, as informed by the best available science.
9. Do not result in harmful ecological changes such as reduction of dependent predator populations, trophic cascades, or phase shifts
Fishing activities must not result in harmful changes such as depletion of dependent predators, trophic cascades, or phase shifts.

This may require fishing certain species (e.g., forage species) well below maximum sustainable yield and maintaining populations of these species well above the biomass that produces maximum sustainable yield.
10. Ensure that any enhancement activities and fishing activities on enhanced stocks do not negatively affect the diversity, abundance, productivity, or genetic integrity of wild stocks Any enhancement activities are conducted at levels that do not negatively affect wild stocks by reducing diversity, abundance or genetic integrity.

Management of fisheries targeting enhanced stocks ensure that there are no negative impacts on the wild stocks, in line with the guiding principles described above, as a result of the fisheries.

Enhancement activities do not negatively affect the ecosystem through density dependent competition or any other means, as informed by the best available science.

## Seafood Watch Criteria and Scoring Methodology for Fisheries

Sustainable wild-capture fisheries should ensure that the abundance of both targeted and incidentally caught species is maintained in the long term at levels that allow the species to fulfill its ecological role* while the structure, productivity, function and diversity of the habitat and ecosystem are all maintained. A management system should be in place that enforces all local, national and international laws to ensure long-term productivity of the resource and integrity of the ecosystem by adhering to the precautionary approach and responding to changing circumstances.

## Scope

Seafood Watch ${ }^{\circledR}$ recommendations apply to a single stock or species caught in a single fishery as defined by gear type, region and management body. Fisheries assessments generally focus on a single fishery, as defined by region and target species (which may include multiple target species, in the case of a multispecies fishery). A single assessment may contain multiple recommendations to address different gear types, biological stocks, or regional variations in ecological impacts and management, as needed. If a portion of a fishery is eco-certified to a standard that benchmarks as equivalent to Seafood Watch yellow or better, we may create an assessment that addresses only the uncertified portion of the fishery. The certified portion will be listed separately on our website.

## Criterion 1 - Impacts on the Species Under Assessment

## Guiding principles

Ensure all affected stocks are healthy and abundant. Abundance, size, sex, age and genetic structure should be maintained at levels that do not impair the long-term productivity of the stock or fulfillment of its role in the ecosystem and food web.

Fish all affected stocks at sustainable levels. Fishing mortality should be appropriate given current abundance and inherent vulnerability to fishing while accounting for scientific uncertainty, management uncertainty, and non-fishery impacts such as habitat degradation.

The cumulative fishing mortality experienced by affected species must be at or below the level that produces maximum sustainable yield for single-species fisheries on typical species that are at target levels.

Fishing mortality may need to be lower than the level that produces maximum sustainable yield in certain cases such as multispecies fisheries, highly vulnerable species, or fisheries with high uncertainty.

For species that are depleted below target levels, fishing mortality must be at or below a level that allows the species to recover to its target abundance.

## Assessment instructions

Evaluate Factors 1.1-1.2 under Criterion 1 to score the stock for which you want a recommendation. Evaluate Factors 2.1-2.3 under Criterion 2 to score all other main species in the fishery, including both bycatch and retained species as well as any overfished, depleted, endangered, threatened or other species of concern that are regularly caught in the fishery.

## Factor 1.1 Abundance

Goal: Stock abundance and size structure of native species is maintained at a level that does not impair recruitment or productivity

Score according to table below. In cases where there is no quantitative stock assessment available, use the data-limited stock assessment decision tree. Where directed by the data-limited stock assessment decision tree, calculate the inherent vulnerability using the Productivity-Susceptibility Analysis (MSC 2014) described below this table and in Table 1.1.2. Additional guidance on scoring forage species abundance in $\mathbf{C 1}$ and $\mathbf{C 2}$ is available in Appendix 8.

Table 1.1.1

| Conservation Concern | Description | Score |
| :---: | :---: | :---: |
| Very Low | 1. a. There is a recent stock assessment or update that has been approved through a robust scientific peer review process, AND <br> b. Biomass is estimated to be above or fluctuating around a target reference point appropriate for the species (that is based on up to date life-history and spatial distribution information) with no scientific controversy; <br> OR <br> 2. Stock is at or very near its historic high or virgin biomass; <br> OR <br> 3. Species is non-native | 5 |
| Low | 1. There is a quantitative stock assessment that is no more than 10 years old AND the biomass does not meet all the requirements for very low concern, but: <br> a. is above a limit reference point appropriate for the species, (that is based on up-to-date life-history and spatial distribution information) and at least $75 \%$ of the target reference point. (i.e., biomass may be below a target reference point); or <br> b. is estimated to be above a target reference point appropriate for the species (that is based on up-to-date life-history and spatial distribution information); <br> OR <br> 2. Quantitative stock assessment is lacking, but there is a data-limited assessment available and there is confidence that the stock is healthy and no conflicting indicators ${ }^{2}$ (see Appendix 7). | 3.67 |
| Moderate | 1. Species is above a limit reference point but below $75 \%$ of the target reference point; <br> OR <br> 2. When: <br> a. there are no stock data available, OR | 2.33 |

[^1]|  | b. there are no appropriate reference points, $\mathbf{O R}$ <br> c. a data-limited assessment is available and there is little confidence in the result due to high uncertainty, $\mathbf{O R}$ <br> d. data limited assessment methodologies provide conflicting conclusions, <br> SUCH THAT: <br> e. the Data Limited Stock Assessment Decision Tree requires that a Productivity Susceptibility Analysis be conducted, and it is determined that the stock is NOT highly vulnerable, <br> OR <br> 3. Stock is classified by management body as not overfished or has IUCN least concern status, <br> OR <br> 4. For forage species, stock biomass is above the limit reference point, but reference points are not, or it is unknown whether they are appropriate for the species. |  |
| :---: | :---: | :---: |
| High | 1. It is probable that stock is below the limit reference point, depleted/overfished, or determined to be a stock of concern, vulnerable, endangered or threatened by a state, national, or international scientific body (including COSEWIC designations of Endangered or Threatened and IUCN listings of Critically Endangered, Endangered, Vulnerable or Near Threatened; however, more recent or more regional/stock specific data can override these determinations); <br> OR <br> 2. One or more available appropriate data-limited assessment method(s) (see Appendix 7) suggest status of stock is poor; OR <br> 3. When: <br> a. there are no stock data available, $\mathbf{O R}$ <br> b. there are no appropriate reference points, OR <br> c. a data-limited assessment is available and there is little confidence in the result due to high uncertainty, OR <br> d. data limited assessment methodologies provide conflicting conclusions, | 1 |


|  | SUCH THAT: <br> e. the Data Limited Stock Assessment Decision Tree (see Appendix 7) <br> requires that a Productivity Susceptibility Analysis be conducted and it is <br> determined that the stock is highly vulnerable. |  |
| :--- | :--- | :--- |

## Instructions for Productivity-Susceptibility Analysis (for determining vulnerability)

To determine whether a species is highly vulnerable (only if needed for rating the species using the table above): If the species is a shark, sea turtle, seabird, marine mammal or coral, it is automatically considered to have "high" inherent vulnerability. The default "high vulnerability" score for these taxa can be overridden in cases where there is evidence that the population's status is not of high concern. For teleost fish and invertebrate species, score inherent vulnerability according to the PSA method described below, adapted from the Marine Stewardship Council (MSC) 2014 (available at https://www.msc.org/documents/scheneme-documents/fisheries-certificication-s-scheme-documents/fisheries-certification-requirements-version-2.0) with revisions made in 2020. Productivity attributes used in this methodology differ for fish and invertebrate species. When data are insufficient to score any given productivity attribute, that attribute can be left blank. Susceptibility attributes are assigned default values in cases where data are insufficient for scoring (see tables below).

Adapted steps from MSC instructions on conducting a PSA (for reference see description starting on page 87 of the MSC Fisheries Certification Requirements v2.0) with revisions made in 2020.

1. The analyst will use the "Seafood Watch PSA scoring tool" to calculate productivity and susceptibility scores
2. For each data-deficient stock combination (gear type, location, body of water) that is assessed using PSA, a separate PSA score will be calculated with this tool. Both productivity and susceptibility will be scored on a three-level risk scale: low, medium and high. Where there is limited or conflicting information for a productivity or susceptibility attribute, use the more precautionary (higher value) score.
3. For Productivity: See the productivity table for guidance. Productivity attributes can vary within a species depending on a number of factors including location and sex; where possible use values most applicable to the fishery being assessed. If there is variation across sexes and the fishery is not sex-selective, an average of the two values should be taken. Note that lower productivity corresponds to higher risk (and vice versa). Additional information below for certain attributes:

- Score the von Bertalanffy Growth Coefficient (K), the average maximum size and average size at maturity for fish species primarily, and invertebrate species as appropriate to their growth type. Use published literature to identify K where possible.
- Score the average maximum age for invertebrates and for finfish score average maximum age only when average maximum length is unavailable.
- Score density dependence for invertebrate species only.
- Score habitat quality for diadromous species only
- If data are unavailable for a particular attribute, leave it unscored.

4. For Susceptibility: See the susceptibility table for guidance. Note that lower susceptibility corresponds to lower risk (and vice versa). Additional information below for certain attributes:

- When scoring "areal overlap," "vertical overlap," and "seasonal overlap," consider all fisheries impacting the species.
- "Selectivity" and "post-capture mortality" should be scored with reference to the fishery under assessment only.
- Default values are provided in the table. Default values should be used unless there is evidence to the contrary.
- For "post-capture mortality" (PCM) in the absence of observer data or other verified field observations made during commercial fishing operations that indicate the individuals are released alive and post-release survivorship is high, the default value should be high. The analyst may adjust the default value when 1 ) a high score is allocated for selectivity and 2) a large portion of animals are returned alive and survive the encounter.

5. To calculate the overall score:

- Productivity score $(P)=$ arithmetic mean of the productivity attribute scores (p1, p2, p3, p4, p5, p6, p7, and p8, where p8 is only used for invertebrates)
$\ominus \quad$ Susceptibility score $(S)=$ arithmetic mean of the susceptibility attribute scores (s1, s2, s3, s4, s5)
- Vulnerability score $(\mathrm{V})=$ the Euclidean distance of 1 and 2 using the following formula: $V=\sqrt{P^{2}+S^{2}}$

6. Vulnerability Score range:

- <2.64 = Low vulnerability
$0 \geq 2.64$ and $\leq 3.18=$ Medium vulnerability
- > 3.18 = High vulnerability

7. Seafood Watch uses the high vulnerability threshold in the scoring table for 1.1 (effectively grouping low and medium vulnerability stocks).
8. PSA results of low to moderate vulnerability may be overridden with a "high vulnerability" score in cases where either:

- the species has one or more extremely vulnerable attributes under "productivity" (e.g., produces fewer than ten young per year or lives greater than 40 years), OR
- available evidence suggests a high concern with the status of similar species and/or neighboring related stocks"

Table 1.1.2a. Productivity attributes and rankings from Marine Stewardship Council 2014:

| Productivity Attribute | High productivity <br> (low risk, score = 1) | Medium productivity <br> (medium risk, score $=$ <br> 2) | Low productivity <br> (high risk, score = 3) |
| :--- | :--- | :--- | :--- |
| Average age at <br> maturity | <5 years | $5-15$ years | $>15$ years |
| Average maximum age <br> (Use for invertebrates. <br> Do not use for finfish <br> when maximum length <br> is available) | $<10$ years | $10-25$ years | $>25$ years |
| Von Bertalanffy <br> (Brody) Growth <br> Coefficient (K) (to be <br> used for species that <br> exhibit first order <br> growth) | $>0.25$ (Patrick et al. <br> 2009 ) | $0.15-0.25$ (Patrick et al. <br> 2009) | $<0.15$ (Patrick et al. <br> 2009) |
| Fecundity | $>20,000$ eggs per year | $100-20,000$ eggs per <br> year | <100 eggs per year |
| Average maximum size <br> (not to be used when <br> scoring invertebrate <br> species) | $<100$ cm | $100-300$ cm | $>300$ cm |
| Average size at <br> maturity (not to be <br> used when scoring <br> invertebrate species) | <40 cm | Broadcast spawner | Demersal egg layer or <br> brooder |
| Reproductive strategy | Live bearer <br> No depensatory or <br> compensatory <br> dynamics <br> demonstrated or likely | Depensatory dynamics <br> at low population sizes <br> (Allee effects) <br> demonstrated or likely |  |
| Density dependence <br> (to be used when <br> scoring invertebrate <br> species only) | Compensatory <br> dynamics at low <br> population size <br> demonstrated or likely | Habitat has been <br> moderately altered by <br> non-fishing impacts <br> substantially <br> compromised from <br> non-fishery impacts <br> and thus has reduced <br> capacity to support the <br> species, for example, <br> from dams, pollution, <br> or coastal <br> development. |  |
| Quality of Habitat (for <br> diadromous species <br> only) | Habitat is robust, no <br> known degradation <br> from non-fishery <br> impacts. | 40-200 cm | $>200$ cm |

Table 1.1.2b. Susceptibility attributes and rankings from Marine Stewardship Council 2014.

| Susceptibility Attribute | Low S (score = 1) | Medium S (score = 2 ) | High S (score = 3) |
| :---: | :---: | :---: | :---: |
| Areal overlap (Considers all fisheries) | Vast majority (>90\%) of species concentration (main geographic range) is unfished (considering all fisheries) (must have evidence) | Most (70\%-90\%) of species concentration is unfished by any fishery (must have evidence) | $>30 \%$ of the species concentration is fished, considering all fisheries. <br> Default score if unknown |
| Vertical overlap (Considers all fisheries) | Low overlap between fishing depths and depth range of species, i.e. most of the species depth range (>=66\%) is unfished (considering all fisheries) (Must have evidence; unlikely for any "main species") | Medium overlap between fishing depths of depth range of species, considering all fisheries, i.e. species has considerable portion (>=33\%) of depth range that is unfished (must have evidence) | High degree of overlap between fishing depths and depth range of species <br> Default score for target species, as well as any air-breathing animal, or when unknown |
| Seasonal Availability (Considers all fisheries; score using the most conservative option) | Fisheries overlap with species <3 months/year (Griffiths et al 2017) OR <br> Seasonal migrations decrease overlap with the fishery (Patrick et al 2009) | Fisheries overlap with species 3-6 months/year (Griffiths et al 2017) <br> OR <br> Seasonal migrations do not substantially affect overlap with the fishery (Patrick et al 2009) | Fisheries overlap with species $>6$ months/year (Griffiths et al 2017) <br> Default score if unknown <br> OR <br> Seasonal migrations increase overlap with the fishery (Patrick et al 2009) |
| Selectivity of fishery (Specific to fishery under assessment) | Species is not targeted AND is not likely to be captured by gear (e.g., average body size at maturity is smaller than mesh size (net fisheries), or species is not attracted to the bait used (line fisheries), or is too large to enter trap (pot/trap fisheries), etc. (if known, <33\% of individuals of this | Species is targeted, or is incidentally encountered AND is not likely to escape the gear, BUT conditions under "high risk" do not apply <br> Default score when conditions under "high risk" do not apply | Species is targeted or is incidentally encountered AND Attributes of the fishery, in combination with the species' biology or behavior, e.g. migratory bottlenecks, spawning aggregation, site fidelity, unusual attraction to gear, sequential |


|  | species encountering gear are captured) <br> Must have evidence |  | hermaphrodite, semelparity, segregation by sex, etc. increase its susceptibility to the gear: e.g. net mesh size allows retention of individuals below size at maturation, or fishery targets spawning aggregations or BOFFFFs (Hixon et al 2014) <br> If effective management measures are in place to mitigate the effect of the behavior or requirement, the behavior and/or requirement need not be considered. |
| :---: | :---: | :---: | :---: |
| Post-capture mortality (Specific to fishery under assessment) | Evidence of majority of captured individuals (>66\%) released and survive post-capture | Evidence of some (3366\%) individuals released and survive post-capture | Retained species, or majority dead when released, or unknown <br> Default score for retained species or unknown |

## Factor 1.2 Fishing Mortality

Goal: Fishing mortality is appropriate for current state of the stock.

NOTE: Ratings are based on fishing mortality/exploitation rate, e.g., $\mathrm{F} / \mathrm{F}_{\mathrm{MSy}}$ or other appropriate reference point. When determining whether a fishery is a substantial contributor, and/or whether fishing mortality is at or below a sustainable level, err on the side of caution when there is uncertainty. For further guidance, see Appendix 1 (guidance on evaluating fishing mortality). Additional guidance on scoring forage species fishing mortality in $\mathbf{C 1}$ and $\mathbf{C 2}$ is available in Appendix 8 .

Table 1.2.1

| Conservation Concern | Description | Score |
| :---: | :---: | :---: |
| Low Concern | 1. Probable (>50\% chance) that fishing mortality from all sources (including commercial, recreational, subsistence, and ghost fishing, if applicable) is at or below a sustainable level that is appropriate for the species (i.e., a level that will allow a population to maintain abundance at or rebuild to $\mathrm{B}_{\text {MSY }}$ or another appropriate reference point); <br> OR <br> 2. Species is non-native; <br> OR <br> 3. For species assessed under C2: Fishery is not a substantial contributor to fishing mortality or its contribution to mortality is expected to be low enough to not adversely affect population. | 5 |
| Moderate Concern | 1. F is fluctuating around a reference point that is appropriate for the species, <br> OR <br> 2. $F$ is above a target reference point and below a limit reference point that is appropriate for the species ${ }^{3}$ <br> OR <br> 3. Unknown ${ }^{4}$ <br> OR <br> 4. $F$ is below reference points and appropriateness of reference points is unknown. This includes most forage species, unless overfishing is occurring, or there is evidence that fishing mortality is below a reference point that is appropriate for the species. | 3 |
| High Concern | 1. Probable ( $>50 \%$ chance) or suspected that fishing mortality from all sources (including commercial, recreational, subsistence, and ghost fishing, if applicable) is above a sustainable level that is appropriate for | 1 |

[^2]|  | the speciess (i.e., a level that will allow a population to maintain abundance <br> at or rebuild to $\mathrm{B}_{\text {Msy }}$ or another appropriate reference point) (e.g., <br> overfishing is occurring); AND |  |
| :--- | :--- | :--- |
| 2.For species assessed under Criterion 2: individual fishery's contribution is <br> unknown or fishery is a substantial contributor. |  |  |

## Criterion 1 Score and Rating

Score = geometric mean (Factors 1.1, 1.2).
Rating is based on the Score as follows:

- $>3.2=$ Green
- $>2.2$ and $\leq 3.2=$ Yellow
- $\leq 2.2=$ Red


## Criterion 2 - Impacts on Other Capture Species

## Guiding principles

Ensure all affected stocks are healthy and abundant. Abundance, size, sex, age and genetic structure should be maintained at levels that do not impair the long-term productivity of the stock or fulfillment of its role in the ecosystem and food web.

Fish all affected stocks at sustainable levels. Fishing mortality should be appropriate given current abundance and inherent vulnerability to fishing while accounting for scientific uncertainty, management uncertainty, and non-fishery impacts such as habitat degradation.

The cumulative fishing mortality experienced by affected species must be at or below the level that produces maximum sustainable yield for single-species fisheries on typical species that are at target levels.

Fishing mortality may need to be lower than the level that produces maximum sustainable yield in certain cases such as multispecies fisheries, highly vulnerable species, or fisheries with high uncertainty.

For species that are depleted below target levels, fishing mortality must be at or below a level that allows the species to recover to its target abundance.

Minimize bycatch. Seafood Watch ${ }^{\circledR}$ defines bycatch as all fisheries-related mortality or injury other than the retained catch. Examples include discards, endangered or threatened species catch, bait species, pre-catch mortality and ghost fishing. All discards, including those released alive, are considered bycatch unless there is valid scientific evidence of high post-release survival and there is no documented evidence of negative impacts at the population level.

The fishery optimizes the utilization of marine and freshwater resources by minimizing post-harvest loss and by efficiently using marine and freshwater resources as bait.

Have no more than a negligible impact on any threatened, endangered or protected species The fishery avoids catch of any threatened, endangered or protected (ETP) species. If any ETP species are inadvertently caught, the fishery ensures and can demonstrate that it has no more than a negligible impact on these populations.

## Assessment instructions

The Criterion 2 score for the stock for which you want a recommendation is the lowest score of all the other main species caught with it (including both target and non-target, retained and discarded species), multiplied by the discard + bait use rate. A species is a main species if it meets any of the following conditions ("catch" here includes landings plus discards):

- A common component of the catch (as guidance, $>5 \%$ of the catch in most cases), or
- Overfished, endangered, threatened, undergoing overfishing, or otherwise a species of concern, where catch occurs regularly and may significantly contribute to the conservation concern (i.e., more than a negligible and/or sporadic level of catch). As guidance, mortality of the species caused by this fishery is $>5 \%$ of a sustainable level, or
- Fishery under assessment is one of the main sources of fishing mortality for the species, including bait species if known (as guidance, approx. $20 \%$ or more of total fishing mortality), and
- In fisheries that use bait, the bait species should be treated as a bycatch species if it meets the main species criteria outlined above. If the species used as bait are unknown but together account for greater than 5\% of the catch and no other main species have been identified, then add "unknown finfish" with abundance and fishing mortality both scored as "moderate concern".

Note: Main species should include only those species that can be caught together in a set. It should not include species that are caught during separate hauls/harvest actions/attempts/sets, even though they may be targeted or caught opportunistically in the same area, using the same gear, and potentially on the same trip. Exceptions can be made based on a case by case basis depending on the fishing method. To help identify whether a species caught or used as bait in the fishery should be considered a main species, please use the decision tree below:


## Identifying unknown species

If the main species are unknown or information on species composition is incomplete, use the Unknown Bycatch Matrices in Appendix 2 to identify those taxa that are most likely to interact substantially with the fishing gear, defined as scoring a $\mathbf{3 . 5}$ or below in the tables. Species with scores above 3.5 from the Unknown Bycatch Matrices do NOT need to be assessed. Like taxa should be grouped together such that there are assessments for "finfish," "benthic invertebrates," etc., with species identified in the text to the extent possible. Main taxa identified above can be modified using the following additional information where available and appropriate:

1. Geographic range of the fishery
2. Degree of overlap, if any (with foraging areas, breeding grounds, etc.) between fishing and potential bycatch species
3. Fishing depth
4. Whether the fishery is operating in coastal (some coastal areas may have a greater impact on some species) or open-ocean systems
5. Whether the fishery operates seasonally and coincides with breeding season, and other concerns based on fishing region and the conservation concern for the potential bycatch species.

If there is no bycatch and no other main species landed, the fishery receives a score of five for this criterion, the remaining questions in Criterion 2 can be skipped, and the assessor can continue with Criterion 3.

## Factor 2.1 Abundance

Goal: Stock abundance and size structure of all main bycatch species/stocks is maintained at a level that does not impair recruitment or productivity.

Overview: This factor is based on Factor 1.1, with additional guidance for cases where there is bycatch of unknown species. This section includes guidance on use of the Unknown Bycatch Matrices (see Appendix 2).

## Known and assessed species

When bycatch species are known and have stock assessments, follow the assessment instructions for Factor 1.1 above (the Factor for Abundance is identical for all main species caught in the fishery, whether target, other retained, or discarded).

## Known but data-limited species

In cases where bycatch species are known but not formally assessed: when a given species has no formal stock assessment but there is a data-limited assessment available (See Appendix 7.), stock abundance should be assessed as in Factor 1.1. If there are no appropriate data-limited methods that can be used to indicate stock status, follow guidance for "unassessed species" below.

## Unassessed species

In cases where bycatch species are known, but there is no indication of stock status (for example, there is no formal stock assessment or data-limited assessment), stock abundance should be assessed as in

Factor 1.1, which will require use of the Productivity-Susceptibility Analysis in most cases. Highly vulnerable taxa such as sea turtles, marine mammals, sharks, and seabirds with unknown abundance should be scored as "high concern" for abundance. Analysts should list specific bycatch species in the text, and appropriately consider species likely to be of higher vulnerability.

## Unknown bycatch composition

In cases where bycatch composition is unknown or data-limited, use the Unknown Bycatch Matrices to assess the likely bycatch species (as defined by using the instructions for "identifying unknown species" under "main species").

## To use the Unknown Bycatch Matrices:

1. Determine which taxa to include: begin by assessing each taxon listed in Appendix $2 /$ Unknown Bycatch Matrices for this type of fishery with a score of 3.5 or below. This list can be adjusted as appropriate, taking into account conditions of the particular fishery. When bycatch species are known, all taxa that include species that meet the "main species" threshold should be scored.
2. Score Factor 2.1 as "high concern" if the taxon is comprised largely of species that are either:
a. Of high vulnerability (i.e., sharks, sea turtles, marine mammals, seabirds, and coral, as well as families or genera of fish or invertebrates that are known to have high vulnerability (see list in Appendix 2),
b. Unassessed in the fishery area, but closely related species or neighboring stocks of known status are generally of high concern, or
c. Are overfished, endangered or threatened within the range of the fishery

Note: The score of "high concern" can be overridden based on data that indicate a particular species is not highly vulnerable or a specific fishery is operating differently from the standard operating procedures.
3. Score Factor 2.1 as "moderate concern" for teleost fish or invertebrates that are not from highly vulnerable taxa as defined in \#2 above.

## Factor 2.2 Fishing Mortality

Goal: Fishing mortality is appropriate for the current state of all main bycatch species/stocks.

Overview - Generally, Criterion 2.2 follows the structure of Criterion 1.2.

## Known species

Follow the assessment for Factor 1.2 above (the Factors for Abundance and Fishing Mortality are identical for all main species caught in the fishery, whether target, other retained, or discarded AND are identical for bait species used in the fishery).

## Marine Mammals in U.S. Fisheries

Additional guidance for scoring of marine mammals caught in U.S. fisheries is given below (due to the availability of data on potential biological removal (PBR) and fishing mortality rates on all bycaught
marine mammals, available in marine mammal stock assessments and List of Fisheries reports, see https://www.fisheries.noaa.gov/national/marine-mammal-protection/list-fisheries-summary-tables)

If PBR or fishery mortality relative to PBR is not known, score conservatively given what is known (e.g., fishery and/or species classification) or score as "moderate." Example: if it is unknown but the fishery is classified as Category II and the species is not strategic (see https://www.fisheries.nopaa.gov/laws-and-policies/glossary-marine-mammal-protection-act (US U.S.C. 1362[19]), score as "low."

Use Table 2.2.1.b to score marine mammals caught in non-U.S fisheries that are on the List of Foreign Fisheries (LOFF). Assess marine mammals in non-U.S. Fisheries that are NOT listed in the LOFF similarly, scoring low concern when evidence shows the fishery is not negatively impacting the recovery/stability of the marine mammal population.

Table 2.2.1.a.

| \% of PBR taken by <br> fishery | Cumulative <br> fisheries <br> mortality $>$ <br> PBR? | Seafood <br> Watch <br> Rating |
| :--- | :--- | :--- |
| $<50 \%$ | No | Low |
| $50-100 \%$ | No | Moderate |
| $<10 \%$ | Yes | Low |
| $10-50 \%$ and not <br> one of the main <br> contributors to <br> total mortality | Yes | Moderate |
| $>50 \%$ OR a main <br> contributor to <br> total fisheries- <br> related mortality | Yes | High |

Table 2.2.1.b.

| LOFF <br> Category | Presence of injury or <br> mortality (P/A in LOFF)? | Seafood Watch Rating |
| :--- | :--- | :--- |
| Exempt | No | Not included |
|  | Category III | Low concern |
|  | No or unknown | Use the UBM |
|  | Yes | Score as you would score a U.S. <br> fishery with no PBR described <br> below |

## Unknown or data limited species

If a Productivity-Susceptibility Analysis is used to score abundance for species with no abundance and fishing mortality data, use Table 1.2.1 to score fishing mortality for that species (this will most likely result in a score of "moderate" for unknown fishing mortality). If bycatch includes marine mammals, sea turtles, seabirds and/or highly vulnerable sharks, and there is no assessment of the fishery's impact on these species, use the Unknown Bycatch Matrix (UBM) to score fishing mortality for these species. Where bycatch species are not fully known, but taxonomic groups at risk are known or can be inferred, group species by taxon and use the UBM to score each group. Note that the UBM score can be
overridden if the evidence suggests that bycatch species caught in the particular fishery being assessed are not of high concern.

Taxa scored using the UBMs should be scored according to the table in Appendix 2 and the table below. As with determining main species and scoring abundance, if there are data that indicate a specific fishery is operating differently from the standard operating procedures, the Unknown Bycatch Matricestes can be overruled.

Table 2.2.2

| Bycatch score from Unknown Bycatch <br> Matrices (1-5) | Fishing Mortality |
| :--- | :--- |
| $>=3.5$ | Low Concern |
| $2.5-3$ | Moderate Concern |
| $1-2$ | High Concern |

## Factor 2.3 Modifying Factor: Discards and Bait Use

Goal: Fishery optimizes the utilization of marine and freshwater resources by minimizing post-harvest loss. For fisheries that use bait, bait is used efficiently.

Overview: While the rest of Criterion 2 focuses on the population impacts on bycatch and other capture species, Factor 2.3 addresses the issue of the waste associated with high discards or bait use in capture fisheries. The score is adjusted downward based on high discards + bait use and the color rating of Criterion 2 is affected accordingly.

Because bait use is considered in 2.3 but is rarely quantified, we aim to provide default scores for bait use, based on literature review, for a variety of fishery types (target species and gear). We will provide an opportunity to override these default scores if data specific to the fishery can be provided.

Instructions: This weighting factor is addressed once for each fishery under assessment. Both bait and dead discards are considered relative to total landings. This ratio refers to the total dead discards and/or bait use relative to total landings of all species caught in the fishery. The discard mortality rate is generally assumed to be 100\% (i.e., all discards count as dead discards). Exceptions include cases where research has demonstrated high post-release survival, including invertebrates caught in pots and traps. Research that demonstrates high post-release survival for the same or similar species caught with the same or comparable gear types may qualify as showing high post-release survival. When discard mortality rates are known, multiply these rates by the amount of discards for the relevant species to determine the amount of dead discards. If the bycatch-to-landings ratio and/or bait use are unknown, refer to average bycatch rates for similar fisheries (based on gear type, target species and/or location) as given in review papers (e.g., Kelleher 2005 and Alverson et al. 1994, NMFS 2013). Bait use, if unknown, need only be addressed in cases where it is likely to be substantial relative to landings (e.g., lobster pot fishery). Err on the side of caution when there is no information.

If the amount of dead discards plus bait use relative to total landings (in biomass or numbers of fish, whichever is higher) exceeds $100 \%$ (i.e., discards plus bait exceeds landings), modify the total score for

## Criterion $\mathbf{2}$ by multiplying by a factor of $\mathbf{0 . 7 5}$. Other fisheries are unaffected (given a score of $\mathbf{1}$ ).

Table 2.3.1

| Ratio of bait + <br> discards/landings | Factor 2.3 <br> score |  |
| :--- | ---: | :---: |
| $<100 \%$ | 1 |  |
| $\geq 100 \%$ | 0.75 |  |

## Criterion 2 Score and Rating

Criterion 2 Score for the stock for which you want a recommendation = Subscore * Discard Rate (Factor 2.4).

- $\quad$ Subscore = lowest subscore of all other assessed species caught.
- Subscore for each species = geometric mean (Factors 2.1, 2.2).

Rating is based on the lowest Score as follows:

- $>3.2=$ Green
- $>2.2$ and $\leq 3.2=$ Yellow
- $\leq 2.2=$ Red


## Criterion 3 - Management Effectiveness

## Guiding principles

The fishery is managed to sustain the long-term productivity of all affected species
Management should be appropriate for the inherent resilience of affected aquatic life and should incorporate data sufficient to assess the affected species and manage fishing mortality to ensure little risk of depletion. Measures should be implemented and enforced to ensure that fishery mortality does not threaten the long-term productivity or ecological role of any species in the future.

## Assessment instructions

Generally, 3.1 assesses management strategies for retained species and 3.2 assesses management strategies for discarded species. However, a single species can both be retained and discarded, which complicates this clear division. Therefore, the division between 3.1 and 3.2 can be based on different types of management strategies rather than by species.

- 3.1: Strategies for managing catch - i.e. fishery stock management, such as setting total allowable catches etc., should be evaluated under 3.1. If the fishery lacks regulations to either manage or prevent catch of a particular species, that should be addressed under 3.1 if the species is ever retained or sold, even if it's a relatively minor species.
- 3.2: Strategies for preventing catch - i.e. avoiding undesired, endangered or protected species, including gear modifications, etc., should be evaluated under 3.2.

The lack of regulations preventing catch of any protected or endangered species, marine mammals, etc. that are not retained and that are vulnerable to the fishery should always be considered under 3.2.

Assess Factors 3.1 through 3.5 once for each fishery. See table below to calculate final C3 score.

Step 1: Assign a rating for each of the five management subfactors using the table below:
(Note: if a "Critical" is scored for 3.1, you can continue to the overall scoring table for Management without needing to score other subfactors)

## Factor 3.1 Management Strategy and Implementation

Goal: Management strategy has a high chance of preventing declines in stock productivity by taking into account the level of uncertainty, other impacts on the stock, and the potential for increased pressure in the future. See Appendix 3 for more guidance.

Table 3.1.1

| 3.1 <br> Management strategy and implementation | Description |
| :---: | :---: |
| Highly Effective <br> Goal: Fishery has highly. appropriate strategy and goals and there is evidence that the strategy is being implemented successfully | 1. For more than $70 \%$ of the fishery's main targeted and retained, native species/stocks (by number), appropriate management/conservation targets have been defined (e.g., reference points); <br> AND <br> 2. More than $70 \%$ of the fishery's main targeted and retained native species/stocks (by number) have precautionary policies in place; these are based on scientific advice and incorporate uncertainty and environmental variability; they are characteristic of flexible and resilient fisheries management, are risk averse, include regulations to control fishing mortality over the full extent of the stock, prevent localized depletions, and respond to the state of the stock within appropriate timeframes ${ }^{5}$ (see Appendix 3 for examples of highly effective management strategies); <br> AND <br> 3. Effective strategies are in place for targeted/retained, overfished, depleted, endangered or threatened species that will allow for recovery with a high likelihood of success in an appropriate timeframe; <br> AND <br> 4. There is evidence that the strategy is being implemented successfully; <br> AND <br> 5. Management is responsive to changes in stock productivity and/or biomass; <br> AND |

[^3]|  | 6. Harvest control rules include conservative buffers appropriate for the species (e.g. forage species) that are accepted without scientific controversy and are demonstrated to be effective; <br> OR <br> 7. For NON-NATIVE species, <br> a. strategies are in place that: <br> 1) prevent further spread of and reduce biomass over time or suppress biomass to low levels (e.g., below $\mathrm{B}_{\text {msy }}$ or appropriate reference point); or <br> 2) include mechanisms to allow for recovery of species impacted by the nonnative; <br> AND <br> b. Management does not exacerbate concern with the non-native, e.g., through stocking or seeding. |
| :---: | :---: |
| Moderately effective | Fishery does not meet all the standards of "highly effective" management, but <br> 1. For more than $70 \%$ of the fishery's main targeted and retained, native species/stocks (by number), management measures in place still exceed those for "Ineffective" or "Critical" management; <br> AND <br> 2. For more than $70 \%$ of the fishery's main targeted and retained, native species/stocks (by number), measures that are expected to be effective are in place (see Appendix 3), but: <br> a. There is a need for increased precaution (e.g., stronger reductions in TAC when biomass declines, quicker reaction to changes in populations, etc.); <br> OR <br> b. Effectiveness is unknown and it is UNLIKELY that the fishery is having serious negative impacts on any retained populations (e.g., statuses of all main retained populations are known and none are scored red); <br> OR <br> c. Measures have not been in place long enough to evaluate their success; <br> OR |


|  | d. There is uncertainty regarding implementation of the management measures in place, but the instruments to ensure effective implementation exist (e.g. relevant national, regional and local legislation) and there is no evidence of systematic non-compliance; <br> AND <br> 3. Species of Concern, and Overfished or Depleted Stocks <br> a. For all targeted/retained species that are overfished or depleted, management has a rebuilding or recovery strategy in place whose eventual success is probable; or <br> b. Best management practices to minimize mortality of "stocks of concern" are implemented and are believed to be effective; <br> AND <br> 4. Non-Native Species <br> a. Management measures or harvesting prevent increases in stock size and further spreading; and <br> b. If any stocking or seeding occurs, species is already established, and ongoing stocking/seeding activity has been demonstrated not to contribute to growth or spread of non-native population. |
| :---: | :---: |
| Ineffective | Management exceeds the standard of "critical" below, but (for at least $30 \%$ of the fishery's main, primary targeted and retained native species/stocks by number): <br> 1. Management effectiveness is unknown and it is LIKELY that the fishery is having serious negative impacts on retained populations (e.g., Criterion 1 and/or Criterion 2 is scored red due to concerns with the status of one or more main retained populations); <br> OR <br> 2. There is uncertainty regarding implementation of the management measures in place and the necessary instruments to ensure effective implementation are lacking (e.g. relevant national, regional and local legislation) or there is evidence of systematic non-compliance; <br> OR <br> 3. There is no management and it is UNLIKELY that the fishery is having serious, negative impacts on any retained populations (e.g., statuses of all main retained populations are known and none are scored red); |


|  | OR <br> 4. Management sets catch limits above scientifically recommended levels, or otherwise disregards scientific advice; <br> OR <br> Management is not responsive to changes in stock productivity and/or biomass <br> 5. <br> OR <br> 6. The fishery lacks management measures that are reasonably expected to be effective, appropriate strategies for rebuilding species of concern, or appropriate control of non-native fished species (where applicable) as detailed under "moderately effective" (\#2-5) above. |
| :---: | :---: |
| Critical | 1. Management strategy is insufficiently precautionary to protect retained populations or strategies have not been implemented successfully; <br> OR <br> 2. There is no management where clearly needed; <br> OR <br> 3. The fishery targets and/or regularly retains overfished, depleted, endangered or threatened species and the fishery is a substantial contributor to mortality of the species, and management lacks an adequate rebuilding or recovery strategy and/or effective practices designed to limit mortality of these species (for example, overfishing is occurring); <br> OR <br> 4. For non-native species, there are known or likely negative impacts on the ecosystem, fishery is maintained in part through stocking/seeding/etc., and/or stock size or further spread are not controlled by harvesting or other strategies; <br> OR <br> 5. Fishery management does not comply with relevant legal requirements; <br> OR |

6. Substantial Illegal fishing; 25\% or more of the product is caught illegally.

## Factor 3.2 Bycatch Strategy

Goal: Management strategy prevents negative population impacts on bycatch species, particularly species of concern.

Table 3.2.1

| 3.2: Bycatch <br> Strategy | Description |
| :--- | :--- |
| Effective | Fishery has no or very low (<5\%) bycatch (including any unintended or unmanaged <br> catch, even if retained), with no bycatch of species of concern; or <br> IF species of concerne are caught or the fishery is not highly selective (i.e., rate of <br> discards, non-target or unmanaged catch exceeds 5\% of landings): |
| 1. The fishery has a highly effective or precautionary strategy and goals designed to |  |
| minimize the impacts of the fishery on bycatch species; |  |


|  | The fishery must have some bycatch management measures in place to meet the "moderately effective" threshold (including implementing an appropriate Take Reduction Plan for U.S fisheries listed as Category I for marine mammal bycatch, and measures to mitigate ghost fishing if there is a demonstrated concern with or high likelihood of ghost fishing), BUT either <br> 1. The strategy or implementation effectiveness is under debate or uncertain (e.g., bycatch limits are imposed based on assumptions, but limits are disputed or unsure); <br> OR <br> 2. Bycatch reduction techniques are used but are of unknown or uncertain effectiveness; <br> OR <br> 3. Management has not been in place long enough to evaluate its effectiveness or is unknown; <br> AND <br> 4. Where applicable, prevent, mitigate or remove ghost gear. Where there is a known issue with ghost gear from a fishery, management measures are being developed. |
| :---: | :---: |
| Ineffective | IF species of concern are caught or the fishery is not highly selective (i.e., rate of discards, non-target or unmanaged catch exceeds $5 \%$ of landings): <br> 1. The bycatch management measures are insufficient given the potential impacts of the fishery (e.g., fishery has bycatch of species of concern or a high discard rate and does not implement best practice measures for all such species, evidence to show that U.S. MMPA Take Reduction plan is ineffective); <br> OR <br> 2. There is strong evidence that shark finning is taking place in this fishery; <br> OR <br> 3. Fishery does not comply with all relevant legal requirements regarding bycatch; |


| OR |
| :---: | :---: |
| 4. If a fishery has a demonstrated concern with or high likelihood of ghost fishing (of <br> target or non-target species), management measures are insufficient to prevent, <br> mitigate or remove potential ghost fishing, or are non-existent and they are not being <br> developed; |
| OR |
| 5. If management of a fishery used to supply bait being used in the fishery under <br> assessment is known to be poor (for example the fishery is rated as Avoid and/or <br> management is scored as ineffective by Seafood Watch). |

## Factor 3.3 Scientific Data Collection and Analysis

Table 3.3.1

| 3.3: Scientific <br> Data <br> Collection and Analysis | Description |
| :---: | :---: |
| Highly Effective | 1. The management process uses an independent and up-to-date scientific stock assessment or analysis, or other appropriate method that seeks knowledge related to stock status; <br> AND <br> 2. This assessment is complete and robust, is peer reviewed by a scientific body, includes all major, relevant sources of fishing mortality (e.g., recreational fishing), and contains both fishery-independent data, including abundance data, and appropriate fishery-dependent data; <br> AND <br> 3. Abundance and geographic range of any non-native target species are monitored; <br> AND <br> 4. Bycatch is appropriately monitored; |


|  | AND <br> 5. Adequate observer coverage or video monitoring and data collection and analysis are sufficient to ensure that goals are being met for both bycatch and retained species; <br> AND <br> 6. Fisheries, especially those using pots/traps and gillnets (and other fisheries employing gears which have demonstrated ghost gear impacts), must collect data on lost gear or otherwise demonstrate a method to include ghost fishing impacts in the assessment of fishing mortality; <br> AND <br> 7. For forage species, stock assessments are conducted with sufficient frequency (based on the specific attributes of the species) to account for their dynamic nature and recognize fluctuations in biomass and/or productivity. |
| :---: | :---: |
| Moderately Effective | 1. Some data related to stock abundance and health are collected and analyzed. Data may not be sufficient to meet "highly effective" category, but are used to monitor and maintain the stock (including monitoring of bycatch) using appropriate datalimited assessment methods and management strategies; <br> OR <br> 2. Management relies on an appropriate strategy that requires only minimal monitoring (e.g., large protected areas including spawning habitat, and other appropriate "data-less" management techniques); <br> AND <br> 3. If there is a high risk of ghost fishing associated with the fishery, research is underway to determine the extent of the problem and identify possible solutions. |
| Ineffective | 1. No data or very minimal data are collected or analyzed; appropriate data-limited assessment and management methods are not used (see Appendix 7); <br> OR |


| 2.Bycatch monitoring or assessment is insufficient given potential bycatch impacts of <br> the fishery (e.g., observer coverage may be needed for fisheries encountering <br> endangered species, whereas fisheries with selective gear, comprehensive studies <br> demonstrating low concern with bycatch or data-limited management strategies <br> such as area closures that limit bycatch potential may not need a high level of <br> monitoring); |
| :--- | :--- |
| OR3. The fishery's main, targeted species are unassessed and regulations to constrain <br> fishing mortality for these species are lacking. |

## Factor 3.4 Enforcement of and Compliance with Management Regulations

Table 3.4.1

| 3.4: <br> Enforcement of and Compliance with management regulations | Description |
| :---: | :---: |
| Highly Effective | 1. The appropriate permits, regulations, requirements of biological opinions, (or equivalent documents for non-U.S. fisheries) and agreed-upon, voluntary arrangements are regularly enforced and independently verified, including VMS, logbook reports, dockside surveillance and other similar measures appropriate to the fishery; <br> AND <br> 2. Capacity to control, ensure, and report compliance are appropriate to the scale of the fishery including the detection and prevention of illegal fishing; <br> AND <br> 3. If applicable, $100 \%$ of at-sea transshipments must be observed. |
| Moderately Effective | Enforcement and/or surveillance are in place to ensure goals are successfully met, although effectiveness of enforcement/surveillance may be uncertain (e.g., regulations are enforced by fishing industry or by voluntary/honor system, but without regular independent scrutiny). |


| Ineffective | Enforcement and/or surveillance is lacking or believed to be inadequate, or compliance <br> is known to be poor. |
| :--- | :--- |

## Factor 3.5 Stakeholder Inclusion

Table 3.5.1

| 3.5: <br> Stakeholder Inclusion | Description |
| :---: | :---: |
| Highly Effective | The management process is transparent and includes stakeholder input, which means managers: <br> 1. Involve all major user groups; <br> AND <br> 2. Provide a mechanism to effectively address user conflicts; <br> AND <br> 3. Encourage high participation in both the assessment and management process; <br> AND <br> 4. Make transparent decisions; <br> AND <br> 5. There is an effective and constructive relationship between managers, scientists, and fishermen. |
| Moderately Effective | 1. The management process is transparent and includes stakeholder input; <br> BUT <br> 2. All user groups are not effectively considered, or there is no mechanism in place to effectively address user conflicts. |
| Ineffective | 1. Stakeholders are not included in decision-making; <br> OR <br> 2. Decisions are not made transparently. |

Step 2: Assign a rating and a score for management effectiveness (Criterion 3) based on the five factors rated above.

Table 3.6

| Conservation Concern: | Description | Score |
| :---: | :---: | :---: |
| Very Low | Meets or exceeds the standard of "highly effective" management for all five factors | 5 |
| Low | 1. Meets or exceeds all the standards for "moderately effective" management for all five subfactors; <br> AND <br> 2. Meets or exceeds the standard of "highly effective" management for, at a minimum, "management strategy and implementation" (3.1); <br> BUT <br> 3. At least one other factor is not "highly effective." | 4 |
| Moderate | 1. Meets or exceeds all the standards for "moderately effective" management for all five subfactors; <br> BUT <br> 2. "Management strategy and implementation" is not "highly effective." | 3 |
| High | 1. Meets or exceeds the standard for "moderately effective" management for, at a minimum, "management strategy and implementation" and "bycatch management"; <br> BUT <br> 2. At least one other factor is "ineffective." | 2 |
| Very High Concern | 1. "Management strategy and implementation" and/or "bycatch management" are "ineffective." | 1 |
| Critical | Fishery scores "Critical" for "management strategy and implementation"; | 0 |

Note: a score of " 0 " (Critical) for Management results in an Avoid ranking overall.

## Criterion 3 Score and Rating

Rating is based on the Score for Criterion 3 as follows:

- Green if $>3.2$
- Yellow if >2.2 and $\leq 3.2$
- Red if $\geq 1$ and $\leq 2.2$

Rating is Critical if scored a 0.

## Criterion 4 - Impacts on the Habitat and Ecosystem

## Guiding principles

Avoid negative impacts on the structure, function or associated biota of marine habitats where fishing occurs. The fishery does not adversely affect the physical structure of the seafloor or associated biological communities.

If high-impact gears (e.g., trawls, dredges) are used, vulnerable seafloor habitats (e.g., corals, seamounts) are not fished, and potential damage to the seafloor is mitigated through substantial spatial protection, gear modifications and/or other highly effective methods.

Maintain the trophic role of all marine life. All stocks are maintained at levels that allow them to fulfill their ecological role and to maintain a functioning ecosystem and food web, as informed by the best available science.

Do not result in harmful ecological changes such as reduction of dependent predator populations, trophic cascades, or phase shifts. Fishing activities must not result in harmful changes such as depletion of dependent predators, trophic cascades, or phase shifts.

This may require fishing certain species (e.g., forage species) well below maximum sustainable yield and maintaining populations of these species well above the biomass that produces maximum sustainable yield.

Ensure that any enhancement activities and fishing activities on enhanced stocks do not negatively affect the diversity, abundance, productivity, or genetic integrity of wild stocks. Any enhancement activities are conducted at levels that do not negatively affect wild stocks by reducing diversity, abundance or genetic integrity.

Management of fisheries targeting enhanced stocks ensure that there are no negative impacts on the wild stocks, in line with the guiding principles described above, as a result of the fisheries.

Enhancement activities do not negatively affect the ecosystem through density dependent competition or any other means, as informed by the best available science.

Follow the principles of ecosystem-based fisheries management. The fishery is managed to ensure the integrity of the entire ecosystem, rather than solely focusing on maintenance of single species stock productivity. To the extent allowed by the current state of the science, ecological interactions affected by the fishery are understood and protected, and the structure and function of the ecosystem is maintained.

## Assessment instructions

Address Factor 4.1-4.2 for all fishing gears separately.

## Factor 4.1 Physical Impact of Fishing Gear on the Habitat/Substrate

Goal: The fishery does not adversely impact the physical structure of the ocean habitat, seafloor or associated biological communities.

Instructions: Fishing gears that do not contact the seafloor score 5 for this criterion, and Factor 4.2 can be skipped. Use the table below to assign a score for gear impacts (Appendix 5 provides further guidance). If gear type is not listed in the table, use the score for the most similar gear type in terms of extent of bottom contact. Note that if it can be demonstrated that a specific gear is significantly different or has been significantly modified, it can be scored accordingly. Seafood Watch will not assess a fishery using destructive practices such as explosives or cyanide regardless of habitat type and management actions; therefore, those fishing methods are not included in the table. Where multiple habitat types are commonly encountered, and/or the habitat classification is uncertain, score conservatively according to the most sensitive plausible habitat type. See Appendix 5 for further guidance and the methods used in developing the table below.

Table 4.1.1

| Description | Seafood Watch score |
| :---: | :---: |
| Gear does not contact bottom; fishing for a pelagic/open water species | 5 |
| 1. Vertical line fished in contact with the bottom; <br> OR <br> 2. Vertical line used to fish for a benthic/demersal or reef-associated species | 4 |
| 1. Bottom gillnet, trap, bottom longline except on rocky reef/boulder and corals; or <br> 2. Bottom seine (on mud/sand only); or <br> 3. Midwater trawl that is known to contact bottom occasionally (<25\% of the time); or <br> 4. Purse seine known to commonly contact bottom | 3 |
| 1. Scallop dredge on mud and sand; or <br> 2. Bottom gillnet, trap, bottom longline on boulder or coral reef; or <br> 3. Known trampling of coral reef habitat occurs; or <br> 4. Bottom seine (except on mud/sand); or <br> 5. Bottom trawl (mud and sand, or shallow gravel) (includes midwater trawl known to commonly contact bottom) | 2 |
| 1. Hydraulic clam dredge; or <br> 2. Scallop dredge on gravel, cobble or boulder; or <br> 3. Trawl on cobble or boulder, or low energy ( $>60 \mathrm{~m}$ ) gravel; or <br> 4. Bottom trawl or dredge used primarily on mud/sand (or to catch a species that associates with mud/sand habitat), but information is limited and there is the potential for the gear to contact sensitive habitat | 1 |
| Dredge or trawl on deep-sea corals or other biogenic habitat (such as eelgrass and maerl) | 0 |

## Factor 4.2 Modifying Factor: Mitigation of Gear Impacts

Goal: Damage to the seafloor is mitigated through protection of sensitive or vulnerable seafloor habitats, and limits on the spatial footprint of fishing on fishing effort.

Instructions: Assess Factor 4.2 only for fishing gear that contacts the bottom. Scores from Factor 4.2 can only improve the base score from 4.1. A high level of certainty is required to score a strong or moderate mitigation measure, e.g., good quality seafloor maps, VMS and/or observer coverage is required to document that spatial measures are effective and enforced. Further guidance can be found in Appendix 6.

Assess the fishery management's efforts to mitigate the fishery's impact on the benthic habitat. Factor 4.2 allows the habitat score to increase, based on the strength of mitigation measures, by the number of bonus points specified in the table.

Table 4.2.1

1. At least $50 \%$ of the representative habitat is protected from the gear type used in
$+1$
the fishery under assessment (see Appendix 6);

## OR

2. a. For trawl/dredge fisheries, expansion of the fishery footprint into untrawled/undredged habitat is prohibited. A rotational strategy of habitat protection if deemed appropriate is acceptable); and
b. Fishing intensity is constrained to be sufficiently low. Must have scientific evidence (using knowledge of the resilience of the habitat and the frequency of fishing impacts from the gear type used in the fishery under assessment (see Appendix 6)), that at least $50 \%$ of the representative habitat is in a recovered state and will remain so under current management; and
c. Vulnerable habitats are strongly protected;

## OR

3. a. Gear is specifically designed to reduce impacts on the seafloor, and
b. There is scientific evidence that these modifications are effective, and
c. Gear modifications are used on the majority of vessels;

## OR

4. Other measures are in place that have been demonstrated to be highly effective in reducing the impact of the fishing gear, which may include an effective combination of both "moderate" measures described below, e.g., gear modifications + spatial protection.
5. a. A substantial proportion of all representative habitats are protected from all bottom
contact, and
b. For trawl/dredge fisheries, expansion of the fishery's footprint into untrawled/undredged habitat is prohibited (note: this does not prohibit a rotational strategy of habitat protection if deemed appropriate), and
c. Vulnerable habitats are strongly protected;

## OR

2. Gear modifications or other measures are in use in the fishery under assessment that are reasonably expected to be effective.

| Does not meet standard for +0.5 above, or | +0 |
| :--- | :--- |

Not applicable because gear used is benign and fishery received a score of " 5 " for 4.1. +0

## Scoring for Factor 4.1: Impact on the Habitat

The score for Factor 4.1 is the sum of the score for 4.1 and the score for 4.2. The category name for 4.1 is assigned based on score ranges, as below:

Table 4.2.2

| Score (Sum of 4.1 and 4.2) | Category |
| :--- | :--- |
| $>3.2$ | Low Concern |
| $>2.2$ and $\leq 3.2$ | Moderate Concern |
| $\leq 2.2$ | High Concern |

## Factor 4.3 Ecosystem-based Fisheries Management

Goal: All stocks are maintained at levels that allow them to fulfill their ecological role and to maintain a functioning ecosystem and food web. Fishing activities should not seriously reduce ecosystem services provided by any retained species or result in harmful changes such as trophic cascades, phase shifts or reduction of genetic diversity. Even non-native species should be considered with respect to ecosystem impacts. If a fishery is managed in order to eradicate a non-native, the potential impacts of that strategy on native species in the ecosystem should be considered and rated below.

Instructions: Assign an ecosystem-based management score for the fishery. In scoring Factor 4.3, it is important to consider the management of the impact of the fishery upon the ecosystem through its interactions with all capture species. Broadly consider whether there is an ecosystem-based approach for all species caught within a fishery. Where species are released with a high probability of survival, or where a fishery is known to be a non-substantial contributor to a species' mortality, EBFM need not be considered for those species. In situations where main species are selected using the Unknown Bycatch Matrices, EBFM need not be considered for those species but should be scored in a precautionary manner (understanding that the impacts of the fishery on the ecosystem will be unknown).

Table 4.3.1

| Conservation Concern | Description | Score |
| :---: | :---: | :---: |
| Very Low | 1. a. More than $70 \%$ of the fishery's main targeted and retained native species/stocks (by number) have policies in place (e.g., harvest control rules) that are effective at protecting ecosystem functioning and accounting for species' ecological role; and <br> b. Precautionary and effective temporal and spatial management is used, e.g., to protect spawning areas, prevent localized depletion, and protect important foraging areas for predators of fished species, if applicable; and <br> c. Detrimental food web effects due to harvesting the species are highly unlikely (based on the species ecological role); <br> OR <br> 2. An ecosystem study has been conducted and it has been scientifically demonstrated that the fishery has no unacceptable ecological and/or genetic impacts; <br> AND <br> 3. For fisheries on non-native species, policies in place to manage the fishery and/or control the spread of the species do not have long-term, adverse effects on native species. | 5 |
| Low | 1. a. More than $70 \%$ of the fishery's main targeted and retained native species/stocks (by number) have policies to protect ecosystem functioning and account for capture species' ecological role but have not yet proven to be effective; and <br> b. Temporal and spatial management is used to protect ecosystem functioning; and <br> c. Detrimental food web impacts due to harvesting the species are unlikely (based on the species ecological role); <br> OR <br> 2. a. Detrimental food web impacts are possible (based on the species' ecological role); and <br> b. Precautionary and effective temporal and spatial management demonstrated with confidence in scientific peer reviewed literature to be appropriate to the scale of the fishery and ecology of the stock is used, e.g., to protect spawning areas, prevent localized depletion, and protect important foraging areas for predators of fished species, if applicable; and | 4 |


|  | c. For forage fisheries, there is an appropriate conservative, ecological harvest control rule in place that is consistent with the Lenfest Forage Fish Task Force Recommendations, with buffers built-in to account for the needs of dependent predators; <br> AND <br> 3. For fisheries on non-native species, policies in place to manage the fishery and/or control the spread of the species do not have long-term, adverse effects on native species. |  |
| :---: | :---: | :---: |
| Moderate | 1. If the fishery is a substantial contributor to forage species fishing mortality (as identified in C1 or C2), the fishery has temporal and spatial management that is appropriate to the scale of the fishery and ecology of the stock that is likely to be effective with little scientific controversy, AND uses an appropriate conservative, ecological harvest control rule that is consistent with the Lenfest Forage Fish Task Force Recommendations, with buffers built in to account for the needs of dependent predators; <br> OR <br> 2. If the fishery is not a substantial contributor to forage species fishing mortality (as identified in C1 or C2), either: <br> a: The fishery lacks temporal and spatial management or other policies to protect ecosystem functioning and account for capture species' ecological role, but detrimental food web impacts are not likely; <br> OR <br> b: Detrimental food web impacts are possible (based on the species' ecological role), and there is spatial and temporal management that is appropriate to the scale of the fishery and ecology of the stock that is likely to be effective with little scientific controversy; <br> OR <br> c: For fisheries on non-native species, the policies to manage the fishery and/or control the spread of the non-native species have an unknown effect on native species. | 3 |
| High | 1. a. The fishery lacks temporal and spatial management appropriate to the scale of the fishery and ecology of the stock or other policies to protect ecosystem functioning and account for capture species' ecological role; and <br> b. The likelihood of trophic cascades, alternative stable states, or other detrimental food web impacts resulting from the fishery are high, but | 2 |


|  | conclusive scientific evidence specifically related to the fishery are <br> lacking; <br> OR |  |
| :--- | :--- | :--- |
| 3. For forage species: A conservative, ecological harvest control rule that <br> is consistent with the Lenfest Forage Fish Task Force <br> Recommendations, with buffers built-in to account for the needs of <br> dependent predators, is NOT in place. Reference points that are <br> appropriate given species ecological role have not been defined and <br> the fishery lacks a precautionary strategy that accounts for the needs <br> of dependent predators; |  |  |
| AND |  |  |
| 4. For fisheries on non-native species, policies in place to manage the |  |  |
| fishery and/or control the spread of the species have adverse effects |  |  |
| on native species. |  |  |$\quad$| Very High |
| :--- |
| Critical |
| Trophic cascades, alternate stable states, or other detrimental food web <br> impacts are resulting from the fishery. |
| For forage species, fishery has resulted in trophic cascades, alternate stable <br> states, or other detrimental food web impacts. |

## Criterion 4 Score and Rating

Score $=$ Geometric Mean (Factors 4.1+4.2, Factor 4.3)
Rating is based on the Score as follows:

For fisheries targeting forage species:

- $>3.2$ and 4.3 is a Moderate Concern or better = Green
- $\quad>2.2$ and $\leq 3.2$ and 4.3 is a Moderate Concern or better $=$ Yellow
- $\leq 2.2$ and/or 4.3 is a High Concern or worse = Red
- 4.3 is a Critical Concern = Assessment is Red Overall

For all other fisheries:

- $>3.2$ = Green
- $>2.2$ and $\leq 3.2=$ Yellow
- $\leq 2.2=$ Red


## Overall Score and Final Recommendation

Overview: The final scoring system combines the individual criterion scores to produce a numerical final score from 0-5, but also applies decision rules based on the number of "high concerns," i.e., "red" scoring criteria as outlined below.

Specifics: The following sections show how the final score and final recommendation are calculated from the individual criterion scores. It is the current philosophy of the Seafood Watch criteria that regardless of the final numerical score, if there is one red criterion (with a numerical score $\leq 2.2$ ), then the highest possible final recommendation is a yellow "Good Alternative." If there are two red criteria, then the overall final recommendation will be red "Avoid" regardless of the numerical score. If there is one or more "critical concern" then the final recommendation is red "Avoid" regardless of the numerical score. A fishery must score a green in either Criterion 1 or Criterion 3 (or both) in order to be a "Best Choice" overall.

Final Score = geometric mean of the four Scores (Criterion 1, Criterion 2, Criterion 3, Criterion 4).
The overall recommendation is as follows:

- Best Choice = Final Score >3.2, and either Criterion 1 or Criterion 3 (or both) is Green, and no Red Criteria, and no Critical scores
- Good Alternative = Final score >2.2, and no more than one Red Criterion, and no Critical scores, and does not meet the criteria for Best Choice (above)
- Avoid = Final Score $\leq 2.2$, or two or more Red Criteria, or one or more Critical scores.


## Glossary

## Adequate observer coverage or video monitoring:

Observer coverage required for adequate monitoring depends on the rarity of the species caught, with fisheries that interact with rare species requiring higher coverage (Babcock et al. 2003). Similarly, species that are "clumped" instead of being evenly distributed across the ocean also require higher levels of coverage. In addition, fisheries using many different gear types and fishing methods require higher levels of coverage. Poorly sampled areas, gears, and seasons can bias results. For these reasons, the exact level of coverage required for a particular fishery depends on the associated discard and target species, the species' distribution within the fishery, the absolute number of sets, how common the bycatch is, and whether that level of bycatch is problematic for the species (Babcock et al. 2003). The analyst will need to determine what level of observer coverage is adequate for the fishery of interest; coverage of $17-20 \%$ (or as high as $50 \%$ for rare species bycatch) may be required in some cases but may not be necessary in all cases. There needs to be a decent sample size in terms of absolute sets and catches. If the observer coverage is sufficient to provide reliable bycatch estimates for priority species in that fishery, then the coverage is good, regardless of the percent (and vice-versa). Must present evidence that video monitoring meets the criteria for "adequate observer coverage" as defined here.

## Appropriate Quantitative Stock Assessment:

A stock assessment that has been peer reviewed and approved, for example through a scientific and statistical committee, and includes appropriate reference points that are not known to be less conservative than MSY-based approaches. There may be some uncertainty associated with such assessments; however, if they have been approved or accepted by peer reviewers, they can be considered quantitative. Stock assessments that do not meet this guidance would be considered datalimited approaches.

## Appropriate reference points:

Determination of the appropriateness of reference points depends on two questions:

1) Is the goal appropriate? Appropriate biomass reference points are designed with the goal of maintaining stock biomass at or above the point where yield is maximized (target reference points; TRPs) and safely above the point where recruitment is impaired (limit reference points; LRPs). Fishing mortality reference points should be designed with the goal of ensuring that catch does not exceed sustainable yield and has a very low likelihood of leading to depletion of the stock in the future.
2) Is the calculation of the reference points credible? There may be a concern if reference points have been lowered repeatedly or if there is scientific controversy regarding the reference points or the calculations of biomass and fishing mortality relative to reference points.
See the guidance for each type of reference point below and in Appendix 1.

Target reference point: Reference points need to be evaluated on a case-by-case basis, but in general: Biomass target reference points (TRPs) below about $\mathrm{B}_{35 \%}$ require strong scientific rationale. TRP values below about $\mathrm{B}_{35 \%}$ may not be acceptable, as deterministic reference points may not be adequately precautionary accounting for stochasticity and environmental variability. See Appendix 1 for more details.

Limit reference point: The point where recruitment would be impaired. Reference points need to be evaluated on a case-by-case basis, but in general: Biomass limit reference points (LRPs) should be no less than $1 / 2 \mathrm{~B}_{\mathrm{MSY}}$, or $1 / 2$ an appropriate target reference point such as $\mathrm{B}_{40 \%}$. LRPs below $\mathrm{B}_{20 \%}$ or $1 / 2 \mathrm{~B}_{\text {MSY }}$ require strong scientific rationale. Limit reference points set at $50 \%$ of $\mathrm{B}_{35 \%}$ may not be acceptable, as deterministic reference points may not be adequately precautionary accounting for stochasticity and environmental variability. Where the LRP is not set at 50\% of the TRP, it is important to consider the appropriateness of each RP when determining the appropriate score.

Spawning potential ratio/fraction of lifetime egg production (SPR/FLEP) reference point: The SPR/FLEP limit reference point should either be derived through scientific analysis to be at or above replacement \%SPR for the species (the threshold level of SPR necessary for replacement) based on its productivity and S-R relationship (viz., Mace and Sissenwine 1993), or should be set at about 35-40\% of LEP. An exception can be made for species with very low inherent productivity (e.g., rockfish, sharks), in which case a reference point of 50-60\% of LEP is more appropriate (Mace and Sissenwine 1993, Myers et al. 1999, Clark 2002, Botsford and Parma 2005).

## Appropriate for the species:

Whether a reference point is appropriate for a species depends on its life history characteristics, its productivity dynamics and its role in the ecosystem.

In respect to forage species: Most modern assessments use a stock-recruitment curve that is described by stationary parameters, including virgin biomass or $\mathrm{B}_{0}$ and are not appropriate for species with dynamic productivity that shifts in response to environmental conditions. While it is possible to calculate reference points based on dynamic virgin biomass (acknowledging that the carrying capacity of the environment for these species is different based on favorable to unfavorable environmental conditions), to date, none exist in practice for any species and the effectiveness of dynamic reference points is not well understood. While static reference points do not describe the shifts in productivity of forage species (instead, at best, they represent a long term average), they can be used effectively in management when 1) the harvest strategies based upon them account for volatility AND 2) when the harvest strategy outcomes have been tested using a proven, robust Management Strategy Evaluation framework, demonstrating that fishing mortality is set low enough to prevent collapse during periods of low stock productivity. Given these considerations, unless harvest strategies account for volatility and have been tested and proven to prevent stock collapse (i.e., in most situations), Seafood Watch considers forage stock biomass and fishing mortality to be highly uncertain.

Note that the best reference point to minimize the probability and severity of collapse for forage species depends on the specific attributes of the species. See Siple, Essington and Plagányi (2019). Forage fish fisheries management requires a tailored approach to balance trade-offs. Fish and Fisheries, 20(1) 110124.

## Bycatch:

Seafood Watch defines bycatch as all fisheries-related mortality or injury other than the retained catch. Examples include discards, endangered or threatened species catch, pre-catch mortality and ghost fishing. All discards, including those released alive, are considered bycatch unless there is valid scientific evidence of high post-release survival and there is no documented evidence of negative impacts at the population level.

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## Catch:

The catch of a fishery refers to any species that interacts with the fishing gear and is brought to the vessel during hauling. In some instances, it may not be brought onboard, for example purse seiners are able to release catch prior to bailing onto the deck. The catch can be further divided and referred to using the following terms:

- Retained Catch - Catch that is retained onboard vessel and landed. Traditionally retention occurred based on the economic value of the catch, however full-retention fisheries are becoming increasingly popular as a way of ensuring all catch is quantified and to reduce discarding.
- Non-retained catch - Catch that is discarded, typically at sea. This may include lowvalue species or species that cannot be landed for regulatory reasons (lack of quota, prohibited species, protected species).
- Target Catch - Species that fishers aim to catch when setting gear. This is typically higher value or more abundant species within the catch composition. Some fisheries will have one target species e.g. purse seine, while others may have several e.g. bottom trawl.
- Non-target Catch - Species that are encountered in the fishery but are not the focus of the fishing effort. Such species may form part of the retained catch (if they have some value and/or can legally be landed) or non-retained catch (e.g. marine mammals, low value species)


## Critically endangered:

An IUCN category for listing endangered species. A taxon is considered "critically endangered" (CE) when it faces an extremely high risk of extinction in the wild in the immediate future, as defined by any of the relevant IUCN criteria for "critically endangered" (FAO Glossary; IUCN).

## Data-moderate:

Reliable estimates of Target Reference Point quantities are either unavailable or not useful due to life history, a weak stock-recruit relationship, high recruitment variability, etc. Reliable estimates of current stock size, life history variables and fishery parameters exist. Stock assessments include some characterization of uncertainty (Restrepo and Powers 1998; Restrepo et al. 1998).

## Data-limited:

Refers to fisheries for which there are no estimates of MSY or relevant reference points, stock size, or certain life history traits. There may be minimal or no stock assessment data, and uncertainty measurements may be qualitative only (Restrepo and Powers 1998; Restrepo et al. 1998).

## Data-rich:

Refers to fisheries with reliable estimates of MSY-related or relevant target quantities and current stock size. Stock assessments are sophisticated and account for uncertainty (Restrepo and Powers 1998; Restrepo et al. 1998).

## Depleted:

A stock at a very low level of abundance compared to historical levels, with dramatically reduced spawning biomass and reproductive capacity. Such stocks require particularly energetic rebuilding strategies. Recovery times depend on present conditions, levels of protection and environmental conditions. May refer also to marine mammals listed as "depleted" under the Marine Mammal Protection Act (FAO Glossary). Classifications of "overfished" or "depleted" are based on assessments by the management agency and/or FAO, but analysts can use judgment to override the classification, especially where the prior assessment may be out of date (also includes IUCN listings of "near threatened", "special concern" and "vulnerable"). Inclusion in this classification based on designations such as "stock of concern" is determined on a case-by-case basis, as such terms are not used consistently among management agencies. Stocks should be classified as "depleted" if the stock is believed to be at a low level of abundance such that reproduction is impaired or is likely to be below an appropriate limit reference point. Marine mammals classified as "depleted" under the Marine Mammal Protection Act fall into this category, if not listed as endangered or threatened. Also includes stocks most likely (>50\% chance) below the level where recruitment or productivity is impaired. Note: Official IUCN listings should be overridden by more recent and/or more specific classifications, where available (e.g., NMFS stock assessment showing stock is above target levels).

## Ecological harvest control rule:

For certain taxa, like forage species that have an exceptionally important role in the ecosystem, harvest control rules (HCRs) should be based on ecosystem considerations (i.e., maintaining enough biomass to allow the species to fulfill its ecological role), rather than MSY or single-species considerations. For forage species, HCRs should be consistent with the precautionary principles and recommendations of the Lenfest Forage Fish Task Force (see specific guidance under the "Lenfest Forage Fish Task Force" entry. See also the fact sheet regarding ecological reference points at https://www.lenfestoceaan.org/.-/media/assets/2018/10/buhheister-fact-sheet-pdf.pdf).

## Ecological role:

The natural trophic role of a stock within the ecosystem under consideration in an assessment (MSC 2010).

## Effective:

Management or mitigation strategies are defined as "effective" if: a) the management goal is sufficient to maintain the structure and function of affected ecosystems in the long-term, and b) there is scientific evidence that they are meeting these goals.

## Effective mitigation or gear modification:

A strategy that is "effective" as defined above, either in the fishery being assessed or as demonstrated in a very similar system (See Appendix 3 and Appendix. 4 for a partial list of effective mitigations; this list will be continually developed).

## Endangered/threatened:

Taxa in danger of extinction and whose survival is unlikely if causal factors continue operating. Included are taxa whose numbers have been drastically reduced to a critical level or whose habitats have been so drastically impaired that they are deemed to be in immediate danger of extinction (FAO Fisheries Glossary). This classification includes taxa listed as "endangered" or "critically endangered" by IUCN or "threatened", "endangered" or "critically endangered" by an international, national or state
government body (e.g., Canada's Committee on the Status of Endangered Wildlife in Canada COSEWIC, and Species at Risk Act - SARA), as well as taxa listed under CITES Appendix I. This classification does not include species listed by the IUCN as "vulnerable" or "near threatened". Marine mammals listed as "strategic" under the Marine Mammal Protection Act are also considered as endangered/threatened if they are listed because "based on the best available scientific information, [the stock] is declining and is likely to be listed as a threatened species under the Endangered Species Act of 1973 (ESA, 16 U.S.C. 1531 et seq.) within the foreseeable future." However, marine mammal stocks listed as "strategic" because "the level of direct human-caused mortality exceeds the potential biological removal level," or because they are listed as "depleted" under the Marine Mammal Protection Act, are instead classified as species of concern. If there is more recent information to suggest that the status of the population under consideration is healthier than suggested by IUCN, for example from a data-limited stock assessment, and the IUCN assessment is greater than 10 years old the IUCN status can be overridden. If local wildlife protection listings, for example the U.S. Endangered Species Act or Canadian Species at Risk Act, are being used to override the IUCN listing, the local status must be based on biological evidence rather than a political decision not to list the species.

## Exceptional importance to the ecosystem:

A species that plays a key role in the ecosystem that may be disrupted by typical levels of harvesting, including: keystone species (those that have been shown or are expected to have community-level effects disproportionate to their biomass), foundation species (habitat-forming species, e.g., oyster beds), basal prey species (including krill and small pelagic forage species such as anchovies and sardines), and top predators, where the removal of a small number of the species could have serious ecosystem effects. Species that do not fall into any of these categories but that have been demonstrated to have an important ecological role impeded by harvest (e.g., studies demonstrating trophic cascades or ecosystem phase shifts due to harvesting) shall also be considered species of exceptional importance to the ecosystem (Paine 1995; Foley et al. 2010).

## Fish Aggregating Device (FAD):

Any floating object strategically placed in the ocean to attract and aggregate fish (from https://www.fisheries.noaa.gov/inport/item/28652). FADs may be static (anchored) or free-floating (untethered).

## Fluctuating biomass:

- If a stock is trending upwards (based on the most recent assessment) and has just recently exceeded the target reference point (TRP), it can be rated as Very Low Concern. If a stock is not trending but is truly fluctuating around the TRP (exceeding in some years and falling short in others, but with no clear trend), it can be rated as Very Low Concern. However, if a stock is fluctuating around the limit reference point (LRP), it cannot be considered Very Low Concern.
- If a stock is trending downward and is currently below the TRP, the rating can be no better than Low Concern.
- If a stock is below the LRP, it is considered a High Concern.


## Fluctuating fishing mortality:

- If $F$ is trending downwards or was previously above $F_{\text {MSY }}$ (or an appropriate reference point) but has recently gone below $\mathrm{F}_{\text {Msy }}$ (in the most recent assessment), fishing mortality should be rated as Low Concern.
- If F is fluctuating around $\mathrm{F}_{\mathrm{MSY}}$ (or an appropriate reference point), or if F has been consistently
below $\mathrm{F}_{\text {Msy }}$ and has just recently (in the latest assessment) risen above $\mathrm{F}_{\text {msy }}$ for just this one year (potentially due to management error or a new stock assessment and the consequent adjustment in reference points or estimates), fishing mortality should be rated as Moderate Concern.
- If F is trending upwards and has just risen above $\mathrm{F}_{\text {MSY }}$ (or an appropriate reference point), fishing mortality should be rated as High Concern unless there is a substantial plan to bring F back down. Such a plan would need to differ substantially from the existing harvest control rules (HCR), as those evidently did not keep $F$ at a sufficiently low level.


## Forage Species:

Forage species are small schooling fish or invertebrate species that serve as prey for larger and often commercially and recreationally important fish or invertebrates, as well as for mammals, birds, and other predators.

Forage species play an important role in food webs because they 1) exhibit high connectance ${ }^{6}$ to other organisms in the ecosystem and 2) a large amount of energy is channeled through that species. Forage species typically exhibit highly variable productivity such that there may be high uncertainty in their reference points, making it difficult to evaluate their stock status. The drivers of this variability in productivity ${ }^{7}$ may be environmental forcing and/or other factors. As a result of their importance in food webs these stocks require management that is tailored to their specific life histories and ecological roles.

Species that generally qualify as forage species include sandeels, sandlances, herrings, menhaden, pilchards, sardines, sprats, anchovies, krill, lanternfish, smelts, capelin, mackerels, silversides, sand smelts, Norway pout (adapted from MSC Fisheries Standard V2.01, p. 14). Other species or stocks may qualify if they meet the definition above. Due to differences in food web structure and function between marine and aquatic systems (including species richness and the effect of ecosystem size) it may require additional information to determine whether freshwater species qualify as forage species.

We note that in some food webs, several species may fulfill the ecological role as forage as a guild rather than as single forage species.

## Ghost fishing:

Gear that is abandoned, lost, derelict, or discarded that continues to catch, entrap, or entangle marine species.

## Harvest control rule (HCR):

HCRs are the operational component of a harvest strategy or management procedure/strategy. They are pre-agreed guidelines that determine how much fishing can take place, based on indicators of the targeted stock's status. These indicators can be based on either monitoring data or models. For example, a harvest control rule can describe the various values of fishing mortality, which will be aimed at for various values of the stock abundance. It formalizes and summarizes a management strategy. Constant catch and constant fishing mortality are two types of simple harvest control rules.

[^4]
## Harvest strategy:

Synonymous with management strategy or procedure.

## Highly appropriate management strategy:

Management that is appropriate for the stock and harvest control rules takes into account major features of the species' biology and the nature of the fishery. Such a management strategy incorporates the precautionary approach while also taking uncertainty into account and evaluating stock status relative to reference points, as these measures have been shown to be robust (modified from MSC 2010). As an example, if management is based on Total Allowable Catch, these limits are set below MSY and/or scientifically advised levels, accounting for uncertainty, and lowered if $B<B_{\text {MSY }}$. However, alternatives to TAC-based management, such as area-based (closures), 3S (size, sex and/or season limitations) or other appropriate methods may also apply (Appendix 3).

## Historic high:

Refers to near-virgin biomass, or highest recorded biomass, if biomass estimates predate the start of intensive fishing. If a fishery has been historically depleted and then rebuilt, the rebuilt biomass is not considered a "historic high" even though it may be higher than historical levels.

## Inherent vulnerability:

A stock's vulnerability to overfishing based on inherent life history attributes that affect the stock's productivity and may impede its ability to recover from fishing impacts. All sea turtles, marine mammals, and seabirds are considered "highly vulnerable". Marine invertebrates' vulnerability is based on the average of several attributes of inherent productivity.

One of the first key papers on this subject (Musick 1999) summarizes the results of an American Fisheries Society (AFS) workshop on the topic and offers proposed low, medium and high "productivity index parameters" (for marine fish species) based on available life history information: the intrinsic rate of increase $r$, the von Bertalanffy growth function $k$, fecundity, age at maturity and maximum age. Notably, although a species' intrinsic rate of increase is identified as the most useful indicator, it is also difficult to estimate reliably and is often unavailable (Cheung et al. 2005). To enable more timely and less data-intensive and costly identification of vulnerable fish species, Cheung et al. (2005) used fuzzy logic theory to develop an index of the intrinsic vulnerability of marine fishes based on life history parameters: maximum length, age at first maturity, longevity, von Bertalanffy growth parameter K, natural mortality rate, fecundity, strength of spatial behavior and geographic range (input variables). The index also uses heuristic rules defined for the fuzzy logic functions to assign fish species to one the following groups: very high, high, moderate or low level of intrinsic vulnerability.

In this framework, intrinsic vulnerability is also expressed by a numerical value between 1 and 100 with 100 being most vulnerable. This index of intrinsic vulnerability was then applied to over 1300 marine fish species to assess intrinsic vulnerability in the global fish catch (Cheung et al. 2007). FishBase, the online global database of fish, uses the numerical value from this index as a "vulnerability score" on the profiles of fish species for which it has been evaluated (Froese and Pauly 2010). Formerly, Seafood Watch used the FishBase vulnerability score to determine inherent vulnerability of fish species.

## Large portion of the stock is protected:

At least $50 \%$ of the spawning stock is protected, for example through size/sex/season regulations or the inclusion of greater than $50 \%$ of the species' habitat in marine reserves. Future guidance will improve the integration of marine reserve science into the criteria, based on ongoing research.

## Lenfest Forage Fish Task Force Recommendations:

In the 2012 report "Little Fish, Big Impact: Managing a Crucial Link in Ocean Foodwebs," the Lenfest Forage Fish Task Force (LFFTF) recommended the following precautionary buffers for forage fisheries: for fisheries with an intermediate level of information (which will include most well-managed forage fisheries), there must be at least $40 \%$ of virgin or unfished biomass ( $B_{0}$ ) left in the water, and fishing mortality should be no higher than $50 \%$ of $F_{\text {msy }}$. Low information fisheries should leave at least $80 \%$ of $B_{0}$ in the water. High information fisheries (which have a high information not just on the fished stock, but the full ecosystem), may exceed these reference points if justified by the science, but in no case should fishing mortality exceed $75 \%$ of $F_{\text {Msy }}$ or biomass fall below $30 \%$ of $B_{0}$. (see pages 90-91 of Lenfest Forage Fish Task Force guidelines and pages 8-9 of the Lenfest summary document). SFW incorporates the LFFTF recommendations into Criterion 4; they were formerly incorporated into Criterion 1 and 2.

## Likelihood:

Likely: 60\% chance or greater, when quantitative data are available; may also be determined according to expert judgment and/or plausible argument (modified from MSC 2010 and based on guidance from MSC FAM Principle 2).

Probable: Greater than 50\% chance; can be based on quantitative assessment, plausible evidence or expert judgment. Examples of "probable" occurrence for fishing mortality:

There may be some uncertainty or disagreement among various models; fishing mortality may be above $75 \%$ of a sustainable level and/or catch may be above $75 \%$ of a sustainable catch level (e.g., MSY) for stocks at $\mathrm{B}_{\mathrm{Ms}}$.

## Main species:

A species is included in the assessment as a main species if:

- A common component of the catch (as guidance, $>5 \%$ of the catch in most cases), or
- Overfished, endangered, threatened, undergoing overfishing, or otherwise a species of concern, where catch occurs regularly and may significantly contribute to the conservation concern (i.e. more than a negligible and/or sporadic level of catch) (as guidance, mortality of the species caused by this fishery is $>5 \%$ of a sustainable level), or
- Fishery is one of the main sources of mortality for the species, including bait species if known (as guidance, approx. $20 \%$ or more of total mortality).
- If bait species are unknown and no other main species have been identified, then add "unknown finfish" with an abundance score of 3 and a fishing mortality score of 3 .


## Managed appropriately:

Management uses best available science to implement policies that minimize the risk of overfishing or damaging the ecosystem, taking into account species vulnerability along with scientific and management uncertainty.

## Management Strategy

A management strategy or procedure or harvest strategy is the sum of all the management measures selected to achieve the biological, ecological, economic and social objectives of the fishery. SFW focuses on the biological and ecological objectives of management in the Fisheries Standard. Management or harvest strategies generally include: management objectives, a monitoring program, indicators of the fishery's status and population health, a method to assess those indicators, and harvest control rules. Robust harvest strategies are tested through a process called Management Strategy Evaluation before they are implemented.

## Management Strategy Evaluation

Definition from: Punt, Butterworth, de Moor, De Oliveira and Haddon (2016). Management strategy evaluation: Best practices. Fish and Fisheries 17(2): 303-334.
Figure from: Siple, Essington, and Plagányi (2019). Forage fish fisheries management requires a tailored approach to balance trade-offs. Fish and Fisheries, 20(1), 110-124.

Management strategy evaluation (MSE) involves using simulation to compare the relative effectiveness for achieving management objectives of different combinations of data collection schemes, methods of analysis and subsequent processes leading to management actions. MSE can be used to identify a 'best' management strategy among a set of candidate strategies, or to determine how well an existing strategy performs. The ability of MSE to facilitate fisheries management achieving its aims depends on how well uncertainty is represented, and how effectively the results of simulations are summarized and presented to the decision-makers. Key challenges for effective use of MSE therefore include characterizing objectives and uncertainty, assigning plausibility ranks to the trials considered, and working with decision-makers to interpret and implement the results of the MSE.


From Siple et al. 2019

## Negligible:

Mortality is insignificant or inconsequential relative to a sustainable level of total fishing mortality (e.g., MSY or PBR); less than or equal to $5 \%$ of a sustainable level of fishing mortality

## No management:

A fishery with no rules or standards for regulating fishing catch, effort or methods. Management does not need to be enforced through government regulation or official management agencies but may also include voluntary action taken by the fishery, as long as there is general compliance.

## Non-native species:

A non-native organism is an organism occurring outside its natural past or present range and dispersal potential, including any parts of the organism that might survive and subsequently reproduce, whose dispersal to the non-native area is caused by direct human action (e.g., introduced through ballast water or intentional translocation of organisms to a new area, but not including indirect anthropogenic effects, such as range shifts due to climate change). Modified from Falk-Petersen et al 2006.

## Overfished:

A stock is considered "overfished" when exploited past an explicit limit where abundance is considered too low to ensure safe reproduction. In many fisheries, the term "overfished" is used when biomass has been estimated below a biological reference point used to signify an "overfished condition". The stock may remain overfished (i.e., with a biomass well below the agreed limit) for some time even though fishing mortality may have been reduced or suppressed (FAO Glossary). Classification as "overfished" or "depleted" (including IUCN listing as "near threatened," "special concern," and "vulnerable") is based on evaluation by the management agency and/or FAO, but an analyst can use judgment to override this classification, especially where the classification may be out of date as long as there is scientific justification for doing so. Inclusion in the "overfished" category based on designations such as "stock of concern" are determined on a case-by-case basis, as such terms are not used consistently among management agencies. Stocks should be classified as "overfished" if the stock is believed to be at such a low level of abundance that reproduction is impaired or is likely to be below an appropriate limit reference point. Marine mammals classified as "depleted" under the Marine Mammal Protection Act also fall into this category if not listed as endangered or threatened. Stocks that are most likely (>50\% chance) below the level where recruitment or productivity is impaired are also considered "overfished". Note: Official IUCN listings should be overridden by more recent and/or more specific classifications where available (e.g., NMFS stock assessment showing that a stock is above target levels).

## Overfishing:

A generic term used to refer to a level of fishing effort or fishing mortality such that a reduction of effort would, in the medium term, lead to an increase in the total catch; or, a rate or level of fishing mortality that jeopardizes the capacity of a fishery to produce the maximum sustainable yield on a continuing basis. For long-lived species, overfishing (i.e., using excessive effort) starts well before the stock becomes overfished. Overfishing, as used in the Seafood Watch ${ }^{\circledR}$ criteria, can encompass biological or recruitment overfishing (but not necessarily economic or growth overfishing).

- Biological overfishing: Catching such a high proportion of one or all age classes in a fishery as to reduce yields and drive stock biomass and spawning potential below safe levels. In a surplus production model, biological overfishing occurs when fishing levels are higher than those
required for extracting the Maximum Sustainable Yield (msy) of a resource and recruitment starts to decrease.
- Recruitment overfishing: When the rate of fishing is (or has been) high enough to significantly reduce the annual recruitment to the exploitable stock. This situation is characterized by a greatly reduced spawning stock, a decreasing proportion of older fish in the catch and generally very low recruitment year after year. If prolonged, recruitment overfishing can lead to stock collapse, particularly under unfavorable environmental conditions.
- Growth overfishing: Occurs when too many small fish are being harvested too early through excessive fishing effort and poor selectivity (e.g., excessively small mesh sizes), and the fish are not given enough time to grow to the size at which maximum yield-per-recruit would be obtained from the stock. Reduction of fishing mortality among juveniles, or their outright protection, would lead to an increase in yield from the fishery. Growth overfishing occurs when the fishing mortality rate is above $F_{\max }$ (in a yield-per-recruit model). This means that individual fish are caught before they have a chance to reach their maximum growth potential. Growth overfishing, by itself, does not affect the ability of a fish population to replace itself.
- Economic overfishing: Occurs when a fishery is generating no economic rent, primarily because an excessive level of fishing effort is applied in the fishery. This condition does not always imply biological overfishing.
(FAO Glossary; NOAA 1997)


## Precautionary approach:

The precautionary approach involves the application of prudent foresight. Taking account of the uncertainties in fisheries systems and considering the need to take action with incomplete knowledge, the precautionary approach requires, inter alia: (i) consideration of the needs of future generations and avoidance of changes that are not potentially reversible; (ii) prior identification of undesirable outcomes and measures to avoid or correct them promptly; (iii) initiation of any necessary corrective measures without delay and on a timescale appropriate for the species' biology; (iv) conservation of the productive capacity of the resource where the likely impact of resource use is uncertain; (v) maintenance of harvesting and processing capacities commensurate with estimated sustainable levels of the resource and containment of these capacities when resource productivity is highly uncertain; (vi) adherence to authorized management and periodic review practices for all fishing activities; (viii) establishment of legal and institutional frameworks for fishery management within which plans are implemented to address the above points for each fishery, and (ix) appropriate placement of the burden of proof by adhering to the requirements above (modified from FAO 1996).

## Productivity is maintained/not impaired:

Fishing activity does not impact the stock, either through reduced abundance, changes in size, sex or age distribution, or reduction of reproductive capacity at age, to a degree that would diminish the growth and/or reproduction of the population over the long-term (multiple generations).

## Productivity-susceptibility analysis (PSA):

Productivity-susceptibility analysis was originally developed to assess the sustainability of bycatch levels in Australia's Northern Prawn fishery (Patrick et al. 2009) and has since been widely applied to assess vulnerability to fishing mortality for as number of fisheries worldwide. Productivity-susceptibility
analysis is used by NOAA and the Australian Commonwealth Scientific Industrial Research Organization (CSIRO) to inform fisheries management. It also constitutes the basis of the risk-based framework used to evaluate data-limited fisheries for both fish and invertebrates under the Marine Stewardship Council Fishery Assessment Methodology (MSC FAM). The PSA approach allows the risk of overfishing to be assessed for any species based on predetermined attributes, even in the most data-limited situations.

The exact sets of productivity and susceptibility attributes vary between PSA methodologies, and different weighting of attributes can be employed based on relative contextual importance. Additionally, scoring thresholds can vary depending on the context in which PSA is employed. In the US methodology, productivity is defined as the capacity for a stock to recover once depleted, which is largely a function of the life history characteristics of the species. Generally, productivity attributes are similar to the life history parameters used for the above index of intrinsic vulnerability.

While PSA analysis is a widely accepted approach for evaluating risk of overexploitation of a fished species, for the purposes of Seafood Watch assessments it is useful to separate the productivity attributes - which are intrinsic to a species and neither dependent on nor influenced by fishery practices-from the susceptibility attributes. Fisheries may influence the susceptibility of impacted stocks through the choice of gear, bait species, hook design, mesh size, area or seasonal closures, and other management measures. In addition, where detailed information on fishing mortality (e.g., estimates of F or harvest rates) is available, these data provide a more complete picture of the fishery impact that the susceptibility attributes are designed to predict.

## Reasonable timeframe (for rebuilding):

Dependent on the species' biology and degree of depletion, but generally within 10 years, except in cases where the stock could not rebuild within 10 years even in the absence of fishing. In such cases, a reasonable timeframe is within the number of years it would take the stock to rebuild without fishing, plus one generation, as described in Restrepo et al. (1998).

## Recent stock assessment:

As a rule of thumb, stock assessments or updates based on data collected within the last five years are considered to be recent. If the data used within an assessment are $>5$ years old, but $<10$ years old, and show that biomass is above target reference points, abundance should be scored as a low concern in most cases, but with consideration of trends and time series; e.g., if the population has been stable and was well above the TRP in the last assessment, and the species is not one that fluctuates greatly in abundance, and the fishery hasn't changed dramatically in recent years, a very low concern may be justified. If the stock assessment is very out of date-as a rule of thumb, data are >10 years old-the stock status should be considered unknown and rated accordingly. It may be considered unknown even when the assessment is less than 10 years old in circumstances where the stock was previously very close to reference points or is very dynamic. If the most recent stock assessment was not accepted by the relevant scientific body for any reason, the stock should be considered unknown. If older data are used within an assessment which is ultimately approved by a relevant scientific body (knowing that the data are old), the results can be considered appropriate and scored a low concern; this outcome is expected to be rare.

## Recruitment is impaired:

Fishing activity impacts the stock-either through reduced abundance, changes in size, sex or age distribution, or reduction of reproductive capacity at age-to a degree that will diminish the growth
and/or reproduction of the population over the long-term (multiple generations), or the stock is below an appropriate limit reference point, if one is defined.

## Regularly monitored:

Fishery-independent surveys of stocks, or other reliable assessments of abundance, are conducted at least every three years.

## Relevant legal requirements:

These include state, national and international laws which pertain to the fishery.

## Reliable data:

Data produced or verified by an independent third party. Reliable data may include government reports, peer-reviewed science, audit reports, etc. Data are not considered reliable if significant scientific controversy exists over the data, or if data are old or otherwise unlikely to represent current conditions (e.g., survey data is several years old and fishing mortality has increased since the last survey).

## Resilience:

Resilience is the ability to recover from or withstand perturbation. In the Standard for Fisheries, Seafood Watch discusses resilience at the following levels:

1. Species - the ability of a species to recover from or withstand perturbations based on its life history traits and strategies.
2. Stock - the ability of a stock to recover from or withstand perturbations as a result of management strategies that build capacity to buffer the impacts of expected/predicted or unexpected changes (see Appendix 3).
3. Habitats - the ability of the benthos to recover from or withstand perturbations from fishing activity, which may be a result of management strategies that mitigate those impacts (see Appendix 6).
4. Ecosystem - The capacity of an ecosystem to absorb perturbations while retaining its essential structure, function and feedbacks. Speed of recovery following perturbation is a common empirical metric of resilience (adapted from EDF's FISHE Glossary).
5. Management in the context of climate change (relevant to all species but which is critical for species (like forage species) that undergo large shifts in productivity as a function of their life history characteristics) - Resilient management in this context includes precautionary, efficient, and responsive policies that address climate uncertainty with consideration of species' life history characteristics.

## Species of concern:

Species about which management has some concerns regarding status and threats, but for which insufficient information is available to indicate a need to list the species as endangered. In the U.S., this may include species for which NMFS has determined, following a biological status review, that listing under the ESA is "not warranted," pursuant to ESA section 4(b)(3)(B)(i), but for which significant concerns or uncertainties remain regarding their status and/or threats. Species can qualify as both "species of concern" and "candidate species" (http://www.nmfs.noaa.gov/pr/glossary.htm\#s). In addition, marine mammal stocks listed as "strategic" under the Marine Mammal Protection Act are
classified as species of concern. The terms "species of concern" or "stock of concern" are used similarly by other federal and state management bodies.

## Stakeholder input:

A stakeholder is an individual, group or organization that has an interest in, or could be affected by, the management of a fishery (modified from MSC 2010). Stakeholder input may include: involvement in all key aspects of fisheries management from stock assessment and setting research priorities to enforcement and decision-making. In addition, stakeholders may take ownership of decisions and greater responsibility for the wellbeing of individual fisheries (Smith et al 1999). Effective stakeholder engagement requires that the management system has a consultation processes open to interested and affected parties and that roles and responsibilities of the stakeholders are clear and understood by all relevant parties (modified from MSC 2010).

## Stock:

A self-sustaining population that is not strongly linked to other populations through interbreeding, immigration or emigration. A single fishery may capture multiple stocks of one or multiple species. Stocks can be targeted or non-targeted, retained or discarded, or some combination thereof (e.g., juveniles are discarded, and adults are retained).

Ideally, the management unit of "stock" should correspond to the biological unit. However, often the fisheries management unit of "stock" may not be the same as the biological unit. If multiple biological stocks are managed as one, and there is insufficient information to assess the stock status of each biological stock, the management unit is assessed. This situation detracts from the fishery's "quality of information" score, as it makes it impossible to assess individual stocks' health. If management occurs on a finer scale than biological stocks so that multiple management unit stocks compose one interbreeding population, the health and abundance of the biological stock should be assessed as a whole, based on information aggregated across the management units. The effectiveness of management can be assessed at the finest scale for which meaningful and verifiable differences in management practice exist.

## Substantial contributor:

A fishery is a substantial contributor to impacts affecting a population, ecosystem or habitat if the fishery is a main contributor, or one of multiple contributors of a similar magnitude, to cumulative fishing mortality. An example of a fishery that is not a substantial contributor includes: catch of the species is a rare or minor component of the catch in this fishery and the fishery is a small contributor to cumulative mortality, relative to other fisheries, particularly because the fishery operates or is managed in a way that reduces its impact. However, if there has been a jeopardy determination for that stock in the fishery being assessed it should be considered a substantial contributor regardless of this definition. This applies to species assessed under Criterion 2 only. In exceptional cases, fishing mortality on a stock of concern may have been reduced to almost zero such that any fisheries interacting with a stock would be considered a substantial contributor to total fishing mortality; however, in these cases it is important to consider other sources of mortality, including natural mortality, which may be more significant. Where total fishing mortality is very low and there is evidence that other sources of mortality are impacting a stock, fishing mortality can be considered a non-substantial contributor. In order to assist in determining whether a fishery is a substantial contributor, please consider the decision tree below which aims to determine the level of impact of a fishery relative to a sustainable level, or other fishing impacts.


## Substantial proportion of habitat:

Refers to a condition when at least $20 \%$ of each representative habitat (where representative habitats can be delineated by substrate, bathymetry, and/or community assemblages), within both the range of the targeted stock(s) and the regulatory boundaries of the fishery under consideration (i.e., within the national EEZ for the fishery under consideration), is completely protected from fishing with gear types that impact the habitat in that fishery.

## Susceptibility

A stock's capacity to be impacted by the fishery under consideration, depending on factors such as the stock's likelihood to be captured by the fishing gear. The susceptibility score is based on tables from MSC's Productivity-Susceptibility Analysis framework (see Criterion 1.1). Examples of low susceptibility include: low overlap between the geographic or depth range of species and the location of the fishery; the species' preferred habitat is not targeted by fishery; the species is smaller than the net mesh size as an adult, is not attracted to the bait used, or is otherwise not selected by fishing gear; or strong spatial protection or other measures in place specifically to avoid catch of the species.

## Sustainable level (of fishing mortality):

A level of fishing mortality that will not reduce stock below the point where recruitment is impaired, i.e., above $F$ reference points, where defined. The $F$ limit reference points should be around either $\mathrm{F}_{\text {MSy }}$ or F35-40\% for moderately productive stocks; low productivity stocks like rockfish and sharks require $F$ in the range of F50-60\% or lower. Higher F values require a strong scientific rationale. For example, the F reference points are limit reference points, so buffers should be used to ensure that fishing mortality does not exceed these levels. Where F is unknown but MSY (or another appropriate reference point) is estimated, fishing mortality at least $25 \%$ below MSY (or another appropriate reference point) is considered a sustainable level (for fisheries that are at or above $\mathrm{B}_{\mathrm{MSY}}$ ).

## Uncertainty:

Most data available to fisheries scientists contains uncertainty. Typically, we are dealing with estimates of catch size, population biomass and levels of natural and fishing mortality. As a result, stock assessments based on these data will also include uncertainty, which needs to be considered when interpreting said results for the purpose of a Seafood Watch assessment. In some cases, the uncertainty has been quantified, for example as a standard deviation or standard error of a biomass estimate. In such cases, these values can be used to determine whether the estimate is above or below a reference point. For example, where an estimate of biomass is greater than the target reference point we would expect factor 1.1 to be scored a very low concern, however if uncertainty is such that the lower limit of the standard deviation falls below the target reference point, a low concern is a more appropriate score, in order to account for the uncertainty in the stock assessment result.

## Up-to-date data/stock assessment:

Complete stock assessments are not required every 1-5 years, but stocks should be regularly monitored at least every 1-5 years, and stock assessments should be based on data not more than five years old. Data may be collected by industry, but analysis should be independent.

## Very low levels of exploitation (e.g., experimental fishery):

A fishery is vastly under-exploited or is being conducted experimentally to collect data or gauge viability, such that exploitation rates are far below sustainable yields (e.g., $20 \%$ or less of sustainable take). Alternatively, when no other information is available, exploitation levels may be considered very low if a fishery falls into the "low" category for all "susceptibility" questions under the Productivity-Susceptibility Analysis.

## References

Alverson, D.L. Freeberg, M.H., Pope, J.G. and S.A. Murawski. 1994. A global assessment of fisheries bycatch and discards. FAO Fisheries Technical Paper. No. 339. Rome, FA. 233p

Araki, H, and C. Schmid. 2010. Is hatchery stocking a help or harm? Evidence, limitations and future directions in ecological and genetic surveys. Aquaculture 308: S2-S11

Babcock, E., E. Pikitch, and C. Hudson. 2004. How much observer coverage is enough to adequately estimate bycatch? Oceana, Washington D.C.

Botsford, L. W., and A. M. Parma. 2005. Uncertainty in Marine Management. Pages 375-392 in E. A. Norse and L. B. Crowder, editors. Marine Conservation Biology: The Science of Maintaining the Sea's Biodiversity. Island Press, Washington, DC.

Caddy, J. F. 2004. Current usage of fisheries indicators and reference points, and their potential application to management of fisheries for marine invertebrates. Canadian Journal of Fisheries and Aquatic Science 61:1307-1324.

Chaffee C., S. Daume S, L. Botsford, D. Armstrong, and S. Hanna. 2010. Oregon Dungeness Crab Fishery Final Report with Certification Decision. Ver. 4, 27 October 2010. Reports available at: https://fisheries.msc.org/en/fisheries/oregon-dungeness-crab/@@assessments

Cheung, W. W. L., T. J. Pitcher, and D. Pauly. 2005. A fuzzy logic expert system to estimate intrinsic extinction vulnerabilities of marine fishes to fishing. Biological Conservation 124:97-111.

Cheung, W. W. L., R. Watson, T. Morato, T. J. Pitcher, and D. Pauly. 2007. Intrinsic vulnerability in the global fish catch. Marine Ecology Progress Series. 333:1-12.

Clark, W.G. 1991. Groundfish exploitation rates based on life history parameters. Canadian Journal of Fisheries and Aquatic Sciences 48: 734-750.

Clark W. G. 2002. F35\% revisited ten years later. North American Journal of Fisheries Management 22:251-257.

DFO. 2009. A fishery decision-making framework incorporating the Precautionary Approach. Fisheries and Oceans Canada. Available at: https://www.dfo-mpo.gc.ca/reports-rapports/regs/sff-cpd/precaution-back-fiche-eng.htm

Dulvy, K., Y. Sadovy, and J. D. Reynolds. 2003. Extinction vulnerability in marine populations. Fish and Fisheries 4(1):25-64.

Environmental Defense Fund (EDF). 2016. FISHE - Framework for Integrated Stock and Habitat Evaluation. Available at: http://fishe.edf.org/.

FAO. 1996. Precautionary Approach to Capture Fisheries and Species Introductions. FAO Technical Guidelines for Responsible Fisheries, 2: 54 p.

FAO Fisheries Glossary. Accessed December 8, 2010. Available at: http://www.faoo.org/fi/glossary/.

Falk-Petersen, J., T. Bøhn and O. T. Sandlund. 2006. On the numerous concepts in invasion biology. Biological Invasions 8:1409-1424.

Field, J., Cope, J. and Key, M. 2010. A descriptive example of applying vulnerability evaluation criteria to California nearshore finfish species. Managing Data-Poor Fisheries: Case Studies, Models \& Solutions, 1, 235-246.

Finnegan, S., S. C. Wang, A. G. Boyer, M. E. Clapham, Z. V. Finkel, M. A. Kosnik, M. Kowalewski, R. A. J. Krause, S. K. Lyons, C. R. McClain, D. McShea, P. M. Novack-Gottshall, R. Lockwood, J. Payne, F. Smith, P. A. Spaeth, and J. A. Stempien. 2009. No general relationship between body size and extinction risk in the fossil record of marine invertebrates and phytoplankton. in GSA Annual Meeting, Portland, OR.

Foley, M.M., Halpern, B.S., Micheli, F. et al. 2010. Guiding ecological principles for marine spatial planning. Marine Policy 34: 955-966.

Froese, R., and D. Pauly. 2010. FishBase. World Wide Web electronic publication. www.fishbase.org.

Goodman, D., M. Mangel, G. Parkes, T. Quinn, V. Restrepo, T. Smith and K. Stokes. 2002. Scientific Review of The Harvest Strategy Currently Used in The BSAI and GOA Groundfish Fishery Management Plans, North Pacific Fishery Management Council, Anchorage, AK. 153 p. Available at: https://pdfs semanticscholar.org/30ab/c09298fc4fc13d017c697511aa1bd11bf739.pdf

Hobday, A., M. J. Tegner, and P. L. Haaker. 2001. Over-exploitation of a broadcast spawning marine invertebrate: Decline of the white abalone. Reviews in Fish Biology and Fisheries 10:493-514.

Kelleher, K. 2005. Discards in the world's marine fisheries. An update. FAO Fisheries Technical Paper. No. 470. Rome, FAO. 131p.

Kostow, K. 2009. Factors that contribute to the ecological risks of salmon and steelhead hatchery programs and some mitigating strategies. Reviews in Fish Biology and Fisheries 19:1, 9-31.

Mace, P.M. and M.P. Sissenwine. 1993. How much spawning per recruit is enough? pp 101-118 in S.J. Smith, J.J. Hunt and D.Revered (eds.) Risk Evaluation and Biological Reference Points for Fisheries Management. Canadian Special Publication of Fisheries and Aquatic Sciences 120. National Research Council of Canada.

Marine Stewardship Council (MSC). 2014. MSC Fisheries Certification Requirements and Guidance, version 2.0. Marine Stewardship Council, London, UK. 120 p. Available at: https://www.msc.org/documents/scheme-documents/fisheries-certification-scheme-documents/fisheries-certification-requirements-version-2.0.

Marine Stewardship Council (MSC). 2010. Fisheries Assessment Methodology and Guidance to Certification Bodies, version 2.1, including Default Assessment Tree and Risk-Based Framework. Marine Stewardship Council, London, UK. 120 p. Available at:
http://opc.ca.gov/webmaster/ftp/project_pages/CSSI/MSC_Fisheries_Assessment_Methodology_v2\%2 01.pdf

McKinney, M. L. 1997. Extinction vulnerability and selectivity: Combining ecological and paleontological views. Annual Review of Ecology, Evolution and Systematics 28:495-516.

MSC. 2009. Marine Stewardship Council Fisheries Assessment Methodology and Guidance to Certification Bodies Including Default Assessment Tree and Risk-Based Framework

Myers R. A., Bowen K. G., and N.J. Barrowman. 1999. Maximum reproductive rate of fish at low population sizes. Canadian Journal of Fisheries and Aquatic Sciences 56:2404-2419.

Musick, J. A. 1999. Criteria to define extinction risk in marine fishes. Fisheries 24(12):6-14.

National Marine Fisheries Service (NMFS). 2013. U.S. National Bycatch Report First Edition Update 1 [L. R. Benaka, C. Rilling, E. E. Seney, and H. Winarsoo, Editors]. U.S. Dep. Commer., 57 p. Available at: https://www.st.nmfs.noaa.gov/observer-home/first-edition-update-1

NOAA. 1997. NOAA Fisheries Strategic Plan. 48p.

O'Malley, S. L. 2010. Predicting vulnerability of fishes. Masters Thesis. University of Toronto, Toronto.

Paine, R.T. 1995. "A Conversation on Refining the Concept of Keystone Species". Conservation Biology 9 (4): 962-964.

Patrick, W. S., P. Spencer, O. Ormseth, J. Cope, J. Field, D. Kobayashi, T. Gedamke, E. Cortés, K. Bigelow, W. Overholtz, J. Link, and P. Lawson. 2009. Use of productivity and susceptibility indices to determine stock vulnerability, with example applications to six U.S. fisheries. U.S. Department of Commerce, NOAA Tech. Memo. NMFS-F/SPO-101, Seattle, WA.

Pinsky, M.L., Jensen, O.P., Ricard and D., Palumbi, S.R. 2011. Unexpected patterns of fisheries collapse in the world's oceans. Proceedings of the National Academy of Sciences 108 (20) 8317-8322.

Pikitch, E., P.D. Boersma, I.L. Boyd, D.O. Conover, P. Cury, T. Essington, S.S. Heppell, E.D. Houde, M. Mangel, D. Pauly, É. Plagányi, K. Sainsbury, and R.S. Steneck. 2012. Little Fish, Big Impact: Managing a Crucial Link in Ocean Food Webs. Lenfest Ocean Program. Washington, DC. 108 pp.

Plagányi, E.E and T.E. Essington. 2014. When the SURFs up, forage fish are key. Fisheries Research 159:68-74.

Punt, A. E., Butterworth, D. S., de Moor, C. L., De Oliveira, J. A. A. and M. Haddon. 2016. Management strategy evaluation: Best practices. Fish and Fisheries 17(2): 303-334.

Restrepo, V.R. and J. E. Powers. 1998. Precautionary control rules in US fisheries management: specification and performance. ICES Journal of Marine Science 56: 846-852.

Restrepo, V.R., G.G. Thompson, P.M. Mace, W.L. Gabriel, L.L. Low, A.D. MacCall, R.D. Methot, J.E. Powers, B.L. Taylor, P.R. Wade and J.F. Witzig. 1998. Technical Guidance on the Use of Precautionary

Approaches to Implementing National Standard 1 of the Magnuson-Stevens Fishery Conservation and Management Act. NOAA Technical Memorandum NMFS-F/SPO-137. 54pps.

Siple, M. C., Essington, T. E., \& É Plagányi. (2019). Forage fish fisheries management requires a tailored approach to balance trade-offs. Fish and Fisheries, 20(1), 110-124.

## Appendices

## Appendix 1 - Further guidance on interpreting the health of stocks and fishing mortality

The tremendous variability among fisheries makes it impossible to define specific appropriate reference points that would be applicable to all assessed fisheries. Instead, criteria are based on the commonly accepted management goal that target biomass should be at or above the point where yield is maximized, and management should ensure a high probability that biomass is at or above a limit reference point (where recruitment or productivity of the stock would be impaired). Three common types of reference points are MSY-based, SPR-based, and ICES reference points. However, other reference points may be used in some fisheries, and should be evaluated in accordance with the management goal articulated above.

## Evaluating Abundance

## MSY-based reference points

While the concept of MSY is far from perfect, MSY-based biomass and fishing mortality reference points are commonly used in some of the most well managed fisheries around the world. When applied appropriately, these reference points are an important tool for maintaining stock productivity in the long term. However, without properly accounting for scientific and management uncertainty, maintaining a stock at $\mathrm{B}_{\text {MSY }}$ (the biomass corresponding to MSY) and harvesting at MSY runs a high risk of unknowingly either overshooting MSY or allowing biomass to drop below $B_{\text {MSY }}$ without reducing harvest rates and thus inadvertently overharvesting (Roughgarden and Smith 1996; Froese et al. 2010). The risk of impacts from inadvertent overharvesting increases with increased uncertainty and with increased inherent vulnerability of the targeted stock. To account for these interactions, the guidance provided for assessing stock health and fishing mortality is based on MSY reference points but requires high scientific confidence that biomass is above target levels and that fishing mortality is below MSY.

Proxies for $\mathrm{B}_{\text {MSY }}$ are acceptable if shown to be conservative relative to $\mathrm{B}_{\text {Msy }}$ for that stock, or if they fit within the guidelines for appropriate target level** Where $B_{\text {MSY }}$ or other appropriate reference points are not known or are not applicable, the stock/population health criteria can be interpreted using relevant indicators that are appropriate as targets and safe limits for abundance of the species (e.g., escapement relative to escapement goals can be evaluated in lieu of biomass relative to limit reference points).

## ICES reference points

The current objective of ICES advice is to achieve MSY through Ecosystem Based Fisheries Management (ICES 2018).

Traditionally, the ICES reference points $\mathrm{F}_{\mathrm{PA}}, \mathrm{F}_{\mathrm{LI}}, \mathrm{B}_{\mathrm{PA}}$, and $\mathrm{B}_{\mathrm{LI}}$ utilized were not equivalent to MSY -based reference points. In fact, comparisons demonstrated that $F_{P A}$ is typically above $F_{M S Y}$ and $B_{P A}$ is typically below $B_{\text {MSY }}$, such that MSY-based reference points are generally more conservative (ICES 2010). In many cases, $B_{P A}$ is well below $B_{M S Y}$ and even below $1 / 2 B_{M S Y}$ (Kell et al. 2005). Therefore, guidance for evaluating stock health using $B_{P A}$ and fishing mortality using $F_{P A}$ is conservative, accounting for the difference between these reference points and MSY-based reference points. ICES plans to has transitioned to an MSY-based approach (ICES 2018) Not all stock assessments may have been updated
to this new approach however; therefore_if $\mathrm{B}>\mathrm{B}_{\text {PA }}$ or $\mathrm{F}<\mathrm{F}_{P \mathrm{PA}}$, the stock should score as a moderate concern, unless a good reason exists to justify a "low concern" score for abundance (i.e., either the reference points have been shown to be conservative or the biomass is well above reference points).

## Proxies

For many fisheries, $\mathrm{F}_{\text {ms }}$ and $\mathrm{B}_{\text {Msy }}$ are unknown, and proxies are often used. Most commonly, biomass proxies are based on the percent of unfished or virgin biomass ( $\mathrm{B}_{0}$ ). Fishing mortality proxies are often based on spawning potential ratio (SPR).

Commonly used and acceptable biomass reference points are typically $35-40 \%$ of $\mathrm{B}_{0}$ for most stocks (Clark 1991; NZ Ministry of Fisheries 2008). This target may vary according to stock productivity; however, justifications for lower target levels are often based on assumptions about "steepness ${ }^{8 "}$ that may be highly uncertain or poorly understood. It is now recognized that stock targets lower than approximately $30-40 \%$ of $B_{0}$ are increasingly difficult to justify (NZ Ministry of Fisheries 2008). For these targets to be considered appropriate reference points, solid scientific justification is required. In addition, stocks reduced to this target level or below (equivalent to removing more than 60-70\% of the stock's biomass) would be unlikely to achieve the ecosystem-based management goal of allowing a stock to fulfill its ecological role and should be scored accordingly under ecosystem-based management.

Alternatively, when unfished biomass cannot be estimated, appropriate biomass reference points may be based on the equilibrium biomass achieved using appropriate fishing mortality reference points, as described below.

A large body of scientific literature addresses appropriate fishing mortality reference points based on spawner biomass per recruit (SPR). Ideally, these should be shown through scientific analysis to be at or above replacement \%SPR (the threshold level of SPR necessary for replacement) for the species, based on its productivity and S-R relationship (Mace and Sissenwine 1993). However, for many species this analysis will not be available. In these cases, guidance is based on the conclusions of numerous analyses demonstrating that, in general, $\mathrm{F}_{35-40 \%}$ (the fishing mortality rate that reduces the SPR to $35-40 \%$ of unfished levels) is appropriate for species with moderate vulnerability, while a more conservative fishing mortality rate of about $\mathrm{F}_{50-60 \%}$ is needed for highly vulnerable species such as rockfish and sharks (Botsford and Parma 2005; Mace and Sissenwine 1993; Clark 2002; Myers et al. 1999; Goodman et al. 2002).

## Data-limited reference points and other metrics

In the absence of stock assessments and MSY-based reference points, the stock health can be evaluated based on CPUE, trends in abundance and size structure, and/or simple, easy to calculate reference points such as fraction of lifetime egg production (FLEP) (equivalent to spawning potential ratio, SPR) and an array of other data-limited assessment approaches. Other data-limited or alternative assessment techniques that provide evidence that stocks are healthy (i.e., productivity and reproduction are not impaired) may be used in place of or to supplement reference points. Examples of simple metrics which provide evidence that a stock's productivity may have shifted include the Froese length-based indicators (2004) and their modifications by Cope and Punt (2009). FISHE (EDF 2016: available in the FISHE Resources section at http:///fishe.edf.org/.) provides information on these and other data-limited

[^5]metrics. Dowling et al. 2019 provides a compendium of many existing data-limited approaches (see Table 1) and describes their uses, assumptions, and limitations. Appendix 7 provides guidance to SFW analysts on how to evaluate data-limited metrics when scoring abundance. SFW holds data-limited fisheries to the same standard of likelihood as data-rich fisheries when stocks are above a level where recruitment would be impaired and fishing mortality is at or below a sustainable level of harvest.

- Examples of evidence that a stock is above the point where recruitment or productivity is impaired, i.e., an appropriate limit reference point, include:
- the current lifetime egg production (LEP) or spawning per recruit (SPR) is above an appropriate SPR or Fraction of Lifetime Egg Production (FLEP)-related reference point;
- spawning potential is well protected (e.g., females are not subject to mortality, and it can be shown or inferred that fertilization is not reduced);
- quantitative analyses conducted by fishery scientists under transparent guidelines indicate sufficient stock;
- Strong, quantitative scientific evidence from the fishery under consideration is required to consider a stock a "very low concern" for abundance. When limited data are available from the fishery, analogy with similar systems, qualitative expert judgments and/or plausible arguments may be used to consider the stock as "low concern";
- Use of CPUE requires the absence of hyperstability, that CPUE is proportional to abundance (or adjusted), and that there have been no major changes in technology;
- The LEP can be estimated from length frequency data from both unfished (or marine reserve) and current populations and does not require catch-at-age data. Reference points based on FLEP should be considered limit reference points.

For "very low concern" for abundance, there must be no evidence that productivity has been reduced through fisheries-induced changes in size or age structure, size or age at maturity, sex distribution, etc. SPR-based and MSY-based reference points should account for these changes as they are based on productivity of the stock rather than simple abundance. If the metric considers abundance only, or if there is evidence that productivity has been reduced through shifts in age, size or sex distributions, the stock cannot be rated higher than "low concern." Moreover, "very low concern," for abundance stock assessments or updates should be no more than five years old; have been approved through an independent scientific peer review process; and include verified fishery dependent and fishery independent abundance data and accurate life history data. Biomass information must be estimated with low uncertainty. In cases where these qualifications may not apply, the analyst must adequately justify his/her reasoning.

## Evaluating Fishing Mortality

Evaluation of fishing mortality should reflect the mortality caused by the fishery, but in the context of whether cumulative impacts on the species (including mortality from other fisheries) are sustainable. When determining whether a fishery is a substantial contributor, err on the side of caution. Unknown or missing data are grounds for classification as a substantial contributor.

## Reference points

Generally, species should be managed with reference points that fit the definition of a sustainable level of fishing mortality_and/or an appropriate SPR or Fraction of Lifetime Egg Production (FLEP)-related reference point. Species that are not commercially fished or managed but make up non-target catch in the fishery will generally not have reference points defined. In lieu of reference points, these stocks should be evaluated relative to a level of mortality scientifically shown not to lead to depletion of the stock. For species with high vulnerability, the reference point must be demonstrated to be appropriate for that species' biology. As a rule of thumb, $\mathrm{F}_{40 \%}$ is not precautionary enough for high vulnerability species; $\mathrm{F}_{50 \%}$ or lower is more appropriate when using SPR-based proxies.

## ICES reference points

Because analysis has shown that the previously utilized ICES reference point $F_{P A}$ is typically above $F_{M S Y}$, ICES stocks using $\mathrm{F}_{\mathrm{PA}}$ as a reference point must be rated more conservatively than stocks using $\mathrm{F}_{\mathrm{MSr}}$. If F > $\mathrm{F}_{\mathrm{PA}}$, rate the stock as "high concern". If F < $\mathrm{F}_{\mathrm{PA}}$, rate the stock as "moderate concern," unless there is additional evidence that $F$ is below a sustainable level such as $F_{\text {MSY }}$. These reference points may appear in older assessments that have not yet been updated utilizing the MSY approach.

## Data-limited stocks

When no formal reference points are available (i.e., in data-limited fisheries), , fishing mortality could be considered a low concern if the fishery has a low likelihood of interacting with a non-target species due to low overlap between the species range and the fishery, or due to low gear selectivity for the species (resulting in low susceptibility; see below). Fishing mortality on target or non-target species may be considered a low concern if there is a very low level of exploitataion.

## Age of Assessment

If the stock assessment, or the data used within it, is greater than 10 years old then there is a high level of uncertainty associated with the result (with respect to how it reflects the current situation). In cases where $\mathrm{F}<\mathrm{F}_{\text {MSY }}$ (or appropriate reference point) and the data are greater than 10 years old, fishing mortality should be considered "unknown" or a moderate conservation concern. In all cases where F>Fmsy (or appropriate reference point), regardless of the age of assessment, fishing mortality should be scored as a high conservation concern.

## Appendix 2 - Matrix of bycatch impacts by gear type

The matrices in this appendix are used to determine the relative impact of a fishery on bycatch species of various taxa for fisheries where species and amounts of bycatch are not available or are incomplete. The matrices represent typical relative impacts of different fishing gear on various taxa based on the best available science. If there are data that indicate a specific fishery is operating differently from the standard operating procedures, the UBM can be overruled.

## Scoring abundance of unknown bycatch species:

Sea turtles, sharks, marine mammals, seabirds, and fish and invertebrate bycatch species from taxa known to be of high inherent vulnerability - including sharks, skates, rays, sturgeon, rockfish, grouper, corals, abalone and conch - should be scored as highly vulnerable, and thus a High Concern under the abundance factor (2.1). Other fish and invertebrates should generally be scored as a Moderate Concern, unless data exist that would indicate an alternative rating. For more guidance, see also "Additional guidance for scoring unknown bycatch species in Criterion 1.1/2.1 (Abundance)", below.

## Scoring fishing mortality of unknown bycatch species:

## Highly vulnerable marine megafauna (sea turtles, marine mammals, seabirds and sharks)

Updated tables for highly vulnerable taxa (sea turtles, marine mammals, seabirds, and sharks) now incorporate a regional component. We generated these values based on an extensive literature review ( 54 reports, peer-reviewed articles) to better reflect the array of bycatch issues that occur using the same gear types in different regions of the world, reflecting the regional susceptibility of the taxa to gear. Only the turtle matrix also incorporates reproductive values because the literature incorporates age-related information that was not available for the other taxa. We incorporated the effect of mitigation measures only to the extent that bycatch studies were of fisheries that used bycatch reduction techniques.

## Gear categories for Unknown Bycatch Matrices

| FAO Gear Category | FAO Methods | FAO <br> Abbreviation | MBA |
| :---: | :---: | :---: | :---: |
| DREDGES | Dredges (nei) | LN | Use this for all dredges |
| GILLNETS AND ENTANGLING NETS | Set gillnets | GNS |  |
|  | Drift gillnets (driftnets) | GND |  |
|  | Encircling gillnets | GNC | Use GND |
|  | Fixed gillnets (on stakes) | GNF | Use GNS |
|  | Trammel nets | GTR | Use GNS |
|  | Combined gillnets - trammel nets | GTN | Use GNS |
|  | Gillnets and entangling nets (nei) | GEN | If on bottom GNS, if not fixed, GND. Could include liftnets and reefnets |
| HOOKS AND LINES | Handlines and hand-operated pole-and lines | LHP |  |
|  | Mechanized lines and pole-and-lines | LHM |  |
|  | Set longlines | LLS | Bottom longlines, Buoy gear |
|  | Drifting longlines | LLD | Pelagic longline, Trotline |
|  | Trolling lines | LTL | Greenstick, Jig |
| MISCELLANEOUS GEARS | Harpoons | HAR |  |
|  | Diving | MDV |  |
| SURROUNDING NETS | Purse seines | PS | Dolphin set (D), Floating Object/whaleshark (F), Unassociated (U) |
|  | Surrounding nets (nei) | SUX | Lampara, non-tuna PS, Danish seine, suripera, ring nets |
| TRAPS | Pots | FPO | Crab rings |
| TRAWLS | Bottom trawls (nei) | TB | Small or large mesh bottom trawl, Magdalena - Artisanal bottom trawl, butterfly trawl |
|  | Midwater trawls (nei) | TM |  |

Sea Turtle Bycatch Susceptibility

| Region | Longline |  | Gillnet |  | Trawl |  | Dredge | Purse Seine |  |  |  | Other |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | LLS | LLD | GNS | GND | TB | TM | LN | SUX | PSF | PSD | PSU | FPO | HARI MDV | LTL LHP/LHM |
| Caribbean/Gulf of Mexico | 1 | 2 | 1 | 1 | 1 | 2 | 3 | 4 | 3 | -- | 4 | 4 | 5 | 5 |
| East Indian Ocean/Southeast Asia | 1 | 1 | 1 | 1 | 1 | 2 | 3 | 3 | 3 | -- | 4 | 5 | 5 | 5 |
| E. Pacific/Eastern Tropical Pacific | 1 | 1 | 1 | 1 | 1 | 1 | 3 | 4 | 3 | 5 | 4 | 5 | 5 | 5 |
| Mediterranean | 1 | 1 | 1 | 1 | 1 | 2 | 3 | 5 | 5 | -- | 5 | 5 | 5 | 5 |
| North Pacific | 2 | 2 | 2 | 2 | 3 | 3 | 3 | 4 | 3 | -- | 4 | 5 | 5 | 4 |
| Northeast Atlantic | 2 | 2 | 3 | 3 | 3 | 3 | 3 | 4 | 3 | -- | 4 | 3 | 5 | 5 |
| Northwest Atlantic | 2 | 2 | 3 | 3 | 1 | 1 | 3 | 4 | 3 | -- | 4 | 5 | 5 | 5 |
| Oceania (West Central Pacific) | 1 | 1 | 4 | 4 | 4 | 5 | 3 | 2.5 | 2.5 | -- | 3 | 5 | 5 | 5 |
| W. Africa/Southeast Atlantic | 1 | 1 | 1 | 1 | 2 | 2 | 3 | 4 | 3 | -- | 4 | 5 | 5 | 5 |
| Southwest Atlantic | 1 | 1 | 1 | 1 | 1 | 1 | 3 | 4 | 3 | -- | 4 | 5 | 5 | 5 |
| Southwest Pacific (Australia/New Zealand) | 3 | 3 | 1 | 1 | 1 | 2 | 3 | 4 | 3 | -- | 4 | 5 | 5 | 5 |
| West Indian Ocean and Red Sea | 1 | 1 | 1 | 1 | 2 | 1 | 3 | 4 | 3 | -- | 4 | 5 | 5 | 5 |

*For known, unassessed spp., $\geq 3.5=$

## low concern

## Marine Mammal Bycatch Susceptibility

| Region | Longline |  | Gillnet |  | Trawl |  | Dredge <br> LN | Purse Seine |  |  |  | Other |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | LLS | LLD | GNS | GND | TB | TM |  | SUX | PSF | PSD | PSU | FPO | HAR/ MDV | LTL/ LHP/LHM |
| Caribbean/Gulf of Mexico | 3 | 3 | 1 | 1 | 3 | 3 | 5 | 3.5 | 4 | -- | 4 | 3 | 5 | 5 |
| Southeast Asia (East Indian) | 3 | 2 | 1 | 1 | 1 | 2 | 5 | 3.5 | 3 | -- | 4 | 2 | 4 | 5 |
| Eastern Tropical Pacific/Eastern Pacific | 2 | 1 | 1 | 1 | 2 | 1 | 5 | 3 | 3 | 2 | 4 | 2 | 3 | 5 |
| Mediterranean | 2 | 2 | 1 | 1 | 1 | 1 | 5 | 3.5 | 4 | -- | 3 | 2.5 | 4 | 5 |
| Northeast Pacific | 2 | 2 | 1 | 1 | 2 | 2 | 5 | 3 | 4 | -- | 4 | 1 | 5 | 5 |
| Northeast Atlantic | 3 | 3 | 1 | 1 | 2 | 2 | 5 | 3.5 | 4 | -- | 4 | 1 | 5 | 5 |
| Northwest Atlantic | 3 | 2 | 1 | 1 | 3 | 3 | 5 | 3.5 | 4 | -- | 4 | 1 | 5 | 5 |
| Northwest Pacific | 1 | 2 | 1 | 1 | 1 | 2 | 5 | 3.5 | 4 | -- | 4 | 1 | 5 | 5 |
| Oceania (Western Central Pacific) | 4 | 3 | 1 | 1 | 4 | 4 | 5 | 4 | 3 | -- | 4 | -- | 5 | 5 |
| Southern Ocean | 4 | 4 | -- | -- | 4 | 4 | 5 | 5 | 5 | -- | 5 | -- | 5 | 5 |
| West Africa/Southeast Atlantic | 3 | 3 | 1 | 1 | 3 | 1 | 5 | 1 | 2 | -- | 2 | 1 | 5 | 5 |
| Southwest Atlantic | 3 | 3 | 1 | 1 | 2 | 2 | 5 | 3.5 | 4 | -- | 4 | 4 | 5 | -- |
| Southwest Pacific (Australia/New Zealand) | 3 | 1 | 1 | 1 | 1 | 3 | 5 | 3 | 4 | -- | 3 | 2 | 5 | 5 |
| West Indian Ocean and Red Sea | 3 | 3 | 1 | 1 | 3 | 2 | 5 | 3.5 | 4 | -- | 4 | 3 | 5 | 5 |

*For known, unassessed spp., $\geq 3.5$ = low concern

Seabird Bycatch Susceptibility

| Region | Longline |  | Gillnet |  | Trawl |  | Dredge <br> LN | Purse Seine |  |  |  | Other |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | LLS | PLL | GNS | DGN | TB | TM |  | SUX | PSF | PSD | PSU | FPO | HAR/ MDV | LTL/ LHP/LHM |
| Caribbean/Gulf of Mexico | 3 | 3 | 4 | 4 | 4 | 4 | 5 | 5 | 5 | -- | 5 | 4 | 5 | 5 |
| East Indian Ocean/Southeast Asia | 2 | 2 | 3 | 3 | 4 | 4 | 5 | 5 | 5 | -- | 5 | 5 | 5 | 5 |
| Eastern Tropical Pacific/Southeast Pacific | 2 | 2 | 1 | 1 | 2 | 2 | 5 | 4 | 4.5 | 5 | 5 | 4 | 5 | 4.5 |
| Mediterranean | 1 | 1 | 3 | 3 | 2 | 3.5 | 5 | 4 | 5 | -- | 4 | 4 | 5 | 4.5 |
| Northeast Atlantic | 1 | 2.5 | 1 | 1 | 3 | 3 | 5 | 3 | 5 | -- | 5 | 4 | 5 | 4.5 |
| Northeast Pacific | 1 | 1 | 3 | 3 | 3 | 3 | 5 | 4 | 5 | -- | 5 | 4 | 5 | 4.5 |
| Northwest Atlantic | 3 | 3 | 3 | 3 | 4 | 3.5 | 4 | 4 | 5 | -- | 5 | 4 | 5 | 4.5 |
| Northwest Pacific | 1 | 1 | 1 | 1 | 2 | 2 | 5 | 4 | 5 | -- | 5 | 4 | 5 | 4.5 |
| Oceania (Western Central Pacific) | 4 | 2.5 | 3 | 3 | 5 | 5 | 5 | 5 | 4.5 | -- | 4.5 | 5 | 5 | 5 |
| Southern Ocean | 1 | 1 | -- | -- | 2 | 2 | 5 | 5 | 5 | -- | 5 | 5 | 5 | 5 |
| West Africa /Southeast Atlantic | 1 | 2.5 | 2 | 2 | 1 | 1 | 5 | 4 | 5 | -- | 5 | 4 | 5 | 4.5 |
| Southwest Atlantic | 1 | 1 | 2.5 | 2.5 | 2 | 2 | 5 | 4 | 5 | -- | 5 | 4 | 2 | 2 |
| Southwest Pacific (Australia/New Zealand) | 1 | 1 | 2 | 2 | 1 | 1 | 5 | 5 | 5 | -- | 5 | 4 | 5 | 4.5 |
| West Indian Ocean and Red Sea | 2 | 2 | 1 | 1 | 3 | 3 | 5 | 4 | 5 | -- | 5 | 4 | 5 | 4.5 |

*For known, unassessed spp., $\geq 3.5$ = low concern

Unknown Bycatch Matrix - sharks
Shark Bycatch Susceptibility

| Region | Longline |  | Gillnet |  | Trawl |  | Dredge <br> LN | Purse Seine |  |  |  | Other |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | LLS | PLL | GNS | DGN | TB | TM |  | SUX | PSF | PSD | PSU | FPO | HAR/ MDV | LTL/ LHP/LHM |
| Caribbean/Gulf of Mexico | 2 | 1 | 2 | 2 | 2 | -- | 5 | 3.5 | 2 | -- | 3 | 5 | 5 | 4 |
| East Indian Ocean/Southeast Asia | 3 | 3 | 2 | 2 | 1 | 1 | 5 | 3.5 | 1 | -- | 2 | 5 | 5 | 3.5 |
| Eastern Tropical Pacific/Eastern Pacific | 1 | 1 | 2 | 2 | 1 | 2 | 5 | 3.5 | 1 | 3 | 2 | 5 | 5 | 3.5 |
| Mediterranean/Black Sea | 3 | 2 | 3 | 2 | 1 | 1 | 5 | 3.5 | 1 | -- | 3 | 4 | 5 | 3.5 |
| Northeast Atlantic | 1 | 1 | 3 | 3 | 2 | 2 | 5 | 3.5 | 2 | -- | 3 | 3 | 5 | 3.5 |
| Northeast Pacific | 1 | 1 | 2 | 2 | 1 | 2 | 5 | 3.5 | 2 | -- | 3 | 5 | 5 | 3.5 |
| Northwest Atlantic | 1 | 1 | 3 | 3 | 1 | 1 | 5 | 3.5 | 2 | -- | 3 | 5 | 5 | 3.5 |
| Northwest Pacific | 1 | 1 | 2 | 2 | 1 | 1 | 5 | 3.5 | 2 | -- | 3 | 5 | 5 | 3.5 |
| Oceania (Western Central Pacific) | 1 | 1 | 3 | 3 | 2 | 2 | 5 | 3.5 | 1 | -- | 2 | 5 | 5 | 3.5 |
| Southern Ocean | 4 | 4 | -- | -- | 2 | 4 | 5 | 3.5 | 5 | -- | 5 | 5 | 5 | 3.5 |
| West Africa/Southeast Atlantic | 1 | 1 | 2 | 2 | 1 | 1 | 5 | 3.5 | 1 | -- | 2 | 5 | 5 | 3 |
| Southwest Atlantic | 1 | 1 | 2 | 2 | 1 | 2 | 5 | 3.5 | 1 | -- | 2 | 5 | 5 | 3.5 |
| Southwest Pacific (Australia/New Zealand) | 1 | 1 | 2.5 | 2.5 | 2 | 3 | 5 | 1 | 1 | -- | 2 | 5 | 5 | 3.5 |
| West Indian Ocean and Red Sea | 3 | 2 | 1 | 1 | 2 | 2 | 5 | 3.5 | 1 | -- | 2 | 5 | 5 | 3.5 |

*For known, unassessed spp., $\geq 3.5$ = low concern

## Benthic invertebrates, finfish, forage fish, and corals

The values in the matrix of invertebrates, finfish, forage fish, and corals were developed initially by averaging the findings of two studies that ranked the relative ecological impacts of fishing gear (Fuller et al. 2008; Chuenpagdee et al. 2003). Some values in the matrix have been updated based on a survey of scientific experts on bycatch from around the world to increase the global relevance of the matrix.

The findings of the studies used to construct this matrix were pulled from literature searches, fisheries data and expert opinion. In general, these studies ranked the severity of fishing gear impacts as shown in this table (in order of severity):

| Chuenpagdee et al 2003 | Fuller et al 2008 |
| :--- | :--- |
| Bottom trawl | Bottom trawl |
| Bottom gillnet | Bottom gillnet |
| Dredge | Dredge |
| Midwater gillnet | Bottom longline |
| Pot and traps | Midwater trawl |
| Pelagic longline | Pot and trap |
| Bottom longline | Pelagic longline |
| Midwater trawl | Midwater gillnet |
| Purse seine | Purse seine |
| Hook and line | Hook and line |
|  | Dive |
|  | Harpoon |

Because these studies were based on fisheries operating in Canadian and United States waters, we also conducted a review of literature and expert opinion on bycatch severity by gear type from different regions of the world. Some of the initial values from Fuller et al. (2008) and Chuenpagdee et al. (2003) were adjusted accordingly. These changes are intended to better reflect the array of bycatch issues that occur using the same gear types in different regions of the world.

Bycatch severity for biogenic habitats (coral and sponges) by gear type was determined by averaging the values given in Fuller et al. (2008) and Chuenpagdee et al. (2003). Chuenpagdee et al. (2003), named this category "biological habitat" and Fuller et al. (2008) called it "coral and sponges." We did not change these values because it is likely that gear types that contact the bottom have the same potential for severe impacts throughout the world's oceans. Impacts from fishing on the benthos occur on virtually all continental shelves worldwide (Watling 2005).

We increased the number of trawl types from only bottom and midwater (used in both Fuller et al. (2008) and Chuenpagdee et al. (2003)) to also include bottom trawl categories for tropical/subtropical fish, tropical/subtropical shrimp, coldwater fish, and coldwater shrimp. Shrimp trawls are not designed to drag along the bottom and herd fish, so they receive a lower impact score in the matrix for finfish bycatch.

Other changes to the findings of Fuller et al. (2008) and Chuenpagdee et al. (2003) include separating the different purse seine techniques into FAD/log sets, dolphin/whale sets and unassociated school sets
based on the variable bycatch rates found in a study by Hall (1998). Hall (1998) found that log (FAD) sets have the overall greatest bycatch for some species, followed by school sets and dolphin sets.

Bottom seines or demersal seines (including Danish seines, Scottish fly-dragging seines and pair seines) were not included in the Fuller et al. (2008) and Chuenpagdee et al. (2003) studies because these gear types are not commonly used in the U.S. or Canada. Like purse seines, these gear types encircle a school of fish, but they are operated in contact with the seafloor. A study by Palsson (2003) compared haddock discards among three demersal gear types in Icelandic waters and found fish bycatch to be lowest in Danish seines when compared with demersal trawls and longline gear. Danish seines targeting benthic fish species can incidentally catch non-target species such as flatfish, cod, and haddock (Icelandic Ministry of Fisheries 2010). Alverson et al. (1996) found that Danish seines generally fell into a lowmoderate bycatch group of gear, with lower bycatch ratios than the majority of gear types, including bottom trawls, longlines and pots, but with higher bycatch than pelagic trawls and purse seines. Based on these findings, the bycatch score of Danish seines was estimated from the score for purse seines with an increase in the effects on shellfish to account for Danish seines being operated on the seafloor, an increase in the effect on finfish to account for greater bycatch of benthic fish such as flatfish, cod and haddock, and a decrease in the effect on forage fish, which are typically pelagic.

Unknown Bycatch Matrix - benthic invertebrates, finfish, forage fish, and corals and other biogenic habitats

Highest impacts receive a score of 1 and lowest impacts receive a score of 5. Key: $\mathrm{B}=$ Bottom, $\mathrm{P}=\mathrm{Pelagic} \mathrm{M}=$,Mid -water, $\mathrm{BTF}=\mathrm{Bottom}$ tropical fish, BTS = Bottom tropical shrimp, BCF = Bottom coldwater fish, BCS = Bottom coldwater shrimp, PF = Purse FAD/log (tuna), PD = Purse dolphin/whale (tuna), PU = Purse unassociated (tuna), Pot = Pot and trap, HD = Harpoon/diver, TP = Troll/pole and line

|  | Longline |  | Gillnet |  | Trawl |  |  |  |  |  | Dredge | Seine |  |  |  |  | Other |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | B | P | B | M | B | BTF | BTS | BCF | BCS | M |  | B | P | PF | PD | PU | Pot | HD | TP |
| Benthic Inverts | 4.5 | 5 | 3 | 5 | 2 | 2 | 2 | 2 | 2 | 5 | 1 | 3 | 5 | 5 | 5 | 5 | 3.5 | 5 | 5 |
| Finfish | 2 | 2 | 2 | 1 | 2 | 1 | 2 | 2.5 | 2 | 1 | 3 | 2 | 4 | 2 | 3 | 2 | 3.5 | 5 | 3 |
| Forage Fish | 5 | 4 | 2 | 2 | 3 | 2 | 2 | 2 | 2 | 1 | 5 | 3 | 3 | 2 | 3 | 2 | 4 | 5 | 4 |
| Corals and other biogenic habitats | 3 | 5 | 2 | 5 | 1 | 1 | 1 | 1 | 1 | 5 | 1 | 1 | 5 | 5 | 5 | 5 | 3.5 | 5 | 4.5 |

## Additional guidance for scoring unknown bycatch species in Criterion 1.1/2.1 (Abundance)

Sea turtles - all endangered/threatened: See Wallace et al. $(2010,2013)$ for global patterns of marine turtle bycatch. In addition, a global program, Mapping the World's Sea Turtles, created by the SWOT (State of the World's Sea Turtles) database is a comprehensive global database of sea turtle nesting sites around the world. The SWOT map is highly detailed and can be customized, allowing location filters and highlights of both species and colony size with variously colored and shaped icons. This map together with the paper by Wallace et al. (2010) can help to determine if the fishery being assessed has potential interactions with sea turtles.

Sharks, marine mammals and seabirds: Identify whether the fishery overlaps with any endangered/threatened or overfished species and err on the side of caution if species-specific and geographic information is inconclusive. For example, if shark populations are data deficient, err on the side of caution and rate as "overfished" or "depleted."

Sharks: Select "overfished" or "depleted" when data deficient or select "endangered/threatened" when data exist to support this (see Camhi et al. 2009). Globally, three-quarters (16 of 21) of oceanic pelagic sharks and rays have an elevated risk of extinction due to overfishing (Dulvy et al. 2008). See Camhi et al. (2009) for geographic areas, IUCN status and conservation concerns by shark species. Table 1 illustrates additional resolutions, recommendations and conservation and management measures by RFMO for sharks. Additional region and species-specific shark conservation information associated follows Table 1 in list format (Camhi 2009; Bradford 2010).

Marine mammals: The global distribution marine mammals and their important conservation areas are given by Pompa et al. (2011), who also used geographic ranges to identify 20 key global conservation sites for all marine mammal species (123) and created range maps for them (Figure 1; Table 2; Pompa et al. 2011 and supplement).

Seabirds: Figures 2 and 3 illustrate the distribution of threatened seabirds throughout the world (Birdlife International 2011). Also see Birdlife International (2010) to locate Marine Important Bird Areas (MIBA). Albatross are the most highly threatened family, with all 22 species either globally threatened or near threatened. The penguins and shearwaters/gadfly petrels also contain a high proportion of threatened species (Birdlife International 2010).

Table 1. Active resolutions, recommendations, and conservation and management measures by RFMO for sharks. Table from Camhi et al. (2009). ${ }^{\text {a }}$ ICCAT $=$ International Commission for the Conservation of Atlantic Tunas; NAFO = North Atlantic Fisheries Organization; GFCM = General Fisheries Commission for the Mediterranean; SEAFO = South East Atlantic Fisheries Organization; IATTC = Inter-American Tropical Tuna Commission; WCPFC = Western and Central Pacific Fisheries Commission; IOTC = Indian Ocean Tuna Commission; CCAMLR = Commission for the Conservation of Antarctic Marine Living Resources. ${ }^{\text {b }}$ The weight of recommendations and resolutions varies by RFMO. For example, all ICCAT recommendations are binding, whereas resolutions are not.

| Ocean/ <br> RFMO ${ }^{\text {a/ }}$ Year | Res/Rec No. ${ }^{\text {b }}$ | Title | Main actions |
| :---: | :---: | :---: | :---: |
| Atlantic, ICAAT |  |  |  |
| 1995 | Res. 95-2 | Resolution by ICCAT on cooperation with the FAO to study the status of shark stocks and by-catches | - Urges members to collect species-specific data on biology, bycatch and trade in shark species and provide these data to FAO |
| 2003 | Res. 03-10 | Resolution by ICCAT on the shark fishery | - Requests all members to submit data on shark catch, effort by gear, landings and trade in shark products <br> - Urges members to fully implement a NPOA |
| 2004 | Rec. 04-10 | Recommendation by ICCAT concerning the conservation of sharks caught in association with fisheries managed by ICCAT | - Requires members to annually report shark catch and effort data <br> - Requires full utilization <br> - Bans finning <br> - Encourages live release <br> - Commits to reassess shortfin mako and blue sharks by 2007 <br> - Promotes research on gear selectivity and identification of nursery areas |
| 2005 | Rec. 05-05 | Recommendation by ICCAT to amend Recommendation 04-10 concerning the conservation of sharks caught in association with fisheries managed by ICCAT | - Requires annual reporting of progress made toward implementation of Rec. 04-10 by members <br> - Urges member action to reduce North Atlantic shortfin mako mortality |


| 2006 | Rec. 06-10 | Supplementary recommendation by <br> ICCAT concerning the conservation of <br> sharks caught in association with <br> fisheries managed by ICCAT | • Acknowledges little progress in quantity and quality of shark catch statistics <br> - Reiterates call for current and historical shark data in preparation for blue and <br> shortfin mako assessments in 2008 |
| :---: | :--- | :--- | :--- |
| 2007 | Rec. 07-06 | Supplemental recommendation by <br> ICCAT concerning sharks | • Reiterates mandatory data reporting for sharks <br> - Urges measures to reduce mortality of targeted porbeagle and shortfin mako |
| 2008 | Rec. 08-07 Encourages research into nursery areas and possible time and area closures |  |  |


| Med., GFCM |  |  |  |
| :---: | :---: | :---: | :---: |
| 2005 | GFCM/2005/3 | Recommendation by ICCAT concerning the conservation of sharks caught in association with fisheries managed by ICCAT | - Same provisions as ICCAT Rec. 04-10 |
| 2006 | GFCM/2006/8(B) | Recommendation by ICCAT to amend Recommendation [04-10] concerning the conservation of sharks caught in association with fisheries managed by ICCAT | - Same provisions as ICCAT Rec. 05-05 |
| Indian, IOTC |  |  |  |
| 2005 | Res. 05/05 | Concerning the conservation of sharks caught in association with fisheries managed by IOTC | - Requires members to report shark catches annually, including historical data <br> - Plans to provide preliminary advice on stock status by 2006 <br> - Requires full utilization and live release <br> - Bans finning - Promotes research on gear selectivity and to ID nursery areas |
| 2008 | Res. 08/01 | Mandatory statistical requirements for IOTC members and cooperating non-contracting parties (CPCs) | - Requires members to submit timely catch and effort data for all species, including commonly caught shark species and less common sharks, where possible |
| 2008 | Res. 08/04 | Concerning the recording of catch by longline fishing vessels in the IOTC area | - Mandates logbook reporting of catch by species per set, including for blue, porbeagle, mako and other sharks |
| Pacific, IATTC |  |  |  |
| 2005 | Res. C-05-03 | Resolution on the conservation of sharks caught in association with fisheries in the Eastern Pacific Ocean | - Promotes NPOA development among members <br> - Work with WCPFC to conduct shark population assessments <br> - Promotes full utilization <br> - Bans finning <br> - Encourages live release and gear-selectivity research <br> - Requires species-specific reporting for sharks, including historical data |


| 2006 | $\begin{aligned} & \text { Res. C-04-05 } \\ & \text { (REV 2) } \end{aligned}$ | Consolidated resolution on bycatch | - Requires prompt release of sharks, rays and other non-target species <br> - Promotes research into methods to avoid bycatch (time-area analyses), survival rates of released bycatch and techniques to facilitate live release <br> - Urges members to "provide the required bycatch information as soon as possible" |
| :---: | :---: | :---: | :---: |
| Pacific, WCFPC |  |  |  |
| 2008 | Cons. \& Mgt. <br> Measure 2008- <br> 06 (replaces <br> 2006-05) | Conservation and management measure for sharks in the Western and Central Pacific Ocean | - Urges members to implement the IPOA and report back on progress <br> - Requires annual reporting of catches and effort <br> - Encourages live release and full utilization <br> - Bans finning for vessels of all sizes <br> - Plans to provide preliminary advice on stock status of key sharks by 2010 |
| Southern, CCAMLR |  |  |  |
| 2006 | 32-18 | Conservation of sharks | - Prohibits directed fishing of sharks <br> - Live release of bycatch sharks |

## Additional shark information and citations (Bradford 2010)

- In the Gulf of Mexico, Baum and Myers (2004) found that between the 1950s and the late1990s, oceanic whitetip and silky sharks (formerly the most commonly caught shark species in the Gulf of Mexico) declined by over 99 and $90 \%$, respectively.
- In the Northwest Atlantic, Baum et al. (2003) estimated that scalloped hammerhead, white, and thresher sharks had declined by over $75 \%$ between the mid-1980s and late-1990s. The study also found that all recorded shark species in the Northwest Atlantic, with the exception of mako sharks, declined by over 50\% during the same time period.
- Myers et al. (2007) reported declines of $87 \%$ for sandbar sharks, $93 \%$ for blacktip sharks, $97 \%$ for tiger sharks, $98 \%$ for scalloped hammerheads, and 99\% or more for bull, dusky, and smooth hammerhead sharks along the Eastern seaboard since surveys began along the coast of North Carolina in 1972.
- The International Union for the Conservation of Nature (IUCN) has declared that "32\% of all pelagic sharks and rays are threatened." The IUCN has declared another 6\% to be Endangered, and $26 \%$ to be Vulnerable.
- In the Mediterranean Sea, Ferretti et al. (2008) found that hammerhead, blue, mackerel, and thresher sharks have declined between 96 and 99.99\% relative to their former abundance levels.
- Ward and Myers (2005) report a $21 \%$ decline in abundance of large sharks and tunas in the tropical Pacific since the onset of commercial fishing in the 1950s.
- Meyers and Worm (2005) indicate a global depletion of large predatory fish communities of at least 90\% over the past 50-100 years. The authors suggest that declines are "even higher for sensitive species such as sharks."
- Dulvy et al. (2008) state that "globally, three-quarters (16 of 21) of oceanic pelagic sharks and rays have an elevated risk of extinction due to overfishing."
- Graham et al. (2001) found an average decrease of $20 \%$ in the catch rate of sharks and rays off New South Wales, Australia, between 1976 and 1997.

Table 2. Marine mammal species in important conservation sites. "Irreplaceable areas" contain species found nowhere else. Figures from Pompa et al. (2011; supplt. material). ${ }^{1}$ Monachus schauinslandi, ${ }^{2}$ Arctocephalus galapagoensis, ${ }^{3}$ A. philippii, ${ }^{4}$ Inia geoffrensis, Trichechus inunguis (both freshwater) and Sotalia fluviatilis, ${ }^{5}$ Monachus monachus, ${ }^{6}$ Platanista minor (freshwater), ${ }^{7}$ Platanista gangetica (freshwater), ${ }^{8}$ Lipotes vexillifer (freshwater), ${ }^{9}$ Pusa sibirica (freshwater), ${ }^{10}$ Pusa caspica, ${ }^{11}$ Cephalorhynchus commersonii and $A$. gazella. *VU = Vulnerable, EN = Endangered, CR = Critically Endangered, LR = Lesser Risk, EX = Extinct, CE = Critically Endangered; V = Vulnerable, RS = Relatively Stable or Intact. Data from Olson and Dinerstein (2002).

| Key conservation sites | Number of species | Endemic/ <br> small- <br> range | Risk category for each ecoregion* | Number and name of the ecoregion* | Estimated conservation status of the ecoregion* |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Highest richness |  |  |  |  |  |
| South African | 16 | 4 | VU, EN | 209: Benguela Current <br> 211: Agulhas Current | $\begin{aligned} & \hline \mathrm{V} \\ & \mathrm{RS} \end{aligned}$ |
| Argentinean | 15 | 4 | VU, EN | 205: Patagonian Southwest Atlantic | V |
| Australian | 14 | 4 | VU, EN | 206: Southern Australian <br> 222: Great Barrier | $\begin{aligned} & \hline \text { RS } \\ & \text { RS } \end{aligned}$ |
| Baja Californian | 25 | 7 | VU, EN, CR | 214: Gulf of California | CE |
| Peruvian | 19 | 5 | VU, EN | 210: Humboldt Current | V |
| Japanese | 25 | 7 | VU, EN, LR | 217: Nansei Shoto | CE |
| New Zealand | 13 | 2 | VU, EN, LR | 207: New Zealand | V |
| Northwestern African | 25 | 7 | VU, EN, LR | 216: Canary Current | CE |
| Northeastern American | 25 | 7 | VU, EN, LR | 202: Chesapeake Bay | V |
| Irreplaceable |  |  |  |  |  |
| Hawaiian Islands | $1^{1}$ | 1 | EN | 227: Hawaiian Marine | V |
| Galapagos Islands | $1^{2}$ | 1 | VU | 215: Galapagos Marine | V |
| San Félix and Juan Fernández Islands | $1^{3}$ | 1 | VU | 210: Humboldt Current | V |
| Amazon River | $2^{4}$ | 1 | VU | 147: Amazon River/Flooded Forests | RS |
| Mediterranean Sea | $1^{5}$ | 1 | CR | 199: Mediterranean Sea | CE |
| Indus River | $1^{6}$ | 1 | Not Listed | Not Listed | Not Listed |
| Ganges River | $1^{7}$ | 1 | EN | Not Listed | Not Listed |
| Yang-tse River | $1^{8}$ | 1 | EX | 149: Yang-Tse River And Lakes | CE |
| Baikal Lake | $1^{9}$ | 1 | LR | 184: Lake Baikal | V |
| Caspian Sea | $1^{10}$ | 1 | VU | Not Listed | Not Listed |
| Kerguelen Islands | $1^{11}$ | 1 | Not Listed | Not Listed | Not Listed |



Figure 1. Geographic distribution of marine mammal species richness (left column) for A. Pinnipeds; B. Mysticetes; C. Odontocetes. Figure from Pompa et al. (2011).


Figure 2. At-sea distribution of threatened seabirds around the globe. Each polygon represents the range map for one threatened species. Areas of darkest blue show the areas of the ocean where the ranges of the greatest number of threatened species overlap. Figure from Birdlife International (2011).


Figure 3. Worldwide distribution of albatross and petrels. Figure from Birdlife International (2011).

## Appendix 3 - Appropriate management strategies

Appropriate management procedures may vary greatly between different fisheries, regulatory frameworks and species. To some extent, assessment of harvest control rules and other management strategies must therefore be addressed on a case-by-case basis. However, general guidelines for appropriate management are still relevant and useful. For fisheries managed using catch limits or TACs, these guidelines have been derived largely from the guidance provided for implementation of the Magnuson-Stevens Fishery Conservation and Management Act used for fishery management in the U.S. (Restrepo and Powers 1998; Restrepo et al. 1998). While other countries have different regulatory frameworks, similar strategies to those suggested in Restrepo et al. (1998) are used throughout the world where stock assessments are available and catch limits are employed (e.g., Australian Department of Agriculture, Fisheries and Forestry 2007; NZ Ministry of Fisheries 2008; DFO 2009). Commonly accepted strategies include setting fishing mortality rates safely below $\mathrm{F}_{\text {MSY }}$ (or other appropriate reference point) to account for uncertainty; reducing F when stocks fall below biomass target reference points (generally around $\mathrm{B}_{\text {MSY }}$ or $40 \%$ of unfished biomass); and reducing fishing mortality when stock falls below a critical level where recruitment is impaired. Management reference points are assumed to be valid unless scientific information exists to suggest otherwise, e.g., a scientific assessment or controversy that strongly suggests current reference points are not appropriate for the species under assessment.

In general, the minimal attributes of an appropriate management strategy include:

1. A process for monitoring and conducting "assessments" (not necessarily formal stock assessments). Monitoring should occur regularly, though the frequency of assessments needed may vary depending on the variability of the stock.
2. Rules that control the intensity of fishing activity or otherwise ensure the protection of stock productivity.
3. A process to modify rules according to assessment results, as needed.

## Some effective management strategies

For data-rich or data-moderate stocks that have quota-based management, a "highly effective" management strategy is one that:

- Incorporates an up-to-date, scientific stock assessment that allows managers to determine if stocks are healthy and to set appropriate quotas;
- Uses appropriate limit and target reference points for stock and fishing mortality;
- Chooses risk-averse policies rather than risky, yield-maximizing policies;
- Includes buffers in the TAC to account for uncertainty in stock assessments
- Set Allowable Biological Catch ( ABC ) and Annual Catch Limit ( ACL ) at less than the OverFishing Level (OFL = long term mean of MSY) to account for scientific uncertainty (survey data on stock size, etc. can reduce scientific uncertainty);
- Set Total Allowable Catch (TAC) at less than ABC to account for management uncertainty (monitoring catch, etc. can reduce management uncertainty);
- As a rule of thumb, TAC should have less than $30 \%$ p* (likelihood) of exceeding OFL; or TAC should be set such that F is $25 \%$ below the threshold fishing pressure, e.g. $\mathrm{F}_{\text {MSY }}$ (Restrepo et al. 1998)
- Stocks with low biomass, high vulnerability, and high uncertainty warrant greater protection against overfishing (e.g., more conservative harvest control rules/ greater buffers in setting TAC and/or closer monitoring of stocks).
- Takes into account other sources of mortality (e.g., recreational fishery, bycatch of juveniles, etc.) and environmental factors that affect stock, such as oceanographic regime;
- Incorporates a strategy for maintaining or rebuilding stock productivity:
- A no-fishing point when biomass is below the limit reference point;
- A decrease in F when biomass is below the target reference point or is declining (whether declines are due to fishery or environmental factors).
- Employs an effective strategy to prevent overcapitalization;
- Has been demonstrated effective (e.g., stock productivity has been maintained over multiple generations), or if stock productivity has not been maintained or is declining, have adjusted management accordingly.


## Effective management in data-limited fisheries

(more information on data-limited evaluation methods below)

Whether managed stocks are data-rich or data-limited, management must include a strategy to ensure that stock productivity is maintained in order to be considered effective. This strategy should include a process for monitoring and conducting "assessments" of some kind (not necessarily formal stock assessments), rules that control the intensity of fishing activity or otherwise ensure the protection of a portion of the spawning stock, and a system of adaptive management, such that rules are modified according to assessment results, as needed (Smith et al. 2009; Phipps et al. 2010).

There are some relatively reliable methods for setting catch limits in data-limited fisheries, including: An Index Method (AIM), which involves fitting a relationship between population abundance indices and catch; Depletion-Corrected Average Catch (DCAC), which allows managers to estimate a sustainable yield based on average catch over a set time period, adjusting for initial declines in abundance due to harvesting; and extrapolation methods, or relying on inferences from related or "sister" stocks, with the use of precautionary buffers in case the data-limited stocks are more vulnerable than the related datarich stocks (Honey et al. 2010). Other techniques recommended for data-limited stocks include the use of productivity-susceptibility analysis (PSA) to highlight stocks that are particularly vulnerable to overexploitation (Patrick et al. 2009; Honey et al. 2010) and setting catch limits based on historical catch from a period of no declines, with targets set at $75 \%$ of average catch if biomass is believed to be healthy, $50 \%$ of average catch if biomass is expected to be below target levels but above the point where recruitment would be impaired, and $25 \%$ of average catch if the stock is depleted (Restrepo and Powers 1998).

Other than constraining fishing mortality (e.g., through TACs), fisheries may be credited for employing alternative strategies that are widely believed to be help maintain stock productivity. Some examples of effective alternative strategies are spatial management, including protecting a large proportion of coastline in reserves and/or protecting known spawning aggregations with seasonal or spatial closures (e.g., Johannes 1998), or protecting females, which preserves the spawning per recruit of the population as long as fertilization does not decrease (e.g., Dungeness crab; Chaffee et al. 2010). Finally, stocks may be subject to low mortality in a data-limited fishery as a result of low susceptibility, e.g., if the species is small enough to fit between the mesh of the nets or is not attracted to the type of bait used (low susceptibility is generally more applicable as protection for non-target stocks).

For data-limited stocks, management should:

- Include a process for monitoring and assessment, such as recording trends in CPUE and size structure, or estimating FLEP, or comparison of abundance index to historical high (see glossary), unfished, or marine reserve levels:
- Trends in CPUE are appropriate only if technology has not changed, there is no hyperstability, and abundance is shown to be proportional to CPUE;
- Trends in size structure must also be monitored to avoid depletion of large individuals.
- Include a strategy for protecting spawning stock, such as:
- Estimate sustainable yield based on Depletion Corrected Average Catch (DCAC), An Index Method (AIM), or another accepted strategy;
- Protect a large portion of spawning stock in marine reserves (at least 50\%, including important spawning areas if applicable) or close hotspots to fishing (for bycatch species);
- Enforce size, sex, and/or season limitations that are likely to be effective in protecting spawning stock productivity (e.g., Dungeness crab 3-S management);
- Extrapolate based on data-rich related or "sister" stocks, with precautionary buffers in place to account for potential differences in the stocks' life histories;
- Maintain exploitation rates at very low levels (e.g., experimental fishery) until more data can be collected, or
- Base TAC on average historical catch during a period of time with no declines in abundance (TAC should be set at no more than $75 \%$ of average catch if stock is believed to be healthy, $50 \%$ if believed to be below target levels, and $25 \%$ if believed to be overfished. Note: as there is generally no data to assess whether a stock is healthy, TAC should not be more than $50 \%$ of historical catch unless there is a strong scientific reason to believe that stocks are above $\mathrm{B}_{\text {Msy }}$ ).
- Allow for adaptive management so that fishing strategy is adjusted if assessment/monitoring indicates that stock is declining or below target levels;
- Have been demonstrated effective (e.g., stock productivity has been maintained over multiple generations) or, if stock productivity has not been maintained or is declining, management has been adjusted accordingly.

Procedures for monitoring/assessing stocks and procedures for protecting spawning stock must be in place, and be demonstrated effective, to qualify management strategy as "highly effective." If measures are expected to be effective, e.g., through analogy with similar systems, but have not been demonstrated effective in this fishery, management is "moderately effective." If measures are not expected to be effective, management strategy is "ineffective."

Appropriate management also depends on the conservation concern associated with the stock. In addition to the precautionary elements listed above, stocks that are endangered or threatened also require a recovery plan and/or best management practices designed and demonstrated to reduce mortality and allow the stock to recover. Overfished and depleted stocks require a rebuilding plan.

## Data-limited fishery evaluation methods

## Sequential trend analysis (index indicators)

Sequential analysis comprises a broad suite of techniques used to analyze time series data in order to detect trends in a variable (or in various indices) and infer changes in the stock or population. Sequential analyses can encompass a wide range of data types and requirements (Honey et al. 2010). Examples include: DCAC, time series of catch statistics, survey/weight/length-based reference points, trophic indices, and spawning potential ratio (SPR) analogues (Honey et al. 2010).

Depletion-corrected average catch (DCAC) uses only catch time-series data supplemented with educated guesses for a few supplementary parameters. Therefore, it is likely of practical use for many data-limited fisheries on long-lived species (e.g., natural mortality, $\mathrm{M}<0.2$ ) (Honey et al. 2010). The ability of this method to identify sustainable yields from simple data input makes DCAC useful as a first-step estimate for an allowable catch level along with other data-limited methods. See: https://nmfs-fishtools.github.io/for the NOAA toolbox to perform DCAC analysis (Honey et al. 2010).

## Vulnerability analysis

Productivity and susceptibility analysis of vulnerability - The Productivity-Susceeptibibility Analysis of vulnerability (PSA) method is used to assess a stock's vulnerability to overfishing, based on relative scores derived from life-history characteristics. Productivity, which represents the potential for stock growth, is rated semi-quantitatively from low to high on the basis of the stock's intrinsic rate of increase $(r)$, von Bertalanffy growth coefficient ( $k$ ), natural mortality rate ( $M$ ), mean age at maturity, and other metrics (Patrick et al. 2009; Patrick et al. 2010; Field et al. 2010; Cope et al. 2011; Honey et al. 2010).

To assist regional fishery management councils in determining vulnerability, NMFS elected to use a modified version of a productivity and susceptibility analysis (PSA) because it can be based on qualitative data, has a history of use in other fisheries, and is recommended by several organizations as a reasonable approach for evaluating risk (Patrick et al. 2010). Patrick et al. (2010) evaluated six U.S. fisheries targeting 162 stocks that exhibited varying degrees of productivity and susceptibility, and for which data quality varied. Patrick et al. (2010) found that PSA was capable of differentiating the vulnerability of stocks along the gradient of susceptibility and productivity indices. The PSA can be used as a flexible tool capable of incorporating region-specific information on fishery and management activity. Similar work was conducted by Cope et al. (2011) who found that PSA is a simple and flexible approach to incorporating vulnerability measures into complex stock designations while also providing information helpful in prioritizing stock- and complex-specific management.

## Extrapolation (Robin Hood Method)

When very limited or no data are available for a stock or specific species in a region, then managers may need to rely on extrapolation methods to inform decisions. Often, low-value stocks are data-limited (Honey et al. 2010). This method is termed the "Robin Hood" approach in Australia because it takes information and scientific understanding in data-rich fisheries and "gives" inferences to the data-limited fisheries (Smith et al. 2009). Data may include: (1) the local knowledge of the fishers and resource users; and/or (2) scientific research and ecosystem understanding from "sister" systems thought to be similar (Honey et al. 2010). Extrapolation from similar systems or related species may offer an informed starting point from which managers can build precautionary management (Honey et al. 2010). In these
situations, life-history characteristics, potentially sustainable harvest levels, spawning behavior, and other information can be gleaned from nearby stocks, systems, or related species (Honey et al. 2010).

## Decision-making methods

## Decision trees

Decision trees provide systematic, hierarchical frameworks for decision-making that can scale to any spatial, temporal, or management context in order to address a specific question. A decision tree may be customized to meet any need (Honey et al. 2010). Trees may include: identification of reference points based on stock characteristics and vulnerability (Cope and Punt 2009); fostering of fine-scale, transparent, and local management (Prince 2010); and, estimation and refinement of an appropriate Total Allowable Catch (TAC) level (Wilson et al. 2010).

## Management strategy evaluation

Management Strategy Evaluation (MSE) is a general modeling framework designed for the evaluation of performance of alternative management strategies for pursuing different objectives (Honey et al. 2010). This approach simulates the fishery's response to different management strategies (e.g., different TAC levels, seasonal closures, or other effort reductions) (Honey et al. 2010). Assuming sufficient quality data exist, MSE may be useful for assessing the effectiveness of different policy options (Honey et al. 2010).

In addition, a study by Dowling et al. (2008) developed harvest strategies for data-limited fisheries in Australia. Strategies included: (i) the development of sets of triggers with conservative response levels, with progressively higher data and analysis requirements at higher response levels, (ii) identification of data gathering protocols and subsequent simple analyses to better assess the fishery, (iii) the archiving of biological data for possible future analysis, and (iv) the use of spatial management, either as the main aspect of the harvest strategy or together with other measures (Honey et al. 2010).

## Cooperative research and co-management to overcome data-limited situations

A recent study by Fujita et al. (2010) identified opportunities for cooperative research and comanagement that would complement (but not replace) existing top-down fishery regulations. They conclude that management and data collection would improve for some small-scale fisheries if they started: collecting data at the appropriate spatial scales; collecting local information, improving the quality of data, and overcoming constraints on data; providing ecosystem insight from a small/local scale for new and different perspectives; reducing conflicts among fishermen, scientists, and regulators; and improving the responsiveness of fisheries management to local needs. Fujita et al. (2010) suggest that scientists and managers should further develop cooperative strategies (e.g., cooperative research and co-management) and include them in the management framework.

## Effective management of a fishery on a non-native species

Effective management of a fishery for a non-native species may include:

- Mitigation strategies aimed at eradication, reversing establishment, or maintenance at low abundance, as deemed appropriate and feasible for that particular case;
- Adaptation strategies that allow for recovery of species impacted by the non-native species;
- Containment measures such as fishing at the boundaries of the stock to prevent further spread, and/or
- Provisions to prohibit further introductions of any other alien species.


## Management strategies to minimize discarding

Discarding of catch can occur for a variety of reasons, including but not limited to low commercial value and falling outside of regulatory requirements (for example, below a minimum landing size or no quota availability). Discarding is a wasteful practice that is undesirable to both fishers and managers alike. There are a number of strategies that can be employed around the world to minimize discarding: The use of bycatch reduction devices (BRDs) can reduce the catch of undersized individuals of the target species and smaller species of fish and have been used to some success in the Gulf of Mexico shrimp trawl fleet. In an attempt to better quantify the impact of fisheries on all fish stocks within the catch, some fisheries are moving to a zero-discard strategy which requires all fish caught to be landed. Typically, these approaches require high levels of observer coverage or electronic monitoring solutions to ensure compliance.

Many discard avoidance or mitigation strategies will relate to bycatch species and should be considered under factor 3.2; however, it is also important to consider the impact of discarding on the retained species (for example the discarding of undersized individuals as a result of regulation or high-grading) and any measures that have been introduced to mitigate/address these concerns. Such measures should be discussed in factor 3.1.

## Flexible and Resilient fisheries management in the face of climate change

This section is a work in progress and will be expanded in the future. Seafood Watch will provide guidance to its analysts on principles and practices of flexible and resilient fisheries management strategies to be applied in Criterion 3. This will be particularly useful when the species' (and their habitats and the ecosystems) under assessment are or may be impacted by climate change, which includes forage species.

The concept of resilience in fisheries management focuses on how to build capacity that can buffer the impacts of unexpected (or predicted/expected) changes before they occur. While it is often ideal for management to anticipate changes, it is often the case that management responds to change once it has occurred. Seafood Watch accepts reactive management as potentially highly effective, as long as it is implemented in an appropriate timeframe (for example, before a stock falls below a critical biomass such that recovery does not or is likely not to occur)

Below are examples of reactive strategies that are a response to change after it has occurred and proactive strategies that plan for changes that may occur and which promote resilience of stocks and ecosystems. These are adapted from Morrison and Termini 2016. This list provides examples, and is not intended as a comprehensive list of strategies to manage for resilience. Due to the diversity of fishery and ecosystem characteristics that may be encountered it is important to consider each one on a case by case basis. Further guidance will be provided as scientific understanding in this area develops.

| Management profile | Management strategies |
| :---: | :---: |
| Reactive | Flexible management systems are in place (systems that identify when management changes are needed and are able to implement these changes in a timely manner) |
|  | Reference Points are adjustable after changes in species productivity or stock structure occurred |
|  | Fisheries Allocations can be adjusted (if species abundances or distributions changed) |
|  | Fishing practices or gears are adjusted (as Fish Community Composition Changes) |
| Proactive management that increase species' resilience | Managing for uncertainty- Scenario Planning |
|  | Managing to promote adaptive capacity |
|  | Protecting age structure and/or old females |
|  | Incorporating environmental parameters into stock assessments and management measures |
|  | Decreasing existing stressors |
|  | Enhancing or translocating stocks |
| Proactive management that increase ecosystem resilience | Protecting key habitats and species |
|  | Applying ecosystem models to better understand species' responses |
|  | Designing appropriate marine reserves |

## Appendix 4 - Bycatch reduction approaches

In general, fisheries should address bycatch with the following approaches:

- Monitor bycatch rates (using adequate observer coverage),
- Have some scientific assessment of impacts on bycatch populations
- Incorporate strategies that assure bycatch is minimized, such as:
- Enforcing effective and appropriate bycatch caps,
- Closing hotspots or implementing seasonal closures,
- Promoting effective gear modifications such as BRDs, TEDs, etc.
- Adopting bycatch-reducing strategies such as night setting,
- Using the best available management techniques that have been demonstrated in this or a similar system to effectively constrain bycatch rates.

The effectiveness of various bycatch reduction approaches is synthesized from primary literature and reviewed below. To be considered "highly effective", all required measures and at least one primary measures should be in place.

Seabird sources are Løkkeborg (2008) (general conclusions and Table 3, including percent effectiveness of some modification/region strata) and SBWG 2010 (Annexes 3-8). *Secondary measures may be useful in conjunction with primary measures. Turtle sources are FAO 2009 (Tables 1 and page 79) and Gilman and Lundin (2008) (Table 3). Shrimp trawl modifications sources are Eayers 2007 and Gillet 2008 (Box 14). Sharks and marine mammals from Gilman and Lundin (2008) (Table 3). General information on fishing technologies can be found at http://www.fao.org/fishery/en, and a list of bycatch reduction literature can be found here: http://www.bycatch.org/articles.

| Gear/taxon/modification | Primary/ secondary measure* | Effectiveness/notes |
| :---: | :---: | :---: |
| General strategies (good for all gears/taxa) |  |  |
| Monitoring and compliance | Requirement | Considerable difference between experimental and real-world effectiveness. "Three common themes to successful implementation of bycatch reduction measures are long-standing collaborations among the fishing industry, scientists, and resource managers; preand post-implementation monitoring; and compliance via enforcement and incentives" (Cox, Lewison et al. 2007). |
| Avoid bycatch hotspots | Primary | Area/time closures. Generally, very effective, though more so when based on data such as tagging or bycatch data. Perhaps only a secondary mitigation measure for birds (Løkkeborg 2008). Alternatively, move when interaction rates are high. Effective for all fisheries, especially with fleet communication. Closures for one taxon without commensurate reduction in effort can increase bycatch of other taxa. |
| Bycatch caps | Primary | l.e., fishery closes when cap exceeded. |
| Bycatch fees, Compensatory Mitigation Strategies for Marine Bycatch (CMSMB) | Secondary, at best | Not effective. "We conclude that, overall, CMMB has little potential for benefit and a substantial potential for harm if implemented to solve most fisheries bycatch problems. In particular, CMMB is likely to be effective only when applied to short-lived and highly fecund species (not the characteristics of most bycatch-impacted species) and to fisheries that take few non-target species, and especially few non-seabird species (not the characteristics of most fisheries). Thus, CMMB appears to have limited application and should only be implemented after rigorous appraisal on a case-specific basis; otherwise it has the potential to accelerate declines of marine species currently threatened by fisheries bycatch" (Finkelstein, Bakker et al. 2008). May be useful, but only as a complementary measure (Žydelis, Wallace et al. 2009). |
| Pelagic longline |  |  |
| Seabirds (albatrosses and petrels) | Best | No single solution to avoid incidental mortality of seabirds in pelagic longline fisheries. Most effective approach is streamer lines combined with branchline weighting and night setting. Best practices are followed for line setting and hauling (e.g., SRWG 2010). |
| Night setting | Primary | Proven effective in Southern Hemisphere. Streamer lines and weighted lines should also be used when interacting with nocturnal birds/fishing during bright moon. |
| Streamer/scarer lines | Primary | Proven to be effective in North Atlantic. Should be paired and/or weighted lines in North Pacific. Paired lines need more testing. Light configuration not recommended. |
| Weighted branch lines | Primary | Must be combined with other measures. |
| Offal discharge management | Secondary | Not yet established but is thought to assist. |
| Sidesetting | Secondary | Insufficiently researched; there have been operational difficulties on some vessels. Effective in Hawaii in conjunction with bird curtain and weighted branch lines. Japanese research conclusions must be combined with other measures. Untested in Southern Hemisphere. |


| Line shooter and mainline tension, bait caster, live bait, thawing bait | - | Not recommended. |
| :---: | :---: | :---: |
| Underwater setting chute, hook design, olfactory deterrents, blue-dyed bait | - | Insufficient research. Blue-dyed bait may be only effective with squid bait. Results inconsistent across studies. |
| Turtles |  |  |
| Replacement of J and tuna hooks with circle hooks | Primary | Wide circle hook with </= 10-degree offset. |
| Bait change | Primary | Use of fish instead of squid. |
| Deep setting | Primary | Set gear deeper than turtle abundant depths (40-100m). |
| Fish bait hooking | Primary | Single hooking fish bait instead of threading hook through bait multiple times. |
| Temporal changes | Primary | Reduce soak time and haul during daylight. |
| Lights on gear | Secondary | Use of intermittent flashing light sticks instead of continuous use nonluminous gear. |
| Handling and release practices | Primary | To reduce mortality of caught turtles. |
| Sharks |  |  |
| Bait change | Primary | Fish instead of squid. |
| Prohibit wire leaders | Primary |  |
| Deeper setting | Primary | Avoid surface waters. |
| Shark repellants | - | Insufficient research. |
| Circle hooks |  |  |
| Marine mammals |  |  |
| Weak hooks, deterrents, echolocation disruption | - | Insufficient research. |
| Other finfish (including juvenile targets) |  |  |
| Circle hooks |  | May help reduce mortality of billfish and tunas. |
| Shellfish |  |  |
| Not problematic |  |  |
| Bottom longline (Many measures similar to pelagic longline) |  |  |
| Seabirds (albatrosses and petrels) | Best | No single solution to avoid incidental mortality of seabirds in demersal longline fisheries. No combination specified: assume streamers, weighted and night setting, or Chilean longline method (vertical line with very fast sink rates-considered effective even without other measures; widely used in South American waters and SW Atlantic). Best practices are followed for line setting and hauling (e.g., SRWG 2010). |
| Streamer/scarer lines | Primary | Effective, but must be used properly (streamers are positioned over sinking hooks). Better when combined with, e.g., night setting, weighting, or offal control. |
| Weighted lines | Primary | Must be combined with other measures, especially streamers, offal control and/or night setting. |
| Night setting | Primary | Same as pelagic. |


| Haul curtain (reduce bird access when line is being hauled) | Secondary | Can be effective but must use strategically as some birds become habituated. Must be used with other measures. |
| :---: | :---: | :---: |
| Offal discharge control <br> (discharge homogenized offal at time of setting) | Secondary | Must be used in a combo, e.g., with streamers, weighting, or night setting. |
| Side setting | Secondary | Insufficiently researched; there have been operational difficulties on some vessels. |
| Hook design, olfactory deterrents, underwater setting chutes, blue-dyed bait, thawed bait, use of line setter | - | Insufficiently researched. Blue-dyed bait, thawed bait, and use of line setter not relevant in demersal gear. |
| Turtles, sharks, mammals, other finfish, shellfish |  |  |
| See pelagic longlines |  |  |
| Trawl |  |  |
| Seabirds (albatrosses and petrels) | Best | Little work has been done on seabird bycatch mitigation in trawl fisheries (pelagic and demersal). There is no single solution to avoid incidental mortality of seabirds in trawl fisheries. The most effective approach is offal discharge and discards control, through full retention of all waste or mealing (the conversion of waste into fish meal reducing discharge into sump water) plus streamer lines. Effectiveness of other offal control measures such as mincing and batching is not clear. |
| Limited waste control | Minimum requirement for + modifier | No discharge of offal or discards during shooting and hauling. |
| Reduce cable strike through bird scaring wires or snatch block | Primary | Scarers recommended even when offal/discard management is in place. Snatch block recommended on theory. |
| Reduce net entanglement through net binding, net weights, net cleaning |  | Recommended on theory. |
| Net jackets | - | Not recommended. |
| Reduced mesh size, acoustic scarers, warp scarers, bird bafflers, cones on warp cables | - | Effectiveness not yet established. |
| Turtles |  |  |


| Turtle excluder device (TED) | Primary | Any modification to the trawl to reduce the capture of turtles, principally in tropical/subtropical shrimp trawls. Typically a grid or large-hole mesh designed to prevent turtles from entering the codend. The only designs approved for use in the US warm-water shrimp fisheries are hard TEDs (i.e., "hooped hard TEDs" such as NMFS, Coulon and Cameron TEDs, "single grid hard TEDs" such as the Matagorda, Georgia, or Super Shooter TED, and the Weedless TED) and the Parker Soft TED (the latter only in offshore and inshore waters in Georgia and South Carolina). Hard TEDs that are not approved for use in the shrimp fisheries are used in the Atlantic summer flounder bottom trawl fishery. TEDs must be used in conjunction with escape hatches, which also vary in size and design. More details on TED/hatch designs and US regulations can be found in Eayers (2007). |
| :---: | :---: | :---: |
| Sharks |  |  |
| TED |  | TEDs generally allow large animals to escape, e.g., sharks (Belcher and Jennings 2010). Highly variable depending on net type and TED used. BRD made little difference (fisheye). |
| Marine mammals |  |  |
| TED/BRD |  | Grids generally allow large animals to escape. |
| Other finfish |  |  |
| Bycatch reduction device <br> (BRD): Catch separators |  | A BRD is any modification designed principally to exclude fish bycatch from shrimp trawls. Catch separator designs include hard grids (e.g., Nordmore grid) and soft mesh panels attached at an angle inside the trawl net as well as the Juvenile and Trash Excluder Device (JTED), which has a grid/mesh design partially covering the inside of the trawl net. Hard grids are generally seen as more effective than soft panels. Effectiveness of JTED unknown. |
| BRD: Active swimmer escape hatches |  | Designed for strong-swimming fish to actively escape (shrimp are more passive swimmers). Most are located in the codend (e.g., fisheye and fishbox) although others can be in the body of the trawl (square mesh window, composite square mesh panel, radial escape section). |
| BRD: Square-mesh codend |  | Square mesh stays open under tow (unlike diamond mesh). |
| BRD assist |  | E.g., the cone. Stimulates fish to swim forward through escape hatches like the fisheye, square mesh window or radial escape section. |
| Coverless trawl |  | Inclusion of increased mesh sizes in the upper wings and upper netting panel immediately behind the headrope crown, coupled with reduced headline height, encourages the escape of fish species such as haddock and whiting in and around the mouth of the trawl. |
| Rigging modification |  | Triangular/diamond-shaped cut in the top of the codend (e.g., flapper), changes to ground chain settings, headline height reduction, a length of twine stretched between the otter boards to frighten fish, large mesh barrier across trawl mouth and large cuts in the top panel of the net ahead of the codend. |
| Semi-pelagic rigging |  | Avoid contact with seabed. |
| Trawl separator (Rhule trawl) |  | Reduces cod catch in haddock trawls by separating catch and releasing cod from the net. |
| Shellfish |  |  |


| TED |  | TEDs generally allow large animals to escape (jellyfish). Downward <br> facing TEDs may also allow benthic invertebrates to escape. |
| :--- | :--- | :--- |
| BRD e.g., Nordmore grid |  | Effective for crabs (Noell et al. 2018) |
| Rigging modification |  | Longer sweeps between the otter board and trawl can reduce <br> invertebrate bycatch. |
| Semi-pelagic rigging |  | Avoid contact with seabed. |
| Other |  |  |
| BRD |  | Seahorses, sea snakes in Australian prawn fisheries. |
| Gillnet |  |  |
| Seabirds |  | Less research than for trawls. |
| Visual and acoustic alerts | - | High visibility panels (upper portion or checkerboard), dropped cork <br> lines for shallow diving spp., attending nets (Wiedenfeld 2015). |
| Turtles | Primary | Reduces entanglement as the net is stiffer. Good for both demersal <br> and drift nets. |
| Use lower profile nets | Primary |  |
| hotspots. |  |  |


| Use of modified FAD <br> designs | - | Insufficient research |
| :--- | :--- | :--- |
| Marine mammals |  |  |
| Backdown maneuver, <br> Medina panel, deploy <br> rescuers | Primary |  |
| Avoid mammals |  | Restrict setting on mammals. |
| Other finfish | - |  |
| Sorting grids |  | Insufficient research. |
| Avoid finfish |  |  |
| Shellfish |  |  |
| Not problematic | Primary | E.g., Diamonding on FADs. <br> and Heinrich 2005). |
| Pots and traps |  |  |
| Turtles | Primary | E.g., northern right whales, NE lobster fishery. |
| BRDs |  |  |
| Marine mammals | Primary |  |
| Weak lines |  |  |
| Finfish, invertebrates |  |  |
| BRDs |  |  |

## Appendix 5 - Impact of fishing gear on the substrate

To assess fisheries for habitat impacts under the Seafood Watch ${ }^{\circledR}$ criteria, we developed a matrix to help determine the potential impacts that different fishing gear may have on various habitat types. The matrix was developed based on similar work done by the New England Fisheries Management Council (NEFMC 2010) and the Pacific Fisheries Management Council (PFMC 2005).

The NEFMC (2010) created a "Swept Area Seabed (SASI) model" that assessed habitat susceptibility and recovery information. Susceptibility and recovery were scored (0-3) based on information found in the scientific literature and supplemented with professional judgment when research results were deficient or inconsistent.
"Vulnerability was defined as the combination of how susceptible the feature is to a gear effect and how quickly it can recover following the fishing impact. Susceptibility was defined as the percentage change in functional value of a habitat component due to a gear effect, and recovery was defined as the time in years that would be required for the functional value of that unit of habitat to be restored (ASFMC 2010)."

The PFMC (2005) created a similar habitat sensitivity scale (0-3) that represents the relative sensitivity of different habitats to different gear impacts. The sensitivity of habitats from the PFMC (2005) was based on actual impacts reported in the scientific literature.

The relative impacts by gear and habitat type used for the Seafood Watch ${ }^{\circledR}$ matrix were based on the sum of sensitivity and recovery values from tables developed by the NEFMC (2010) (substrates) and the PFMC (2005) (biogenic). The NEFMC (2010) excluded deep-sea corals with extreme recovery times. The values for deep-sea corals in this matrix are the sum of the sensitivity and recovery scores from PFMC (2005). The following other biogenic habitats that were not included in the NEFMC (2010) data tables include: seagrass, sponge reefs (rather than individual sponges) and maerl beds. Due to the slow recovery and importance of these habitat types, they have been given the same value as coral and sponge habitats, all of which are listed as "biogenic."

Hall-Spencer and Moore (2000) examined the effects of fishing disturbance on maerl beds. Maerl beds are composed of a calcareous alga and form complex habitats with a high degree complexity. The associated species assemblages have high diversity (Hall-Spencer and Moore, 2000). Hall-Spencer and Moore (2000) showed that four years after an initial scallop-dredging disturbance had occurred, some fauna, such as the bivalve Limaria hians, had still not re-colonized the trawl tracks. Similarly, work by Sainsbury et al. (1998; in Kaiser et al. 2001) suggests that recovery rates may exceed fifteen years for sponge and coral habitats off the western coast of Australia.

Hydraulic clam dredges are rated as a high concern according to Seafood Watch ${ }^{\circledR}$. There are very few studies on the impact of this gear type, so we have relied on expert opinion (NEFMC 2010). Hydraulic clam dredges are used primarily in sand and granule-pebble substrates because they cannot be operated in mud or in rocky habitats (NEFMC 2010). This gear type is effective at pulverizing and/or removing solids and flattening out seafloor topography (NEFMC 2010). In addition, the habitats where this gear type is used are very susceptible to hydraulic dredges; recovery is moderate on average (NEFMC 2010). This leads Seafood Watch ${ }^{\circledR}$ to rate hydraulic dredges as "high concern." Hydraulic dredges do not operate on deep-sea coral or other biogenic habitats.

Neckles et al. (2005) found significant differences in eelgrass biomass between disturbed and reference sites up to seven years after dragging. The authors projected that it would require a mean of 10.6 years for eelgrass shoot density to recover in areas of intense dragging.

Demersal seines were not evaluated in the reports by Fuller et al. (2008), Chuenpagdee et al. (2003), NEFMC (2010) or PFMC (2005). Demersal seines include: Danish seines, Scottish fly-dragging seines and pair seines. These seines are similar to some bottom trawl gear in that they have a funnel shaped net with a groundrope. They are generally hauled by wires or ropes, and although they are lighter than some bottom trawl gear, they create habitat disturbance (Rose et al. 2000; Thrush et al. 1998; Valdemarason and Suuronen 2001). A review of trawling impacts by Jones (1992) grouped bottom trawling, dredges and Danish seines together as having similar impacts on the. However, studies have demonstrated Danish seines to have less impact on the substrate compared to bottom trawls (Gillet 2008). Therefore, in our matrix they are given an intermediate score as more damaging that bottom longlines and bottom gillnets, but less damaging than bottom trawls. Beam trawls also were not included in the reports but were considered to be similar to otter trawls.

The matrix developed from the sources referenced above is shown on the next page. For use in evaluating the Fisheries Criteria, these data have been summarized into categories (low impact, moderate, moderate-severe, severe, and very severe) to simplify use of the table.

Habitat impacts matrix: Relative impacts by gear and habitat type.

|  | Mud |  | Sand |  | Granule-pebble |  | Cobble |  | Boulder |  | Deep-sea corals ** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | low | high | low | high | low | high | low | high | low | high |  |
| Line, Vertical (BL/2) | 0.5 | 0.5 | 0.6 | 0.5 | 0.8 | 0.8 | 1.0 | 0.9 | 1.0 | 1.0 | 1.3 |
| Longline, Bottom**** | 0.7 | 0.7 | 0.9 | 0.8 | 1.4 | 1.3 | 1.6 | 1.5 | 1.7 | 1.7 | 2.0 |
| Trap (lobster and deep-sea red crab) | 1.3 | 1.3 | 1.2 | 1.2 | 1.8 | 1.7 | 2.0 | 1.9 | 2.1 | 2.1 | 1.3 |
| Gillnet, Bottom**** | 1.3 | 1.3 | 1.5 | 1.4 | 2.0 | 1.9 | 2.2 | 2.1 | 2.3 | 2.3 | 3.0 |
| Bottom Longline, Gillnet | 1.0 | 1.0 | 1.2 | 1.1 | 1.7 | 1.6 | 1.9 | 1.8 | 2.0 | 2.0 | 2.5 |
| Seine, Bottom (BL,G+TBO/2) | 1.8 | 1.7 | 2.0 | 1.9 | 2.5 | 2.3 | 2.7 | 2.5 | 2.6 | 2.6 | 3.6 |
| Trawl, Shrimp (BS+TBO/2) | 2.2 | 2.1 | 2.5 | 2.3 | 3.0 | 2.6 | 3.0 | 2.8 | 3.0 | 2.9 | 4.1 |
| Trawl, Bottom Otter | 2.6 | 2.4 | 2.9 | 2.7 | 3.4 | 2.9 | 3.4 | 3.1 | 3.3 | 3.2 | 4.6 |
| Dredge, New Bedford Scallop | 2.6 | 2.4 | 3.0 | 2.8 | 3.5 | 3.0 | 3.5 | 3.2 | 3.3 | 3.2 | 5.1 |
| Dredge, Hydraulic Clam | $\mathrm{n} / \mathrm{a}$ |  | 4.4 | 4.0 | 4.9 | 4.5 |  |  | n/a |  |  |
| Explosives/Cyanide | 6 |  | 6 | 6 |  | 6 | 6 | 6 | 6 | 6 |  |

* Shrimp trawls tend to be lighter than bottom otter trawls for fish and do not need to touch the seabed to be effective.
** Most biogenic habitats (macroalgae, cerianthid anemones, polychaetes, sea pens, sponges, mussel and oyster beds) are incorporated into the scores for each substrate/gear combination in the table. NEFMC 2010 specifially excluded deep-sea corals. The numbers for deep-sea corals in this matrix are the sum of the sensitivity and (standardized) recovery scores in PFMC 2005. Other biogenic habitats that were not included in the NEFMC data tables include seagrass meadows, sponge reefs (rather than individual sponges) and maerl beds. Use the 'deep-sea corals' column for these habitats.
*** Scores not determined for hydraulic dredges in these habitats as the gear is assumed to not operate in them (NEFMC 2010).
**** NEFMC 2010 groups bottom longlines and gillnets as 'fixed gear' (not shown in table). These scores have been disaggregated here for substrate habitats only by adding 0.4 to the aggregated score for gillnets and subtracting 0.4 for longlines, base don the relative impacts shown in PFMC 2005 (i.e. that gillnets are generally more damaging than longlines).

The values above are the sum of sensitivity and recovery values in tables from Section 5.2 in Part 1 of (NEFMC 2010) (substrates) and Tables 4 and 5 in Appendix C, Part 2 of PFMC (2005) (biogenic). Gear types in black are from the Swept Area Seabed Impact (SASI) model used for the NEFMC EFH process (NEFMC 2010). Gear types in red are derived from those in black. Substrate types are self-explanatory except that mud includes clay-silt and muddy sand, and boulder includes rock. The energy regime is used here as a proxy for natural disturbance, with a cutoff between low and high stability at 60 m depth. Most biogenic habitats (macroalgae, cerianthid anemones, polychaetes, sea pens, sponges, mussel and oyster beds) are incorporated into the scores for each substrate/gear combination in the table. NEFMC (2010) specifically excluded deep-sea corals with extreme recovery times. The numbers for deep-sea corals in this matrix are the sum of the sensitivity and (standardized) recovery scores from PFMC (2005). Other biogenic habitats that were not included in the NEFMC data tables include seagrass meadows, sponge reefs (rather than individual sponges) and maerl beds.

## Appendix 6 - Gear modification table for bottom tending gears

## Spatial protection

Reducing the footprint of fishing through spatial management can be one of the most effective ways to mitigate the ecological impact of fishing with habitat-damaging gears (Lindholm et al. 2001; Fujioka 2006). The relationship between gear impacts, the spatial footprint of fishing and fishing effort (i.e., frequency of impact) is complex (Fujioka 2006) and cannot be quantified precisely in Seafood Watch® ${ }^{\circledR}$ assessments. Nevertheless, the criteria acknowledge the benefits of conservative habitat protection efforts by adjusting the habitat score. Thresholds for adjusting the habitat score due to habitat protection from the gear-type used in the fishery ( $50 \%$ protected to qualify as "strong mitigation" and $20 \%$ protected to qualify as "moderate mitigation") are based on recommendations for spatial management found in the scientific literature as noted in Auster (2001). To minimize impacts on vulnerable species and sensitive habitats, Auster (2001) recommends employing the precautionary principle when a threshold level of $50 \%$ of the habitat management area is impacted by fishing, with a minimum of $20 \%$ of regions in representative assemblages and landscape features protected in MPAs.

The table below gives examples of gear modifications that are believed to be moderately effective at reducing habitat impacts based on scientific studies. This table will be continually revised as new scientific studies become available. The main sources for the current table are He (2007) and Valdemarsen, Jorgensen et al. (2007).

| Gear | Modification |
| :--- | :--- |
| Otter | Semi-pelagic trawl rigging (trawl doors, sweeps and bridles off the bottom, also includes <br> Trawls <br> modifications such as short bridles and sweeps-most commonly used for shrimp, <br> nephrops and other species that are not herded by sand clouds and bridles due to poor <br> swimming ability) |
|  | Quasi-pelagic trawl rigging/sweepless trawls (trawl doors remain in contact with the <br> seafloor, remaining gear largely off the bottom, e.g., whiting in New England, flatfish in <br> Alaska, red snapper in Australia) |
|  | Lighter ground gear (e.g., fewer bobbins) |
|  | Use of rollers instead of rockhoppers |
|  | Trawl door modifications such as high aspect (smaller footprint), cambered (generally for <br> fuel efficiency) or soft doors (e.g., self-spreading ground gear) |

## Appendix 7 - Data-limited assessment methods

This appendix offers guidance for scoring Criterion 1 with limited data for a stock.
Note: This guidance is provided for illustration purposes only. Expert input and case-by-case interpretation are necessary to ensure the assessment indicator is appropriate in the context of the specific fishery, and interpretation should take into account specific factors or changes in the fishery that may affect results (e.g., demand-driven factors that affect size of the catch).


When considering how to proceed with limited data, first determine if a quantitative assessment exists and follow appropriate guidance once determined.

## A quantitative assessment is available:

- And has been reviewed and accepted by a scientific committee where reference points have been determined, follow the standard scoring procedures.
- And has been reviewed and accepted by a scientific committee where reference points have not been determined, consider whether management has given the fishery an official status (i.e. overfished, not overfished).
- If an official status has been determined by management score according to criteria.
- If an official status has not been determined, proceed to following guidance as if there is no quantitative assessment available (see guidance above).
- And was rejected by a scientific committee, proceed to following guidance as if there is no quantitative assessment available (see guidance above).


## A quantitative assessment is not available:

- But a data limited assessment is available offering confident results, score accordingly as Low Concern to High Concern.
- But a data limited assessment is available; however, there is a high level of uncertainty that causes low confidence in the results of the assessment.
- When another data limited assessment is available, consider confidence level in the results of the assessment, if confident, score accordingly from Low Concern to High Concern.
- When no other data limited assessment is available, conduct a ProductivitySusceptibility Analysis (PSA) to determine species vulnerability to fishing pressure.

If a data limited assessment is NOT available, conduct a PSA to determine species vulnerability to fishing pressure.

Data limited stock assessments typically have high levels of uncertainty that can affect the confidence in the results. Confidence can be gained by evaluating how the data were collected (was sound scientific method used?), the type of data limited assessment, and the assumptions associated with the assessment approach. When assumptions have not been met and confidence in the results may be low it is important to consider whether there are other data-limited approaches that have been used for the same stock or species. Multiple assessments that show a similar result can increase confidence. It is also important to identify whether a scientifically accepted, peer-reviewed method was used. Further confidence can be gained through independent review of the data-limited assessment and results. Analysts should communicate with Seafood Watch staff to ensure accurate and consistent interpretation of data-limited assessment results.

Several published papers and online resources list potential data-limited methods ranging from very simple metrics to more data-moderate approaches and describe their use, their underlying assumptions, and their limitations. Analysts can refer to Dowling et al. 2019, and EDF's online tool, Framework for Integrated Stock and Habitat Evaluation's (FISHE) summary document as well as its primers on LengthBased Assessment Methods, Froese Length-Based Sustainabibility Indicators, Primer for Cope and Punt Length-Based Reference Point Method. The private data-limited decision supports FishPath; Dowling et al. 2016, while not currently available to SFW analysts, houses a comprehensive reference library of data-limited assessment methods. It may be made public while this Fisheries Standard version is in use.

Additional resources that provide overviews of data limited metrics are provided in the reference section at the end of this document.

## Appendix 8 - Forage Fisheries Guidance

Fisheries that catch forage species, as defined in the glossary and included below, receive special consideration under the Seafood Watch Wild Capture Fisheries Standard due to the exceptional role that these species play in the ecosystem. This Appendix is a guide to assessing these fisheries under the Fisheries Standard

## Forage Species (glossary term)

Forage species play an important role in food webs because they 1) exhibit high connectance ${ }^{9}$ to other organisms in the ecosystem and 2) a large amount of energy is channeled through that species. Forage species typically exhibit highly variable productivity, ${ }^{10}$ such that there may be high uncertainty in their reference points, making it difficult to evaluate their stock status. The drivers of this variability in productivity may be environmental forcing and/or other factors. As a result of their importance in food webs these stocks require management that is tailored to their specific life histories and ecological roles.

Species that generally qualify as forage species include sandeels, sandlances, herrings, menhaden, pilchards, sardines, sprats, anchovies, krill, lanternfish, smelts, capelin, mackerels, silversides, sand smelts, Norway pout (adapted from MSC Fisheries Standard V2.01, p. 14). Other species or stocks may qualify if they meet the definition above. Due to differences in food web structure and function between marine and aquatic systems (including species richness and the effect of ecosystem size), additional information may be required to determine whether freshwater species qualify as forage species.

We note that in some food webs, several species may fulfill the ecological role of forage species as a guild rather than as single forage species.

## Does the fishery catch forage species?

Once the Criterion 1 and Criterion 2 species are determined for the assessment, the analyst, in consultation with their internal SFW reviewer, determines whether a species qualifies as a forage species (either as an individual stock or as a member of a guild) based on the definition and its intent. To make this determination, the analyst will search for evidence to show that within the ecosystem, the species fulfills both of the criteria in the definition: 1) it exhibits high connectance and 2 ) it serves as a channel for a large amount of energy. There are example species in the definition, including various small pelagic fish and krill; some species of squid may also meet this definition. In cases where there are insufficient data to determine if these conditions are met, the analyst should apply the definition based on its intent and work with the internal SFW reviewer to seek out expert knowledge to determine if a stock will be assessed as a forage species. We note that there will be some species within the groups listed in the definition that will not qualify for consideration as forage species under the Seafood Watch Fisheries Standard.

[^6]To understand and apply the concept and intent of "connectance," the analyst should read the paper cited in the definition; see Plagányi and Essington (2014), When the SURFs up, forage fish are key. Fisheries Research 159:68-74.

## Assessing Criterion 1:

Acknowledging that static reference points based on virgin biomass do not adequately characterize stock status for forage species whose productivity and biomass fluctuate with environmental conditions, SFW considers forage species abundance and fishing mortality as highly uncertain unless reference points are "appropriate for the species" (see glossary definition hyperlinked to this term and included below) or there is a formal status that the stock is overfished or overfishing is occurring. The analyst must have a thorough understanding of this term and its intent in order to score C1.1 and C1.2 correctly. To do this SFW requires analysts assessing forage fisheries to read Siple et al. (2018), cited in the definition, and discuss this concept with their internal SFW reviewer. It is important for the analyst to understand that $\mathrm{C1}$ is scored based solely on abundance and fishing mortality of the stock, not on whether abundance is high enough and fishing mortality low enough to ensure that the needs of the ecosystem are met, namely for dependent predators. Ecosystem needs are evaluated under 4.3, the Ecosystem Based Fisheries Management factor.

## Appropriate for the Species (glossary term):

Whether a reference point is appropriate for a species depends on its life history characteristics, its productivity dynamics and its role in the ecosystem.

Regarding forage species: Most modern assessments use a stock-recruitment curve that is described by stationary parameters, including virgin biomass or $B_{0}$ and are not appropriate for species with dynamic productivity that shifts in response to environmental conditions. While it is possible to calculate reference points based on dynamic virgin biomass (acknowledging that the carrying capacity of the environment for these species is different based on favorable to unfavorable environmental conditions), to date, none exist in practice for any species and the effectiveness of dynamic reference points is not well understood. While static reference points do not describe the shifts in productivity of forage species (instead, at best, they represent a long term average), they can be used effectively in management when 1) the harvest strategies based upon them account for volatility AND 2) when the harvest strategy outcomes have been tested using a proven, robust Management Strategy Evaluation framework, demonstrating that fishing mortality is set low enough to prevent collapse during periods of low stock productivity. Given these considerations, unless harvest strategies account for volatility and have been tested and proven to prevent stock collapse (i.e., in most situations), Seafood Watch considers forage stock biomass and fishing mortality to be highly uncertain.

Note that the best reference point to minimize the probability and severity of collapse for forage species depends on the specific attributes of the species. See Siple, Essington, and Plagányi (2019), Forage fish fisheries management requires a tailored approach to balance trade-offs. Fish and Fisheries 20(1): 110-124.

If a forage species is included as a C1 species, analysts must consider the following when assessing C1.1 - Abundance and C 1.2 - Fishing Mortality:
a. How a forage species scores under 1.1 depends on the results of the stock assessment, whether it is a formal assessment or a data limited assessment (see guidance in Appendix 7 on choosing appropriate data limited indicators), and the confidence that those results accurately characterize the health of a forage species stock's abundance, factoring in how that species' biomass fluctuates with environmental variability through periods of low to high productivity. While it is reasonably straightforward to calculate a stock's abundance (provided that surveys are carried out over the full extent of the stock's range at the appropriate frequency), it is not as straightforward to determine whether that abundance is at a healthy level due to the highly fluctuating nature of these species' biomass and productivity, which varies with forage species type, based on their life history characteristics. The analyst must determine if this variability is correctly factored into the application of a biomass reference point/indicator or if the uncertainty is above this level. Where uncertainty is above this level such that there is low confidence in the reference point, which SFW anticipates will be the case in most traditionally managed forage fisheries, the analyst will score as 1.1 moderate concern. If there is a formal determination that the stock is overfished the analyst will score 1.1 as high concern.
b. Similarly to 1.1, how a forage species scores under 1.2 depends on the results of the stock assessment (whether a formal assessment or a data limited assessment) and the confidence that those results accurately characterize fishing mortality of a highly fluctuating stock (see 1.1 above). To score low concern for fishing mortality, there must be evidence that fishing mortality is set low enough to prevent collapse during periods of low productivity and that this level is not exceeded. A robust Management Strategy Evaluation (MSE) that tests the outcomes of the harvest control rule and finds that fishing mortality is set appropriately can provide this evidence. Alternatively, if the analyst can find evidence via consensus in the peer reviewed literature that fishing mortality is so low such that the fishery has no measurable impact on stock health, $\mathrm{s} / \mathrm{he}$ may assign a low concern score. Where this evidence is not available, as SFW anticipates will be the case in most traditionally managed forage fisheries, the analyst will score as moderate concern for 1.2. If there is a formal determination that overfishing is occurring the analyst will score 1.2 as high concern.

## Assessing Criterion 2

If a forage species is included as a C2 species and the fishery is a substantial contributor to fishing mortality of that species, the same guidance described under C1 above applies for C2.

## Assessing Criterion 3

For a fishery that targets or retains forage species, the analyst must address special items under 3.1 and 3.3 and other considerations under each of these factors that are applicable to all fisheries, but carry more importance for forage species based on their life history characteristics.

Specific to forage species, under 3.1 - Management Strategy and Implementation, the analyst needs to determine whether the harvest control rule adequately addresses the fluctuations in biomass and productivity to ensure that catch is sustainable for that stock. This is not the same as the ecosystem considerations under 4.3, which ask the analyst to determine whether precautionary buffers are in place to ensure that dependent predators' needs are met. Harvest control rules, to be effective for forage fisheries, must enable these stocks to recover from periods of low productivity, when these stocks are
most susceptible to collapse. See (Siple, Essington, and Plagányi 2019), "Forage fish fisheries management requires a tailored approach to balance trade-offs," Fish and Fisheries 20(1): 110-124).

Specific to forage species, under 3.3 - Scientific Data Collection and Analysis, the analyst needs to determine whether stock assessments are conducted with sufficient frequency to account for the stock's fluctuations in productivity and or biomass. This will vary based on the assessed species' life history attributes, which result in different types of oscillations in productivity and/or biomass; see Siple, et al. (2018) referenced in the above paragraph).

General to all fisheries but likely more critical for forage species, is the language and guidance in 3.1 on flexible and resilient fisheries management. Especially important is management's responsiveness to the state of the stock (i.e., stock status, whether a stock is overfished or overfishing occurring) within appropriate timeframes ${ }^{11}$ and (under a separate numeric consideration in the scoring tables for 3.1) management's responsiveness to changes in stock productivity and or biomass.

## Assessing Criterion 4 (4.3)

For forage species, special considerations apply when scoring 4.3. Given the key role that forage species play in ecosystems, fisheries for these species (unless the take is an insignificant portion of the fished stock and it has been proven that the take has no measurable impact on the stock's biomass) are expected to detrimentally impact food webs. If detrimental food web impacts are possible, the fishery cannot score better than Low concern for EBFM.

To score 4.3 for forage species (whether they are included in C1 or if in C2 and the fishery is a substantial contributor to their fishing mortality), the analyst will collect information to understand the following:

1) the spatial and temporal management of the fishery relative to the scale of the fishery and the role that the species plays in the ecosystem and any other policies in place to protect ecosystem function. This will include, but is not limited to the placement, spatial extent and regulations (including the level of protection) set for marine protected areas and the timing and extent of temporal closures of the fishery.
2) The information tier the fishery would meet based on the Lenfest Forage Fish Task Force (LEFTF) Recommendations (low, intermediate or high information) and whether the harvest control rules in place are appropriate for that information tier, ensuring sufficient forage in the water so that dependent predators can thrive (so that their growth is not resource limited). Guidance for determining an appropriate ecological harvest control rule is provided in the glossary (see hyperlinked definition or below)

## Ecological harvest control rule (glossary term):

For certain taxa, like forage species that have an exceptionally important role in the ecosystem, harvest control rules (HCRs) should be based on ecosystem considerations (i.e., maintaining enough biomass to allow the species to fulfill its ecological role), rather than MSY or single-species

[^7]considerations. For forage species, HCRs should be consistent with the precautionary principles and recommendations of the Lenfest Forage Fish Task Force (see specific guidance under the "Lenfest Forage Fish Task Force" entry. See also the fact sheet regarding ecological reference points at https://www.lenfestocean.org/-/media/assets/2018/10/buhheister-fact-sheet-pdf.pdf).

To score low concern for 4.3, a forage fishery (which by default has the potential to incur detrimental impacts on the ecosystem) must fulfill both requirements 2 a and 2 b . For 2 a , the analyst must comprehensively survey the scientific literature and provide clear evidence in his/her response as proof that precautionary and effective spatial management appropriate to the scale of the fishery and ecology of the stock is in use. Seafood Watch recognizes that most forage fisheries, based on current best management practices in the industry, will not meet the requirements for low concern. The LFFTF consistency requirement is the same for low and moderate concern.

To score moderate concern for 4.3, a fishery that is a substantial contributor to fishing mortality for a forage fish species (either under C 1 or C 2 ) must meet requirement 1 . While similar to the low concern language for 4.3, the conservation bar is lower for moderate concern. Spatial and temporal management must be appropriate to the scale of the fishery and ecology of the stock for both scoring categories (low and moderate concern), but for moderate concern, management is "likely" to be effective and there should be little scientific controversy around that likelihood, whereas for low concern the analyst must provide evidence from the scientific literature proving effectiveness. As noted in the paragraph above the LFFTF consistency requirement is the same for low and moderate concern.

To score high concern for 4.3, a forage fishery's spatial and temporal management or other policies to protect ecosystem function are not appropriate for the scale of the fishery and the ecology of the stock OR the harvest control rule is not consistent with the LFFTF recommendations. If strong ecosystem impacts are not integrated into management of the forage fishery, 4.3 will score as high concern. If the fishery scores as high concern score for 4.3, C4 is overall red.

To score critical for 4.3 the analyst must provide evidence that detrimental food web impacts are resulting from the fishery. These impacts might include trophic cascades or alternative stable states. Evidence must be credible and drawn from published, scientific, peer reviewed literature. Due to a critical score resulting in an overall avoid recommendation, the analyst must make this determination in consultation with their internal SFW reviewer and peer reviewers must be consulted on this decision.

## Appendix 9 - Document Revision History

VF3.2 of the standard was revised in February 2020 as summarized below. The resulting document is VF4 (this document).

## Criterion 1

- Changed the productivity and susceptibility factors assessed in the PSA and changed the scoring methodology of the PSA. The changes more closely align the PSA with the literature and help ensure that data poor fisheries do not score better than data rich fisheries.
- Delineated the species-specific characteristics that trigger use of the new forage species scoring criteria.
- Resolved challenges in scoring abundance and fishing mortality of forage species based on static reference points while ensuring that fishing mortality thresholds are properly set.
- Revised guidance on reference points to ensure that their appropriateness is judged based on principles of conservation and system resilience, rather than a comparison to MSY.
- Amended the use of data-limited stock assessments such that where there is confidence in the results of a data-limited assessment, a single assessment can be used to score Factor 1.1, independent of a PSA.
- Clarified the definition of a substantial contributor to fishing mortality and created a decision tree to identify whether a fishery is a 'substantial contributor' or a 'nonsubstantial contributor'.


## Criterion 2

- Clarified the process of selecting main species by developing a decision tree that identifies the questions to ask when making this determination.
- Created a scoring table that uses the preliminary Marine Mammal Protection Act List of Foreign Fisheries (LOFF) to help score impacts to marine mammals by non-US fisheries. Once the final LOFF is developed, marine mammal bycatch from imported fisheries will be scored similarly to the way U.S. fisheries are scored in Table 2.2.1.a.
- Developed additional guidance in Criterion 2 outlining when the impact of bait fisheries should be included in a Seafood Watch report.
- Created a methodology to revise the Unknown Bycatch Matrices. Seafood Watch is hosting a series of workshops to elicit expert opinion regarding bycatch susceptibility of finfish, forage fish, batoids, invertebrates and habitat forming species.


## Criterion 3

- Added language to the standard and guidance to Appendix 3 to ensure precautionary, flexible and resilient management practices are considered in the assessment. These edits were made to improve Seafood Watch's assessment of forage fish management and account for the impacts of climate change in management and are relevant to all fisheries.
- Added scoring options under moderately effective and ineffective for Factor 3.1 to better capture those situations where there is uncertainty regarding the implementation of a management system.
- Within factor 3.2 for ineffective, added specific consideration for bait fisheries management when it is known to be poor (to be considered when bait species are identified as a main species in Criterion 2).
- Modified Factor 3.2 Bycatch Strategy to incorporate the best practices recommended by Global Ghost Gear Initiative.
- Amended Factors 3.3 and 3.4 titles and clarified their contents to distinguish 'monitoring' from 'surveillance' to be more consistent with the FAO and Marine Stewardship Council (MSC) definitions. The term 'surveillance' will exclusively be used in Factor 3.4 to refer to observations required to obtain information about compliance with laws and regulations.
Criterion 4
- To better account for the ecosystem level impacts of fishing on forage species, Seafood Watch created a new decision rule to reflect the importance of forage species to the ecosystem; if factor 4.3 scores as high concern, Criterion 4 is considered red. SFW added a critical scoring option for demonstrable ecosystem impacts resulting from the fishery. This critical score would result in an overall rating of Avoid.
- Moved the Lenfest Forage Fish Task Force recommendations on conservative harvest control rules necessary ensure sufficient forage for ecosystem needs (namely the needs of dependent predators) from C 1 to C4.3.
- Added language to 4.3 better account for spatial and temporal management appropriate to the scale of the fishery and ecology of the stock.


## Appendices References

## Appendix 1

Botsford, L. W., and A. M. Parma. 2005. Uncertainty in Marine Management. Pages 375-392 in E. A. Norse and L. B. Crowder, editors. Marine Conservation Biology: The Science of Maintaining the Sea's Biodiversity. Island Press, Washington, DC.

Clark, W.G. 1991. Groundfish exploitation rates based on life history parameters. Canadian Journal of Fisheries and Aquatic Sciences 48:734-750.

Clark W. G. 2002. F 35\% revisited ten years later. North American Journal of Fisheries Management 22:251-257.

Cope, J. M. and A.E. Punt. 2009. Length-based reference points for data-limited situations: applications and restrictions. Marine and Coastal Fisheries: Dynamics, Management, and Ecosystem Science 1:169186.

Dowling N. A., Smith A. D. M., Smith D. C., Parma A. M., Dichmont C. M., Sainsbury K., Wilson J. R, Dougherty, D.T., Cope J.M. 2019. Generic solutions for data-limited fishery assessments are not so simple. Fish and Fisheries 20: 174-188.

Dowling, N.A., J.R. Wilson, M.B. Rudd, E.A. Babcock, M. Caillaux, J. Cope, D. Dougherty, R. Fujita, T. Gedamke, M. Gleason, N. Gutierrez, A. Hordyk, G.W. Maina, P.J. Mous, D. Ovando, A.M. Parma, J. Prince, C. Revenga, J. Rude, C. Szuwalski, S. Valencia, and S. Victor. 2016. FishPath: A Decision Support System for Assessing and Managing Data- and Capacity- Limited Fisheries. In: T.J. Quinn II, J.L. Armstrong, M.R. Baker, J. Heifetz, and D. Witherell (eds.), Assessing and Managing Data-Limited Fish Stocks. Alaska Sea Grant, University of Alaska Fairbanks.

Environmental Defense Fund (EDF). 2016. FISHE - Framework for Integrated Stock and Habitat Evaluation. Available at: http://fishe.edf.org/

Froese, R. 2004. Keep it simple: three indicators to deal with overfishing. Fish and Fisheries 5:86-91.
Froese, R., T.A. Branch, A. Proelß, M. Quaas, K. Sainsbury, and C. Zimmermann. 2010. Generic harvest control rules for European fisheries. Fish and Fisheries: doi:10.1111/j.1467-2979.2010.00387.x

Goodman, D., M. Mangel, G. Parkes, T. Quinn, V. Restrepo, T. Smith and K. Stokes. 2002. Scientific Review of The Harvest Strategy Currently Used in The BSAI and GoA Groundfish Fishery Management Plans, North Pacific Fishery Management Council, Anchorage, AK. 153 p. Available at: https://pdfs.semanticscholar.org/30ab/co9298fc4fc13d017c697511aa1bd11bff39.pdf

Honey, K.T., J.H. Moxley and R.M. Fujita. 2010. From rags to fishes: data-poor methods for fishery managers. Managing Data-Poor Fisheries: Case Studies, Models \& Solutions 1:159-184.

ICES 2010. General context of ICES advice. Available at:
https://stecf.jrc.ec.europa.eu/c/document library/get file?uuid=0b0bd556-7775-4a2b-bec3O9641bb3df61\&groupld=43805

Kell, L. T., Pastoors, M. A., Scott, R. D., Smith, M. T., Van Beek, F. A., O’Brien, C. M., and Pilling, G. M. 2005. Evaluation of multiple management objectives for Northeast Atlantic flatfish stocks: sustainability vs. stability of yield. ICES Journal of Marine Science 62:1104-1117.

Mace, P.M. and M.P. Sissenwine. 1993. How much spawning per recruit is enough? pp 101-118 in S.J. Smith, J.J. Hunt and D.Revered (eds.) Risk Evaluation and Biological Reference Points for Fisheries Management. Canadian Special Publication of Fisheries and Aquatic Sciences 120. National Research Council of Canada.

Myers R. A., K. G. Bowen and N. J. Barrowman. 1999. Maximum reproductive rate of fish at low population sizes. Canadian Journal of Fisheries and Aquatic Sciences 56:2404-2419

O'Farrell, M.R. and L.W. Botsford. 2006. Estimation of change in lifetime egg production from length frequency data. Canadian Journal of Fisheries and Aquatic Sciences 62:1626-1639.

Phipps, K.E., R. Fujita, and T. Barnes. 2010. From paper to practice: incorporating new data and stock assessment methods into California fishery management. Managing Data-Poor Fisheries: Case Studies, Models \& Solutions 1:159-184.

Roughgarden, J. and F. Smith. 1996. Why fisheries collapse and what to do about it. Proceedings of the National Academies of Sciences (USA) 93:5078-5083

## Appendix 2

Abraham, E., J. Pierre, et al. 2009. Effectiveness of fish waste management strategies in reducing seabird attendance at a trawl vessel Fisheries Research 95 (3): 210-219

Ancha, L. 2008. Mediterranean and Black Seas Regional Assessment. In: Division of Marine Science and Conservation, Nicholas School of the Environment. Duke University Beaufort, NC, p. 58.

Anderson, O.R.J., C.J. Small, J.P. Croxall, E.K. Dunn, B.J. Sullivan O. Yates, et al. 2011. Global seabird bycatch in longline fisheries. Endangered Species Research 14:91-106.

Archer, F.I., J.V.Redfern, T. Gerrodette, S.J. Chiversand W.F. Perrin, W.F. 2010. Estimation of relative exposure of dolphins to fishery activity. Marine Ecology Progress Series 410:245-255.

Aylesworth, L. 2009. Oceania Regional Assessment - Pacific Island Fisheries and Interactions with Marine Mammals, Seabirds and Sea Turtles. In: Division of Marine Science and Conservation, Nicholas School of the Environment. Duke University Beaufort, NC, p. 432.

Alverson, D.L., M. H. Freeberg, J. G. Pope, and S. A. Murawski. 1994.A global assessment of fisheries bycatch and discards. FAO Fisheries Technical Paper. No. 339. Rome, FAO. 233p.

Baird, S.J., E. Bradford. 2000. Factors that may have influenced the capture of NZ fur seals (Arctocephalus forsteri) in the west coast South Island hoki fishery, 1991-1998. NIWA Technical Report 92. 35 p.

Baum, J.K. and R. A. Myers. 2004. Shifting baselines and the decline of pelagic sharks in the Gulf of Mexico. Ecology Letters 7:135-145.

Baum, J.K., R. A. Myers, D. G. Kehler, B. Worm, S. J. Harley, and P. A. Doherty. 2003. Collapse and conservation of shark populations in the Northwest Atlantic. Science 299:389-392.

Bensley, N., J. Woodhams, H.M. Patterson, M. Rodgers, K. McLoughlin, I.C. Stobutzki, and Begg, G.A.. 2010. 2009 Shark assessment report for the Australian National Plan of Action for the conservation and management of sharks. Bureau of Rural Sciences, Government of Australia.

Birdlife International. 2011. http://datazone.birdlife.org/sowb/spotseabirds

BirdLife International 2010. Marine Important Bird Areas - Priority for the Conservation of Biodiversity. Cambridge, UK: BirdLife International. ISBN 978-0-946888-74-0.

Bjorkland, R. 2009. The Caribbean Ocean. In: Project GloBAL Regional Assessment of Marine Megafauna. Duke University Beaufort, NC.

Bjorkland, R. and R. Bjorkland. 2013a. A review of bycatch of large marine vertebrates in global fisheries, section 1: Seabirds. In: Monterey Bay Aquarium Seafood Watch Monterey, CA.

Bjorkland, R. and R. Bjorkland. 2013b. A review of bycatch of large marine vertebrates in global fisheries, section 2: Marine mammals. In: Monterey Bay Aquarium Seafood Watch Monterey, CA.

Bjorkland, R. and R. Bjorkland. 2013c. A review of bycatch of large marine vertebrates in global fisheries, section 3: Sea Turtles. In: Monterey Bay Aquarium Seafood Watch Monterey, CA.

Bjorkland, R. and R. Bjorkland. 2013d. A review of bycatch of large marine vertebrates in global fisheries, section 4: Sharks. In: Monterey Bay Aquarium Seafood Watch Monterey, CA.

Brothers, N., A.R. Duckworth, C. Safina, and E.L. Gilman. 2010. Seabird bycatch in pelagic longline fisheries is grossly underestimated when using only haul data. PLOS One, 5, e12491. doi:
12410.11371/journal.pone.0012491.

Camhi, M.D., S. V. Valenti, S. V. Fordham, S. L. Fowler, and C. Gibson. 2009. The Conservation Status of Pelagic Sharks and Rays: Report of the IUCN Shark Specialist Group Pelagic Shark Red List Workshop. IUCN Species Survival Commission Shark Specialist Group. Newbury, UK. x + 78p.

Carboneras Malet, C. 2009. Draft Guidelines for reducing by catch of seabirds in the Mediterranean region. In: Mediterranean Action Plan, United Nations Environment Programme - UNEPDEPI/MED WG331/12 Floriana, Malta.

Chuenpagdee, R., L. E. Morgan, et al. 2003. Shifting gears: Assessing collateral impacts of fishing methods in US waters. Frontiers in Ecology and Environment 1(10): 517-524.

Chapple, T.K., S. J. Jorgensen, S. D. Anderson, P. E. Kanive, A. P. Klimley, L. W. Botsford, and B. A. Block. 2011. A first estimate of white shark, Carcharadon carcharias, abundance off Central California. Biological Letters 00:1-3.

Cramer, K.L., W.L. Perryman, and T. Gerrodette. 2008. Declines in reproductive output in two dolphin populations depleted by the yellowfin tuna purse seine fishery. Marine Ecology Progress Series 369, 273285.

Curry, B.E. 1999. Stress in mammals: the potential influence of fishery-induced stress on dolphins in the eastern tropical Pacific Ocean. In: NOAA Technical Memorandum NMFS-SWFSC-2760. NOAA La Jolla, CA.

Dawson, S.M. 1991. Incidental catch of Hector's dolphin in inshore gillnets. Marine Mammal Science 7:286-295.

Dufresne, S.P., A. Grant, W. S. Norden, and J. Pierre. 2007. Factors affecting cetacean bycatch in a New Zealand trawl fishery. DOC Research and Development series 282. 18pp.

Dulvy, N.K., J. K. Baum, S. Clarke, L. J. V. Compagno, E. Corte's, A. Domingo, et al. 2008. You can swim but you can't hide: the global status and conservation of oceanic pelagic sharks and rays. Aquatic Conservation 18: 459-482.

Ferretti, F., Myers, R.A., Serena, F. \& Lotze, H.K. 2008. Loss of large predatory sharks from the Mediterranean Sea. Conservation Biology 22:952-964.

Fertl, D. and S. Leatherwood. 1997. Cetacean interactions with trawls: a preliminary review. Journal of Northwest Atlantic Fishery Science 22:219-248.

Filippi, D., S. Waugh, and S. Nicol. 2010. Revised spatial risk indicators for seabird interactions with longline fisheries in the Western and Central Pacific. In: West and Central Pacific Fisheries Commission, WCPFC-SC6-2010/EB-IP 01 Nukualofa, Tonga.

Francis, M.P. and M.H. Smith. 2010. Basking shark Cetorhinus maximus bycatch in New Zealand fisheries, 1994-95 to 2007-08. In: New Zealand Aquatic Environment and Biodiversity Report No. 49. Ministry of Fisheries Wellington, New Zealand.

Francis, M.P. and P. Sutton. 2012. Possible factors affecting bycatch of basking sharks Cetorhinus maxiumus in New Zealand trawl fisheries. In: Department of Conservation, Contract No. 4346 Wellington, New Zealand.

Fuller, S.D., C Picco, J. Ford, C. Tsao, L. E. Morgan, D. Hangaard, and R. Chuenpagdee. 2008. How we fish matters: Addressing the ecologicalimpacts of Canadian fishing gear. Ecology Action Centre, Living Oceans Society, and Marine Conservation Biology Institute. 25pp.

Gallagher, A.J., E.S. Orbesen, N. Hammerschlag, and J.E. Serafy. 2014. Vulnerability of oceanic sharks as pelagic longline bycatch. Global Ecology and Conservation 1:50-59.

Gillman, E., S. Clarke, N. Brothers, J. Alfaro-Shigueto, J. Mandelman, J.C. Mangel. et al. 2008. Shark interactions in pelagic longline fisheries Marine Policy 32:1-18.

Graham, K.J., N. L. Andrew, and K. E. Hodgson. 2001. Changes in relative abundance of sharks and rays on Australian south east fishery trawl grounds after twenty years of fishing. Marine and Freshwater Research 52:549-61.

Hall, M. and M. Roman. 2013. Bycatch and non-tuna catch in the tropical tuna purse seine fisheries of the world. In: FAO Fisheries and Aquaculture Technical Paper 568 Rome.

Icelandic Ministry of Fisheries and Agriculture Fishing Gear-Danish Seine. Accessed on January 20, 2011.
James, K., C., R.L. Lewison, P.W. Dillingham, K. Alexandra Curtis, and J.E. Moore. 2015. Drivers of retention and discards of elasmobranch non-target catch. Environmental Conservation, doi: 10.1017/S0376892915000168.

Karpouzi, V., R. Watson, and D. Pauly. 2007. Modelling and mapping resource overlap between seabirds and fisheries on a global scale: a preliminary assessment. Marine Ecology Progress Series 343:97-99.

Kelez, S. 2009. The Eastern Tropical Pacific. In: Project GloBAL Regional Assessment of Marine Megafauna Bycatch. Duke University Beaufort, NC.

Kiszka, J. and R. van der Elst. 2015. Elasmobranchs sharks and rays: A review of teh status, distribution and interaction fiwht fisheries in the Southwest Indian Ocean. In: Offshore Fisheries of the Southwest Indian Ocean: their status and the impact on vulnerable species ed. van der Elst, R. South African Association for Marine Biological Research, Oceanographic Research Institute Durban, South Africa, pp. 365-389.

Lawson, T. 2011. Estimation of catch rates and catches of key shark species in tuna fisheries of the western and central Pacific Ocean using observer data. In: Western and Central Pacific Fisheries Commisson, WCPFC-SC7-2011/EB-IP-02. available at https://meetings.wcpfc.int/node/7337, Federated States of Micronesia.

Lewison, R.L., L.B. Crowder, B.P. Wallace, J.E. Moore, T. Cox, R. Zydelis, et al. 2014. Global patterns of marine mammal, seabird, and sea turtle bycatch reveal taxa-specific and cumulative megafauna hotspots. Proceedings of the National Academy of Sciences 111:5271-5276.

Mapping the World's Sea Turtles. 2011.
http://www.gearthblog.com/blog/archives/2011/07/mapping_the_worlds_sea_turtles.html?utm_sourc e=feedburner\&utm medium=feed\&utm campaign=Feed\%3A+GoogleEarthBlog $+\%$ 28Google + Earth + Blo g\%29

McCarthy, P., S.L. McDonald, and L. Lester. 2009. The Southern Ocean. In: Project GloBAL Regional Assessment of Marine Megafauna Bycatch. Duke University Beaufort, NC.

McCarthy, P. and R. Zydelis. 2009. Southeast Asia. In: Project GloBAL Regional Assessment of Marine Megafauna Bycatch. Duke University Beaufort, NC.

McDonald, S.L., K.C. James, R.L. Lewison, and L.B. Crowder. 2013. Demographic, geographic, and behavioral correlates of fisheries bycatch in marine mammals worldwide Duke University, Nicholas School of the Environment, pp. 1-27.

Moore, J.E. 2009. West Africa. In: Project GloBAL Regional Assessment of Marine Megafauna Bycatch. Duke University Beaufort, NC.

Moore, J.E., W.P Wallace, R.L. Lewison, R. Žydelis, T.M. Cox, and L.B. Crowder, L.B. 2009. A review of marine mammal, sea turtle and seabird bycatch in USA fisheries and the role of policy in shaping management. Marine Policy 33:435-451.

Morizur, Y., S.D. Berrow, N. J. C. TregenzaA. S. Couperus, and S. Pouvreau. 1999. Incidental catches of marine mammals in pelagic trawl fisheries of the northeast Atlantic. Fisheries Research 41:297-307.

Myers, R.A., J. K. Baum, T. Shepherd, S. P. Powers, and C. H. Peterson. 2007. Cascading effects of the loss of apex predatory sharks from a coastal ocean. Science 315:1846-1850.

Myers, R.A. \& Worm, B. 2005. Extinction, survival or recovery of large predatory fishes. Philosophical Transactions of the Royal Society of London B 360:13-20.

Myrick, A.C. and P.C. Perkins. 1995. Adrenocortical color darkness and correlates as indicators of continuous acute premortem stress in chased and purse-seine captured male dolphins. Pathophysiology 2:191-204.

Oliver, S., M. Braccini, S.J. Newman, and E.S. Harvey. 2015. Global patterns in the bycatch of sharks and rays. Marine Policy 54:86-97

Paleczny, M., E. Hammill, V. Karpouzi, and D. Pauly. 2015. Population trend of the world's monitored seabirds, 1950-2010. PLOS One, DOI: 10.1371/journal.pone.0129342.

Palsson, O.K. 2003. A length-based analysis of haddock discards in Icelandic Fisheries. Fisheries Research 59: 437-446.

Pompa, S., P. P. Ehrlich, G. Ceballos. 2011. Global distribution and conservation of marine mammals. Proceedings of the National Academies of Science (USA).
https://doi.org/10.1073/pnas. 1101525108.

ProjectGloBAL 2009a. The Southwest Atlantic Ocean. In: Project GloBAL Regional Assessment of Marine Megafauna Bycatch. Duke University Beaufort, NC.

ProjectGloBAL 2009b. The West Indian Ocean. In: Project GloBAL Regional Assessment of Marine Megafauna Bycatch. Duke University Beaufort, NC.

ProjectGloBAL 2009c. The Western Indian Ocean. In: Project GloBAL Regional Assessment of Marine Megafauna Bycatch. Duke University Beaufort, NC.

Read, A.J., P. Drinker, and S. Northridge. 2006. Bycatch of marine mammals in U.S. and global fisheries. Conservation Biology 20:163-169.

Reeves, R.R., P. Berggren, E.A. Crespo, N. Gales, S.P. Northridge, G. Notarbartolo di Sciara, et al. 2005. Global priorities for reduction of cetacean bycatch. Report to the World Wide Fund for Nature. [Available at http://assets.panda.org/downloads/topninereportenglish.pdf], p. 29 pp.

Reeves, R.R., K. McClellan, and T.B. Werner. 2013. Marine mammal bycatch in gillnet and other entangling net fisheries, 1990 to 2011. Endangered Species Research 20:71-97.

Rowe, S. J. 2007. A review of methodologies for mitigating incidental catch of protected marine mammals. DOC Research \& Development Series 283. 48pp.

Slooten, E. and S.M. Dawson. 2010. Assessing the effectiveness of conservation management decisions: likely effects of new protection measures for Hector's dolphin Cephalorhynchus hectori. Aquatic Conservation: Marine and Freshwater Ecosystems 30:334-347.

Smith, Z., M. Gilroy, M. Eisenson, E. Schnettler, and S. Stefanski, S. 2014. Net loss: The killing of marine mammals in foreign fisheries. In: NRDC Report R:13-11-B Washington, D.C.

South Australian Sardine Industry Association 2012. Code of Practice for mitigation of interactions of the South Australian Sardine Fishery with threatened, endangered, and protected Species. In: ABN 81758 910055 ed. South Australian Sardine Industry Association, Australia.

Stewart, K.R., R.L. Lewison, D.C. Dunn, R.H. Bjorkland, S. Kelez, P.N. Halpin, et al. 2010. Characterizing fishing effort and spatial extent of coastal fisheries. PLOS One : e14451.
doi:14410.11371/journal.pone.0014451.

Suazo, C.G., L.A. Cabezas, C.A. Moreno, J.A. Arata, G. Luna-Jorquera, A. Simeone, et al. 2014. Seabird bycatch in Chile: A synthesis of its impacts, and a review of strategies to contribute to the reduction of a global phenomenon. Pacific Seabirds 41:1-12.

Sullivan, B.J., P. Brickle, T.A. Reid, D.G. Bone and D.A.J. Middleton. 2006. Mitigation of seabird mortality on factory trawlers: trials of three devices to reduce warp cable strikes, Polar Biology 29:745-753.

SWOT (State of the World's Sea Turtles). 2011. http://seamap.env.duke.edu/swot
van der Elst, R. and B. Everett. 2015. Offshore fisheries of the Southwest Indian Ocean: their status and the impact on vunerable species. In: Oceanographic Research Institute, Special Publication 10. South African Association of Marine Biological Research.

Vögler, R., A.C. Milessi, and R.A. Quiñones. 2008. Influence of environmental variables on the distribution of Squatina guggenheim Chondrichthyes, Squatinidae in the Argentine-Uruguayan Common Fishing Zone. Fisheries Research 91:212-221.

Wallace, B.P., R. L. Lewison, S. L. McDonald et al.. 2010. Global patterns of marine turtle bycatch. Conservation Letters 3:1-21.

Wallace, B.P., C. Y. Kot, A. D. DiMatteo, T. Lee, L. B. Crowderand R. L. Lewison. 2013. Impacts of fisheries bycatch on marine turtle populations worldwide: toward conservation and research priorities. Ecosphere 4:1-49.

Watling, L. 2005. The global destruction of bottom habitats by mobile fishing gear. In: Marine Conservation Biology, The science of maintaining the sea's biodiversity. Ed by E. Norse and L Crowder. pp 198-210. Island Press.

Wanless, R.M. 2015. Seabirds. In: Offshore Fisheries of the Southwest Indian Ocean: their status and the impact on vulnerable species eds. van der Elst, R and Everett, B. South African Association for Marine Biological Research, Oceanographic Research Institute Durban, South Africa.

Ward, P. and R. A. Myers. 2005. Shift in open-ocean fish communities coinciding with the commencement of commercial fishing. Ecology 86:835-847.

Watkins, B.P., S.L. Petersen, and P.G. Ryan. 2008. Interactions between seabirds and deep-water hake trawl gear: an assessment of impacts in South African waters. Animal Conservation 11:247-254.

Wiedenfeld, D. 2015. Seabird Bycatch and Sustainable Fisheries. In: Solutions for Seafood ed. Webinar.

Weimerskirch, H, D. Capdeville, and G. Duhamel. 2000. Factors affecting the number and mortality of seabirds attending trawlers and longliners in the Kerguelen area, Polar Biology 23:236-249.

Wieneke, J. and G. Robertson. 2002. Seabird and seal—fisheries interactions in the Australian Patagonian toothfish, Dissostichus eleginoides trawl fishery. Fisheries Research 54:253-265

Žydelis, R., C. Small, and G. French. 2013. The incidental catch of seabirds in gillnet fisheries: A global review. Biological Conservation 162:76-88.

## Appendix 3

Australian Department of Agriculture, Fisheries and Forestry 2007. Commonwealth Fisheries Harvest Strategy: Policy and Guidelines. December 2007. General link to Commonwealth Fisheries Harvest Strategy available at: https://www.agriculture.gov.au/fisheries/domestic/harvest strategy_policy.

Cope, J. 2011. Personal Communication. National Marine Fisheries Service, Northwest Fisheries Science Center.

Cope, J., J. DeVore, E. J. Dick, K. Ames, J. Budrick, D.L. Erickson, J. Grebel, G. Hanshew, R. Jones, L. Mattes, C. Niles, and S. Williams. 2011. An approach to defining stock complexes for U.S. west coast groundfishes using vulnerabilities and ecological distribution. North Atlantic Journal of Fisheries Management 31:589-604.

Cope, J. M. and A. E. Punt. 2009. Length-based reference points for data-limited situations: Applications and restrictions. Marine and Coastal Fisheries: Dynamics, Management, and Ecosystem Science 1:169186.

Dick, E.J and A.D. MacCall. 2010. Estimates of sustainable yield for 50 data-poor stocks in the Pacific coast groundfish fishery management plan. NOAA-TM-NMFS-SWFSC-460. Available at: http:///swfsc.noaaa.gov/publications/TM/SWFSC/NOAA-TM-NMFS-SWFSC-460.pdf

Dowling, N.A., D. C. Smith, I. Knuckey et al. 2008. Developing harvest strategies for low-value and datapoor fisheries: Case studies from three Australian fisheries. Fisheries Research 94:380-390

Field, J., J. Cope and M. Key. 2010. A descriptive example of applying vulnerability evaluation criteria to California nearshore species. Managing data-poor fisheries: Case Studies, Models and Solutions 1:235246

Fujita, R. M., K. T. Honey, A. Morris, H. Russell and J. Wilson. 2010. Cooperative strategies in fisheries management: Transgressing the myth and delusion of appropriate scale. Bulletin of Marine Science 86:251-271.

Honey, K.T., J. H. Moxeley, and R. M. Fujita. 2010. From rages to fishes. Managing Data-Poor Fisheries: Case Studies, Models \& Solutions. California Sea Grant College Program (1):159-184. ISBN number 978-1-888691-23-8

Johannes, R. E. 1998. The case for data-less marine resource management: Examples from tropical nearshore fisheries. Trends in Ecology and Evolution 13: 243-246.

MacCall, A. D. 2009. Depletion-corrected average catch: a simple formula for estimating sustainable yields in data-poor situations. ICES Journal of Marine Science 66: 2267-2271.

Morrison, W.E., and Termini V. 2016. A Review of Potential Approaches for Managing Marine Fisheries in a Changing Climate. U.S. Dept. of Commerce, NOAA. NOAA Technical Memorandum NMFS-OSF-6, 35 p.

NMFS. 2011. Assessment Methods for Data-Poor Stocks Report of the Review Panel Meeting National Marine Fisheries Service (NMFS) Southwest Fisheries Science Center (SWFSC) Santa Cruz, California April 25-29, 2011.

Patrick, W. S., P. Spencer, O. Ormseth, J. Cope, J. Field, D. Kobayashi, T. Gedamke, E. Cortés, K. Bigelow, W. Overholtz, J. Link, and P. Lawson. 2009. Use of productivity and susceptibility indices to determine stock vulnerability, with example applications to six U.S. fisheries. U.S. Department of Commerce, NOAA Tech. Memo. NMFS-F/SPO-101, Seattle, WA

Patrick, W.S., P. Spencer, J. Link, et al . 2010. Using productivity and susceptibility indices to assess the vulnerability of United States fish stocks to overfishing. Fishery Bulletin 108:305-322.

Prince, J. D. 2010. Managing data poor fisheries: Solutions from around the world. Managing data-poor fisheries: Case studies, models and solutions. 1:3-20

Restrepo, V.R. and J. E. Powers. 1998. Precautionary control rules in US fisheries management: specification and performance. ICES Journal of Marine Science 56: 846-852.

Restrepo, V.R., G.G. Thompson, P.M. Mace, W.L. Gabriel, L.L. Low, A.D. MacCall, R.D. Methot, J.E. Powers, B.L. Taylor, P.R. Wade and J.F. Witzig. 1998. Technical Guidance on the Use of Precautionary Approaches to Implementing National Standard 1 of the Magnuson-Stevens Fishery Conservation and Management Act. NOAA Technical Memorandum NMFS-F/SPO-137. 54pps.

Smith, D. et al. 2009. Reconciling approaches to the assessment and management of data-poor species and fisheries with Australia's Harvest Strategy Policy. Marine and Coastal Fisheries: Dynamics, Management, and Ecosystem Science 1:244-254

Wilson, J. R., J. D. Prince and H. S. Lenihan. 2010. Setting harvest guidelines for sedentary nearshore species using marine protected areas as a reference. Marine and Coastal Fisheries: Dynamics, Management, and Ecosystem Science 2:14-27.

## Appendix 4

Belcher, C. N. and C. A. Jennings 2010. Identification and evaluation of shark bycatch in Georgia's commercial shrimp trawl fishery with implications for management. Fisheries Management and Ecology 18:104-112.

Butler, J. and G. Heinrich 2005. The Effectiveness of Bycatch Reduction Devices on Crab Pots at Reducing Capture and Mortality of Diamondback Terrapins and Enhancing Capture of Blue Crabs. NOAA Project Final Report, University of North Florida: 9.

Cox, T., R. L. Lewison, et al. 2007. Comparing effectiveness of experimental and implemented bycatch reduction measures: the ideal and the real. Conservation Biology 21:1155-1164.

Eayers, S. 2007. A Guide to Bycatch Reduction in Tropical Shrimp-Trawl Fisheries, Revised Edition. Rome, FAO: 124.

FAO 2009. Guidelines to Reduce Sea Turtle Mortality in Fishing Operations. Rome, FAO: 139.

Finkelstein, M., V. Bakker, et al. 2008. Evaluating the potential effectiveness of compensatory mitigation strategies for marine bycatch. PLoS ONE 3:e2480.

Gillet, R. 2008. Global Study of Shrimp Fisheries. Rome, FAO.

Gilman, E. and C. Lundin 2008. Minimizing Bycatch of Sensitive Species Groups in Marine Capture Fisheries: Lessons from Tuna Fisheries, IUCN.

Hall, M. 1998. An ecological view of the tuna-dolphin problem: Impacts and tradeoffs. Reviews of Fish Biology and Fisheries 8:1-34.

Løkkeborg, S. 2008. Review and assessment of mitigation measures to reduce incidental catch of seabirds in longline, trawl and gillnet fisheries. Rome, FAO: 33.

SBWG 2010. Report of the Third Meeting of the Seabird Bycatch Working Group. Mar del Plata, Argentina, FAO ACAP.

Wiedenfeld, D. 2015. Seabird Bycatch and Sustainable Fisheries. In: Solutions for Seafood (ed. Webinar).
Žydelis, R., B. P. Wallace, et al. 2009. Conservation of Marine Megafauna through Minimization of Fisheries Bycatch. Conservation Biology 23:608-616.

## Appendix 5

Chuenpagdee, R., L. E. Morgan, et al. 2003. Shifting gears: Assessing collateral impacts of fishing methods in US waters. Frontiers in Ecology and Environment 1:517-524.

Fuller, S.D., C Picco, et al. 2008. How we fish matters: Addressing the ecological impacts of Canadian fishing gear. Ecology Action Centre, Living Oceans Society, and Marine Conservation Biology Institute. 25pp.

Gillet, R. 2008. Global Study of Shrimp Fisheries. Rome, FAO.
Hall-Spencer, J.M., and P.G. Moore, 2000. Scallop dredging has profound, long-term impacts on maerl habitats. ICES Journal of Marine Science 57:1407-1415

Icelandic Ministry of Fisheries and Agriculture Fishing Gear-Danish Seine. Accessed on January 20, 2011. https://www.government.is/topics/business-and-industry/fisheries-in-iceland/.

Jones, J.B. 1992. Environmental impact of trawling on the seabed: a review. New Zealand Journal of Marine and Freshwater Research 26:59-67

Kaiser, M.J., J. S.Collie, S. J Hall, S. Jennings, I. R. Poiner. 2001. Impacts of fishing gear on marine benthic habitats. Reykjavik Conference on Responsible Fisheries in the Marine Ecosystem. Reykjavik, Iceland. ftp://ftp.fao.org/fi/document/reykjavik/pdf/12kaiser.PDF

Neckles, H., F.T. Short, S. Barker , and B.S. Kopp. 2005. Disturbance of eelgrass Zostera marina by commercial mussel Mytilus edulis harvesting in Maine: dragging impacts and habitat recovery. Marine Ecology Progress Series 285:57-73

NEFMC. 2010. Essential fish habitat (EFH) omnibus amendment. The swept area seabed impact (SASI) model: A tool for analyzing the effects of fishing on essential fish habitat. Part 1: literature review and vulnerability assessment. Newburyport, MA. 160 pp. chrome-
extension://efaidnbmnnnibpcajpcg|clefindmkaj/https://archive.nefmc.org/habitat/cte mtg docs/10061 0/100604_SASI Part_1_VA_FINAL.pdf

PFMC. 2005. Pacific coast groundfish fishery management plan for the Califonia, Oregon and Washington groundfish fishery. Appendix C part 2. The effects of fishing on habitat: west coast perspective. PFMC Portland, OR. 48pp.

Rose, C., A. Carr, D. Ferro, R. Fonteyne, P. MacMullen. 2000. Using gear technology to understand and reduce unintended effects of fishing on the seabed and associated communities: background and potential directions. ICES Working Group on Fishing Haarlem, The Netherlands. 19pp.

Sainsbury, K.J., R. A. Campbell, R. Lindholm, and A. W. Whitlaw. 1998. Experimental management of an Australian multispecies fishery: examining the possibility of trawl-induced habitat modification. Global Trends: Fisheries Management (eds E. K.Pikitch, D. D.Huppert \& M. P. Sissenwine), pp. 107-112. American Fisheries Society, Bethesda, Maryland.

Thrush, S.F., J.E. Hewitt, et al. 1998. Disturbance of the marine benthic habitat by commercial fishing: impacts at the scale of the fishery. Ecological Applications 8:866-879

## Appendix 6

Auster, P. 2001. Defining Thresholds for Precautionary Habitat Management Actions in a Fisheries Context. North American Journal of Fisheries Management 21:1-9

He, P. 2007. Technical measures to Reduce Seabed Impact of Mobile Fishing Gears, pp 141-179 in S. Kennelly (ed). Bycatch Reduction in the World's Fisheries.

Fujioka, J.T. 2006. A model for evaluating fishing impacts on habitat and comparing fishing closure strategies. Canadian Journal of Fisheries and Aquatic Sciences 63:2330-2342.

Lindholm, J. B., P. J. Auster, M. Ruth, and L. S. Kaufman. 2001. Modeling the effects of fishing, and implications for the design of marine protected areas: juvenile fish responses to variations in seafloor habitat. Conservation Biology 15:424-437.

Valdemarsen, J., T. Jorgensen, et al.2007. Options to mitigate bottom habitat impact of dragged gears. FAO Fisheries Technical Paper. 506.

## Appendix 7

Babcock, E.A. and A. D. MacCall, A.D. 2011. How useful is the ratio of fish density outside versus inside no-take marine reserves as a metric for fishery management control rules? Canadian Journal of Fisheries and Aquatic Sciences 68:343-359.

Carruthers, T.R., A.E. Punt, C.J. Walters, A. MacCall, M.K. McAllister, E.J. Dick, and J. Cope. 2014. Evaluating methods for setting catch limits in data-poor fisheries. Fisheries Research 153: 48-68.

Chrysafi A. and A. Kuparinen. 2016. Assessing abundance of populations with limited data: lessons learned from data-poor fisheries stock assessment. Environmental Reviews 24: 25-38.

Cope, J.M and A.E. Punt. 2009. Length-based reference points for data-limited situations: applications and restrictions. Marine and Coastal Fisheries: Dynamics, Management, and Ecosystem Science 1:169186.

Data-Limited Methods Toolkit, Fishery Library. 2018. Fishery Library available at: https://www.datalimitedtoolkit.org/fishery library/.

Dowling N. A., Smith A. D. M., Smith D. C., Parma A. M., Dichmont C. M., Sainsbury K., Wilson J. R, Dougherty, D.T., Cope J.M. 2019. Generic solutions for data-limited fishery assessments are not so simple. Fish and Fisheries 20: 174-188.

Environmental Defense Fund (EDF). 2016. FISHE - Framework for Integrated Stock and Habitat Evaluation. Available at: http://fishe.edf.org/.

Summary Document:
Data Limited Analysis Methods: Inputs, outputs and major assumptions:
http://fishe.edf.org/node/40

Primers:
Primer for Length-Based Assessment Methods: http://fishe.edf.org/node/69
Primer for Froese Length-Based Sustainability Indicators: http://fishe.edf.org/node/87

Primer for Cope and Punt Length-Based Reference Point Method: http://fishe.edf.org/node/66 Fair Trade USA Capture Fisheries Standard Version 1.1.0 Available at: https://www.fairtradecertified.org/sites/default/files/filemanager/documents/CFS/FTUSA STD_CFSEN 1.1.0.pdf

Food and Agriculture Organization of the United Nations (FAO). 2003. Fish Stock Assessment Manual. FAO Fisheries Technical Paper, No. 393. Retrieved from http://www.fao.org/docrep/006/x8498e/x8498e00.HTM

Froese, R. 2004. Keep it simple: three indicators to deal with overfishing. Fish and Fisheries 5:86-91

Froese, R., A. Stern-Pirlot, H. Winker, and D. Gascuel. 2008. Size matters: How single-species management can contribute to ecosystem-based fisheries management. Fisheries Research 92:231-241.

Newman, D., J. Berkson, and L. Suatoni. 2015. Current methods for setting catch limits for data-limited fish stocks in the United States. Fisheries Research 164: 86-93.

Newman D, T. Carruthers, A. MacCall, C. Porch and L. Suatoni. 2014. Improving the science and management of data-limited fisheries: an evaluation of current methods and recommended approaches. Natural Resources Defense Council, NRDC report R, 14-09.

Pons M., L.T. Kell, M.B. Rudd, J.M. Cope and F. Lucena Frédou. 2019. Performance of length-based datalimited methods in a multifleet context: application to small tunas, mackerels, and bonitos in the Atlantic Ocean. ICES Journal of Marine Science 76: 960.

## Appendix 8

Berger, A. 2019. Character of temporal variability in stock productivity influences the utility of dynamic reference points. Fisheries Research 217: 185-197.

Cury, P., A. Bakun, R.J.M. Crawford, A. Jarre, R.A. Quinones, L.J. Shannon and H.M. Verheye. 2000. Small pelagics in upwelling systems: patterns of interaction and structural changes in 'wasp-waist' ecosystems. ICES Journal of Marine Science 57, 603-618.

Essington, T.E., P. E. Moriarty; H. E. Froehlich, E. E. Hodgson, L. E. Koehn, K. L. Oken, M. C. Siple, and C. C. Stawitz. 2015. Fishing amplifies forage fish population collapses. Proceedings of the National Academy of Sciences of the USA 112 (21): 6648-6652.

King, J.R., G.A. McFarlane and A.E. Punt, 2015. Shifts in fisheries management: adapting to regime shifts. Philosophical Transactions of the Royal Society B: Biological Sciences 370: 20130277.

Koehn, L.E., T.E. Essington, K.N. Marshall, I.C. Kaplan, W.J. Sydeman, A.I. Szoboszlai and J.A. Thayer. 2016. Developing a high taxonomic resolution food web model to assess the functional role of forage fish in the California Current ecosystem. Ecological Modelling 335:87-100.

Marine Stewardship Council (MSC). 2018. MSC Fisheries Standard v2.01 and MSC Guidance to the Fisheries Standard v2.01. Marine Stewardship Council, London, UK. 156 p. Available at: https://www.msc.org/docs/default-source/default-document-library/for-business/program-documents/fisheries-program-documents/msc-fisheries-standard-v2-01.pdf?sfvrsn=8ecb3272_11.

Marshall, K. N., L.E. Koehn, P.S. Levin, T.E. Essington, and O.P. Jensen. 2019. Inclusion of ecosystem information in US fish stock assessments suggests progress toward ecosystem-based fisheries management. ICES Journal of Marine Science 76: 1-9.

Pikitch, E., P.D. Boersma, I.L. Boyd, D.O. Conover, P. Cury, T. Essington, S.S. Heppell, E.D. Houde, M. Mangel, D. Pauly, É. Plagányi, K. Sainsbury, R.S. and Steneck. 2012. Little Fish, Big Impact: Managing a Crucial Link in Ocean Food Webs. Lenfest Ocean Program. Washington, DC. 108 pp.

Plagányi,E.E and T.E. Essington. 2014. When the SURFs up, forage fish are key. Fisheries Research 159:68-74.

Plaganyi, E., M. Haywood, R. Gorton, M. Siple, Margaret and R. Deng. 2019. Management implications of modelling fisheries recruitment. Fisheries Research 217: 169-184.

Skern-Mauritzen, M., G. Ottersen, N.O. Handegard, G. Huse, G.E. Dingsor, N.C. Stenseth and O.S. Kjesbu 2015. Ecosystem processes are rarely included in tactical fisheries management. Fish and Fisheries 17: 165-175.

Szuwalski, C.S. 2013. Production is a poor metric for identifying regime-like behaviour in marine stocks. Proceedings of the National Academy of Sciences 110 (16):E1436.

Vert-Pre, K.A., R.O. Amoroso, O.P. Jensen and R. Hilborn. 2013. Frequency and intensity of productivity regime shifts in marine fish stocks. Proceedings of the National Academy of Sciences 110 (5): 1779-1784.

Siple, M.C., T.E. Essington, and É. Plagányi. 2019. Forage fish fisheries management requires a tailored approach to balance trade-offs. Fish and Fisheries, 20(1): 110-124.


[^0]:    1 "Affected" stocks include all stocks affected by the fishery, no matter whether target or bycatch, or whether they are ultimately retained or discarded.

[^1]:    ${ }^{2}$ Guidance on "appropriate data-limited assessment methods" to be developed. Until that time, Appendix 7 is provided to illustrate examples only and these indicators and thresholds should not be assumed to be appropriate for all fisheries. Data and assessments to be provided by the fisheries and verified by expert input. See also "recent stock assessment" definition for guidance on consideration of the currency of data-limited assessments.

[^2]:    ${ }^{3}$ In this situation, the limit reference point should be set at the overfishing limit (or equivalent). If target reference point is set at the overfishing limit and is being exceeded, a score of high concern should be given.
    ${ }^{4}$ Where fishing mortality (F) is unknown, or where $F$ is known but there are no available reference points to determine whether $F$ is at an appropriate level.

[^3]:    ${ }^{5}$ When determining an appropriate timeframe it is important to consider the ability of managers to adjust management measures to take into account the latest scientific information and advice, for example, if a stock assessment identifies that overfishing took place in the previous fishing season, do managers adjust the harvest controls for the upcoming season?

[^4]:    ${ }^{6}$ Connectance is the number of links in a food web involving a species, scaled to the total number of links (i.e., complexity) in the food web. A calculated connectance value can be challenging to interpret because it is highly dependent on how a study chose to aggregate species. For reference see: Plagányi, E.E and T.E. Essington. 2014. When the SURFs up, forage fish are key. Fisheries Research 159:68-74.
    ${ }^{7}$ In this context, productivity encompasses both the intrinsic rate of increase ( $r$ ) and the carrying capacity (K).

[^5]:    ${ }^{8}$ Steepness is a key parameter of the Beverton-Holt spawner-recruit model that is defined as the proportion of unfished recruitment produced by $20 \%$ of the unfished spawning biomass. Steepness is difficult to estimate, and the calculation of reference points is often very sensitive to estimates of steepness.

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[^6]:    ${ }^{9}$ Connectance is the number of links in a food web involving a species, scaled to the total number of links (i.e., complexity) in the food web. A calculated connectance value can be challenging to interpret because it is highly dependent on how a study chose to aggregate species. For reference see: Plagányi and Essington (2014), When the SURFs up, forage fish are key. Fisheries Research 159: 68-74.
    ${ }^{10}$ In this context, productivity encompasses both the intrinsic rate of increase ( $r$ ) and the carrying capacity (K).

[^7]:    ${ }^{11}$ (footnote from 3.1 table) When determining an appropriate timeframe it is important to consider the ability of managers to adjust management measures to take into account the latest scientific information and advice, for example, if a stock assessment identifies that overfishing took place in the previous fishing season, do managers adjust the harvest controls for the upcoming season?

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